# The Chief Joseph Hatchery Program 2016 Annual Report 

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This report includes both hatchery production/operations and the corresponding monitoring activities completed through April of 2017. It is structured to meet the RM\&E technical report formatting requirements for BPA, and therefore the hatchery production portion is included in Appendix A.

Reports, program descriptions, annual review materials and background information, news and contact information can be found on our website at: https://www.cct-fnw.com/reports/.

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## Executive Summary

The Colville Confederated Tribes (CCT) Chief Joseph Hatchery (CJH) is the fourth hatchery obligated under the Grand Coulee Dam/Dry Falls project, originating in the 1940s. Leavenworth, Entiat, and Winthrop National Fish Hatcheries were built and operated as mitigation for salmon blockage at Grand Coulee Dam, but the fourth hatchery was not built, and the obligation was nearly forgotten. After the Colville Tribes successfully collaborated with the United States to resurrect the project, planning of the hatchery began in 2001 and construction was completed in 2013. The monitoring program began in 2012 and adult Chinook Salmon were brought on station for the first time in June 2013. Bonneville Power Administration (BPA) is the primary funding source for CJH, and the Mid-Columbia PUDs (Douglas, Grant and Chelan County) have entered into cost-share agreements with the tribes and BPA in order to meet some of their mitigation obligations.

The CJH production level was set at 100\% in 2016 during the fourth year of operation for the Spring and Summer/Fall Chinook programs. Leavenworth National Fish Hatchery (LNFH) provided 639 Spring Chinook broodstock in May 2016. The Leavenworth Spring Chinook program broodstock survival was $99.4 \%$ for females, $98.8 \%$ for males with a combined survival of $99.1 \%$. The total green egg take for the Leavenworth Spring Chinook program was 1,060,200 ( $97 \%$ of full program). Green egg to eyed egg survival was $51.1 \%$. This survival was much lower than the standard ( $90 \%$ ). Survival from incubation to ponding for the Leavenworth groups was $94.5 \%$ which was on par with the standard (95\%). With the lower than anticipated hatchery survival of eggs and juveniles the segregated spring Chinook program was $53 \%$ of full program. The 10 (j) spring Chinook reintroduction program received its full component of 206,138 eyed eggs from the Winthrop NFH in October.

In July and August the CCT used a purse seine vessel to collect 1,124 summer/fall Chinook for broodstock for both the integrated and segregated programs (including Similkameen). Additionally, 14 summer/fall Chinook were collected at the Okanogan adult weir in September. Because we had fallen short of our target goal of late-arriving brood from the weir, hatchery staff beach seined the Kline site on the Similkameen River and collected the remaining 72 brood fish. The summer/fall collected enough brood to meet full production levels. The cumulative pre spawn holding survival, for all Summer/Fall brood collected, was 95.9\% for hatchery-origin broodstock (HOB) and 93.4\% for natural-origin broodstock (NOB). Both hatchery-origin and natural-origin brood met the survival standard (90\%) except Jacks, which are not included in the stated cumulative survival. Total green egg take for the season was 2,675,000 ( $98 \%$ of full program). Egg survival from green egg to eyed egg averaged 75.1\% for NOB and $81.3 \%$ for HOB, both under the survival standard ( $90 \%$ ) for this life stage. Cumulative egg survival from green egg to eyed egg was $85.8 \%$ for NOB and HOB, which is
under the survival standard (90\%) for this life stage. Ponding survival for the integrated subyearling program was $96.7 \%$ and for the segregated subyearling program it was $98.4 \%$, both exceeding the survival standard (95\%) for this life stage. The cumulative ponding survival for the integrated and segregated programs was $96 \%$ for the subyearlings, which was over the survival standard (95\%) and 97\% for the yearlings, which also exceeded the survival standard. After in-hatchery mortalities from pre-spawn holding through ponding there were 1,245,695 fish on hand at the end of April for the yearling releases in 2018 ( $96 \%$ of full program) and 403,753 fish on hand for the sub-yearling releases in May 2017 (58\% of full program).

2016 was the third year for Summer/Fall Chinook sub-yearling hatchery releases from the CJH programs and the second year for yearlings released from Similkameen and Omak acclimation ponds that had been reared at the CJH central facility. In April, 232,353 integrated yearling summer/fall Chinook were released from the Omak acclimation pond (58\% of full program) and 256,294 were released by Washington Department of Fish \& Wildlife (WDFW) from the Similkameen Pond ( $64 \%$ of full program). Subsequently, sub-yearlings from brood year (BY) 2015 were transferred to Omak Pond for short term acclimation and 175,771were released in May ( $59 \%$ of full program). Additionally, 401,215 yearling and 218,383 sub-yearling segregated Chinook were released directly from Chief Joseph Hatchery ( $80 \%$ and $55 \%$ of full program, respectively). For Spring Chinook, 201,821 smolts ( $101 \%$ of full program) were released in the spring of 2017 and mark the third year of implementation of the non-essential experimental population under section 10(j) of the Endangered Species Act. Additionally, 555,636 segregated Spring Chinook smolts were released directly from Chief Joseph Hatchery (79\% of full program).

After release, yearling hatchery spring Chinook survival was 57-80\% to Rocky Reach Juvenile bypass and 50-60\% to McNary Dam and was similar to 2015 and to other nearby programs in the Methow. Summer Chinook sub-yearling survival was 35-44\% to Rocky Reach Dam and $14 \%$ to McNary Dam. The vast majority (82-95\%) of PIT tagged hatchery smolts released from Omak Pond and Riverside Pond migrated to the lower Okanogan River within one week of release. This assessment suggests that the program was successful at releasing actively migrating smolts. Two yearling Summer Chinook migrants with a PIT tag from Omak Pond in 2016 returned to the Okanogan as mini-jacks in September after migrating to below Bonneville Dam.

The CJH monitoring project collected field data to determine Chinook population status, trend, and hatchery effectiveness centered on six major activities; 1) rotary screw traps (juvenile outmigration, natural-origin smolt PIT tagging) 2) beach seine (natural-origin smolt PIT tagging, smolt to adult return) 3) lower Okanogan adult fish pilot weir (adult escapement, proportion of hatchery-origin spawners [pHOS], broodstock) 4) spawning ground surveys (redd and carcass surveys)(viable salmonid population [VSP] parameters) 5) eDNA collection (VSP parameter-distribution/spatial structure) and 6) coded wire tag lab (extraction and reading).

Rotary screw trap operations began on April 4 and continued through June 24, capturing 12,660 natural-origin Chinook and 661 hatchery-origin Chinook. After conducting 4 markrecapture events, the efficiency of the trapping configuration was calculated to be approximately $0.10 \%$. This translated to an overall juvenile natural-origin Chinook outmigration estimate of $10,566,870$ with $95 \%$ confidence intervals of $2,118,348$ to 19,015,391. 70 steelhead ( 0. mykiss) were also captured in the rotary screw trap including 27 natural-origin (adipose fin present and no CWT) and 43 hatchery-origin (adipose fin clipped and/or CWT present). Other species commonly caught in the rotary screw traps included sockeye ( 0 . nerka) (11), yellow perch (P.flavescens) (121), bluegill (L. macrochirus) (82) and mountain whitefish (Prosopium williamsoni)(452).

15,802 juvenile Chinook salmonids were collected with the beach seine, and 13,650 ( $86 \%$ ) were PIT tagged and released. Pre- and post-tag mortality was $2 \%$ and $4 \%$ respectively. In 2016, wild summer Chinook tagged at the mouth of the Okanogan had a minimum apparent survival to Rocky Reach Juvenile Bypass (RRJ) of 0.10. Apparent survival to McNary (MCN), John Day (JDA) and Bonneville (BON) Dams was $0.6 \%, 0.2 \%$ and $0.1 \%$ respectively

The lower Okanogan Adult Fish Weir was deployed on August 8th when discharge was 1,530 cfs. The thermal barrier was present in the lower Okanogan after installation until August 22th when the mean Okanogan River temperature began dropping below $22.5^{\circ} \mathrm{C}$. The mean temperature generally varied between $21.4{ }^{\circ} \mathrm{C}$ and $22.5^{\circ} \mathrm{C}$ until August $25^{\text {th }}$. On August 26 the thermal barrier began to break down, allowing Chinook to migrate up the Okanogan. After reviewing the number of adult Chinook pit tagged at Bonneville and their detections at the Wells Adult Ladder and the Lower Okanogan Pit Array, we suspect that about $45 \%$ of fish passage occurred before the weir trap was operational on August 22. After trapping began, the majority of Chinook (75\%) were trapped between August 31 and September 6. 169 adult Chinook were trapped in 2016. Sixteen natural-origin Chinook were transported to the hatchery and held as broodstock for the integrated program. Adult brood were transported from the weir trap to the hatchery brood truck with the Whooshh ${ }^{\text {TM }}$ fish transport system. There were no immediate mortalities of these fish within the first week after transport to the hatchery. All other naturalorigin fish were released upstream of the weir unharmed. All of the hatchery-origin fish encountered in the weir trap were removed for proportion of hatchery-origin spawner (pHOS) management. Only $1.4 \%$ of the Chinook spawning escapement was detected in the trap. All Chinook and sockeye mortality encountered at the weir were categorized as impinged on the upstream side, indicating that they most likely died upstream and floated down onto the weir. The majority of the Chinook carcasses were encountered a week before the majority of Chinook were encountered in the trap, which suggests that the weir was not causing the mortality. The head differential, river velocity, and trap capacity were within the NOAA standard operating criteria. Water quality information, including dissolved oxygen, turbidity, and total dissolved solids were collected to assess potential impacts to increased fish mortality. Weir trapping operations ceased on September 22.

Spawning ground surveys estimated 5,276 summer/fall Chinook redds and 2,755 carcasses were recovered ( 2,171 natural-origin and 584 hatchery-origin). Adult summer/fall Chinook spawning escapement in 2016 was estimated to be 10,605 , with 8,661 natural-origin spawners. The values for $\mathrm{pHOS}(0.18)$ and proportion of natural influence (PNI) (0.87) in 2016 met the program objectives ( $<0.30$ and $>0.67$, respectively). The five-year average for pHOS (0.27) and PNI ( 0.82 ) met the long-term goal ( $<0.30 \mathrm{pHOS} ;>0.67 \mathrm{PNI}$ ). Selective harvest activities by CCT and WDFW contributed to the reduced pHOS and increased PNI in 2016, along with removals of more than 6,000 surplus hatchery fish at the CJH ladder and trap.

Spatial distribution of spring-Chinook in the Okanogan basin has been monitored using analysis of environmental DNA (eDNA) beginning in 2012. This data is used to assess status and trends in spatial structure and to track the progress of the reintroduction which began in 2015. Results revealed that the Okanogan basin likely saw a limited distribution of spring Chinook, even prior to the reintroduction effort. Following the initial reintroduction, several tributaries have produced consistent annual detections of Chinook eDNA, including Shingle Creek, Vaseux Creek, Salmon Creek and Omak Creek.

PIT tags were also used to evaluate Spring Chinook presence and distribution in the Okanogan. Of the 64 returning fish with a PIT tag to the Okanogan, 12 (19\%) had a final detection in the Methow, and 12 were detected in a U.S. tributary and four (6\%) were detected in Canada with the remaining 36 (56\%) only being detected at a mainstem site (OKL and Zosel Dam). Forty-five (70\%) PIT tags detected in the Okanogan had been tagged as an adult at Wells Dam, thirteen (20\%) were from the 2015 release at Riverside Pond, four (6\%) were from juvenile releases in the Methow and two (3\%) were natural-origin that had been tagged as a juvenile in the Methow and Entiat rivers.

The CJH coded wire tag lab was in its first year of operation in 2016. Coded wire tags were extracted and read from Chinook snout recoveries from broodstock, ladder surplus, purse seine harvest, and creel and spawning ground surveys in 2013-2016. Prior to 2013 WDFW had always processed CWT from the Okanogan program. However, after 2013, WDFW decided they could not process the increasing number of snouts generated by the expanding CCT programs without a separate contract. In 2016, CJH M\&E program was able to set up a CWT lab at the hatchery and process all the snouts, thereby catching up on three years of backlogged CWT data. From this point forward CCT will be able to efficiently process and upload (to RMIS) all of the CJH program and Okanogan monitoring CWT data each year. From 2013-2016 the majority of the summer Chinook adult returns to the CJH ladder were from Wells (26\%), Chelan (29\%) and Dryden (14\%). The Similkameen represented an average of 9\% over the four year period. The high proportion of fish from downstream Upper Columbia River programs is not surprising because adult returns from the CJH segregated program will not return until 2017. Spring Chinook were encountered during the summer Chinook ladder operations (generally early July to October). The majority were from the Leavenworth Hatchery $44 \%$ and Chiwawa Pond (34\%).

The majority (95\%) of hatchery-origin spawners recovered on the spawning grounds in 2016 were from Similkameen (92\%) and Okanogan (3\%). This was very similar to the average (95\%) of recent years (2006-2016). Fish acclimated at the Similkameen Pond site spawned primarily in the lower Similkameen River (55\%, reach S1) with substantial numbers of fish also spawning in habitat within the upper Okanogan reaches, 05 and 06 (45\%). Less than 1\% of the spawning escapements in the non-target basins of Wenatchee, Methow, Chelan or Entiat consisted of hatchery-origin Okanogan summer/fall Chinook in 2016. The most recent brood year that could be fully assessed (through age 5) for stray rate of Okanogan/Similkameen fish to spawning areas outside the Okanogan was 2011. The 2011 brood year had a stray of $1.1 \%$, which was similar to the long term and recent five year average.

An Annual Program Review (APR) was held in March 2017 to share hatchery production and monitoring data, review the salmon forecast for the upcoming year, and develop action plans for the hatchery, selective harvest, and monitoring projects. Based on a strong pre-season forecast of 63,100 Upper Columbia summer/fall Chinook, the plan for 2017 is to operate the hatchery at full program levels of 2 million summer/fall Chinook and 900,000 spring Chinook. To maximize PNI, broodstock for the integrated program would be $100 \%$ natural-origin broodstock (NOB) and CCT would plan to harvest their full allocation with the selective harvest program removing as many adult hatchery Chinook as possible with the purse seine, the weir, and at the hatchery ladder.

## INTRODUCTION

Salmon (Oncorhynchus spp.) and steelhead (O. mykiss) faced many anthropogenic challenges ever since European settlement of the Pacific Northwest. Harvest, hydropower development, and habitat alteration/disconnection have all had a role in reducing productivity or eliminating entire stocks of salmon and steelhead (MacDonald 1894; UCSRB 2007). These losses and reductions in salmon had a profound impact on Native American tribes, including the Confederated Tribes of the Colville Reservation. Hatcheries have been used as a replacement or to supplement the wild production of salmon and steelhead throughout the Pacific Northwest. However, hatcheries and hatchery practices can pose a risk to wild populations (Busack and Currens 1995; Ford 2002; McClure et al. 2008). As more studies lead to a better understanding of hatchery effects and effectiveness, hatchery reform principles were developed (Mobrand et al. 2005; Paquet et al. 2011). The CJHP is one of the first of its kind to be structured using many of the recommendations emanating from Congress's Hatchery Reform Project, the Hatchery Science Review Group (HSRG) and multiple independent science reviews. Principally, the success of the program is not based on the ability to meet the same fixed smolt output or the same escapement goal each year. Instead, the program is managed for variable smolt production and natural escapement. Success is based on meeting targets for abundance and composition of natural escapement and hatchery broodstock (HSRG 2009). Chief Joseph Hatchery Program
(CJHP) managers and scientists are accountable for accomplishments and/or failures, and therefore, have well-defined response alternatives that guide annual program decisions. For these reasons, the program is operated in a manner where hundreds of variables are monitored, and activities are routinely and transparently evaluated. Functionally, this means that directed research, monitoring, and evaluation (RM\&E) are used to determine status and trends and population dynamics, and are conducted to assess the program's progress in meeting specified biological targets, measure hatchery performance, and in reviewing the key assumptions used to define future actions for the entire CJHP.

The actions being implemented by the Colville Tribes, in coordination with regional management partners, represent an extraordinary effort to recover Okanogan and Columbia River natural-origin Chinook Salmon populations. In particular, the Tribes have embraced hatchery program elements that seek to find a balance between artificial and natural production and address the goals of increased harvest and conservation.

Two hatchery genetic management plans (HGMPs) were initially developed for the CJH during the Northwest Power and Conservation Council (NPCC) three-step planning process one for summer/fall Chinook (CCT 2008a) and one for spring Chinook (CCT 2008b). Each of the two plans included an integrated and a segregated component. Integrated hatchery fish have a high proportion of natural origin parents, are released into the Okanogan River system and a proportion of these fish are expected to spawn in the natural environment. Segregated fish have primarily hatchery parents, are to be released from CJH directly into the Columbia River and adult returns are targeted exclusively for harvest.

In 2010 the CCT requested that the National Marine Fisheries Service (NMFS) designate a non-essential experimental population of spring Chinook in the Okanogan utilizing section 10(j) of the Endangered Species Act (ESA). In order to obtain a permit to transfer ESA listed fish from the Methow River to the Okanogan River, a new HGMP was developed (CCT 2013). Biological Opinions (BiOps) and permits have been issued by NMFS for the 2008 HGMPs, and CCT acquired a BiOp and permit for the 2013 spring Chinook in 2014. The program will be guided by all three HGMPs.

At full program the facility will rear up to 2 million summer/fall Chinook and 900,000 spring Chinook. Up to 1.1 million summer/fall Chinook will be released in the Okanogan and Similkameen Rivers as an integrated program and 900,000 will be released from CJH as a segregated program. Up to 700,000 segregated spring Chinook will be released from CJH and up to 200,000 Met Comp spring Chinook from the Winthrop National Fish Hatchery (WNFH) will be used to reintroduce spring Chinook to the Okanogan under section 10(j) of the ESA. In 2015, the summer/fall and spring Chinook program's production level was set at full production capacity.

The CJHP will increase harvest opportunity for all anglers throughout the Columbia River and Pacific Ocean. Additionally, the Colville Tribes and other salmon co-managers have worked
with the mid-Columbia Public Utility Districts to meet some of their hydro-system mitigation through hatchery production (CPUD 2002a; CPUD 2002b; DPUD 2002).

In order to make full use of the best science available the program operates on the following general principles ${ }^{1}$ :

1. Monitor, evaluate and adaptively manage hatchery and science programs
2. Manage hatchery broodstock to achieve proper genetic integration with, or segregation from natural populations
3. Promote local adaptation of natural and hatchery populations
4. Minimize adverse ecological interactions between hatchery- and natural-origin fish
5. Minimize effects of hatchery facilities on the ecosystem
6. Maximize survival of hatchery fish in integrated and segregated programs
7. Develop clear, specific, quantifiable harvest and conservation goals for natural and hatchery populations within an "All-H" (Hatcheries, Habitat, Harvest and Hydro) context
8. Institutionalize and apply a common analysis, planning, and implementation framework
9. Use the framework to sequence and or prioritize actions
10. Hire, train, and support staff in a manner consistent with successful implementation of the program
11. Conduct annual reviews to include peers, stakeholders, and regional managers, and
12. Develop and maintain database and information systems and a highly functional informational web-presence.

The CJHP annual RM\&E activities were focused on five primary field activities to provide data for answering key management questions. These activities included:

1. Rotary screw traps (juvenile outmigration, natural-origin smolt PIT tagging)
2. Beach seine (natural-origin smolt PIT tagging)
3. Lower Okanogan adult fish pilot weir (adult escapement, pHOS , broodstock)
4. Spawning ground surveys (redd and carcass surveys)(VSP parameters)
5. eDNA collection (VSP parameter—distribution/spatial structure)

Additional data compilation activities occurred and were necessary in conjunction with our field efforts to answer the key management questions. These included:

1. Harvest (ocean, lower Columbia, terminal sport, and CCT)
2. Query RMIS for coded wire tag (CWT) recoveries to evaluate strays and stock composition

[^0]3. Query PTAGIS for PIT tag returns at mainstem dams and tributaries
4. EDT model estimates for abundance and productivity (from OBMEP)

In-hatchery monitoring/data collection was focused in five areas (see Appendix A):

1. Broodstock collection and bio-sampling
2. Life stage survival
3. Disease monitoring
4. Tagging, marking, and release
5. Ladder surplus / pHOS reduction

## Study Area

The primary study area of the CJHP lies within the Okanogan River Subbasin and Columbia River near Chief Joseph Dam in north central Washington State (Figure 1). The Okanogan River measures approximately 185 km long and drains 2,316,019 ha, making it the third-largest subbasin to the Columbia River. Its headwaters are in Okanagan Lake in British Columbia, from which it flows south through a series of four lakes before crossing into Washington State at Lake Osoyoos. Seventy-six percent of the area lies in Canada. Approximately 14 km south of the border, the Okanogan is joined by its largest tributary, the Similkameen River. The Similkameen River watershed is 510 km long and drains roughly 756,096 ha. The Similkameen contributes approximately 75\% of the flow to the Okanogan River. The majority of the Similkameen is located in Canada. However, part of its length within Washington State composes an important study area for CJHP. From Enloe Dam (Similkameen rkm. 14) to its confluence with the Okanogan, the Similkameen River contains important Chinook pre-spawn holding and spawning grounds. Downstream of the Similkameen confluence, the Okanogan River continues to flow south for 119 km until its confluence with the Columbia River at Columbia River km 853, between Chief Joseph and Wells dams, near the town of Brewster, Washington.


Figure 1. Map of the U.S. portion of the Okanogan River Basin, the Chief Joseph Hatchery (CJH), Winthrop National Fish Hatchery (WNFH), Okanogan adult weir (Weir), Rotary screw trap (RST), and Chinook Salmon acclimation sites. Horizontal coordinate information is referenced to the North American Datum of 1983 (NAD 83).

Similar to many western rivers, the hydrology of the Okanogan River watershed is characterized by high spring runoff and low flows occurring from late summer through winter. Peak flows coincide with spring rains and melting snowpack (Figure 2). Low flows coincide with minimal summer precipitation, compounded by the reduction of mountain snowpack. Irrigation diversions in the lower valley also contribute to low summer flows. As an example, at the town of Malott, Washington (rkm. 27), Okanogan River discharge can fluctuate annually from less than 1,000 cfs to over 30,000 cfs (USGS 2005).

The Okanogan Subbasin experiences a semi-arid climate, with hot, dry summers and cold winters. Water temperature can exceed $25^{\circ} \mathrm{C}$ in the summer, and the Okanogan River surface usually freezes during the winter months. Precipitation in the watershed ranges from more than 102 cm in the western mountain region to approximately 20 cm at the confluence of the Okanogan and Columbia Rivers (NOAA 1994). About 50\% to $75 \%$ of annual precipitation falls as snow during the winter months.

For most of its length, the Okanogan River is a broad, shallow, low gradient channel with relatively homogenous habitat. There are few pools and limited large woody debris. Fine sediment levels and substrate embeddedness are high and large woody debris is rare (Miller et al. 2013). Towns, roads, agricultural fields and residential areas are adjacent to the river through most of the U.S. reaches.

Near its mouth, the Okanogan River is affected by the Wells Dam on the Columbia River, which creates a lentic influence to the lowermost 27 km of the Okanogan River. Water level fluctuates frequently because of operational changes (power generation, storage) at Wells Dam.


Figure 2. Okanogan River mean daily discharge (blue lines) and water temperature (red lines) at Malott, WA (USGS Stream Gage 12447200).

## METHODS

## Tag and Mark Plan

HATCHERY SUMMER/FALL CHINOOK. -All summer/fall hatchery-origin Chinook were marked with an adipose fin clip to ensure differentiation from natural-origin fish in the field and in fisheries. Additionally, all summer/fall Chinook raised for the integrated program have been/will be tagged with a CWT (with distinct codes differentiated by release location), which is inserted into the snout of fish while in residence at the hatchery. A batch of 200,000 summer/fall Chinook in the segregated program will receive a CWT, so the presence or absence of a CWT in adipose-clipped fish is a partial diagnostic as to which program an ad-clipped, hatchery-origin fish belongs (Table 1).This will allow for selective efforts in broodstock collection, purse seining, and hatchery trapping activities to be program specific by determining
the presence or absence of a CWT in the field. It was decided that losing some resolution on field differentiation of the segregated and integrated populations was a good tradeoff in order to get the harvest information back from the batch of 200,000 CWT in the segregated program.

Under this strategy, a returning adult from the CJH with an adipose fin clip and CWT would be considered part of the integrated program and either collected for broodstock in the segregated program, allowed to escape to the spawning grounds (if pHOS is within acceptable levels), or removed from the population (for harvest or pHOS management). If a fish has an adipose fin clip but no CWT, then it is assumed from the segregated program (or a stray from another hatchery program) and removed for harvest or pHOS management. In this way, CWTs assist with in-season management of hatchery-origin stocks in the field. The 200,000 segregated fish with a CWT represent about $15 \%$ of the combined segregated $(900,000)$ and integrated ( 1.1 million) hatchery fish with a CWT. If smolt to adult survival and adult holding/migration behaviors are identical, this would mean that $15 \%$ of the subsequent generation of segregated fish would have a segregated parent and would not be consistent with the 'stepping stone' approach. However, segregated fish should spend less time holding at the mouth of the Okanogan and therefore have a lower probability of being collected as broodstock in the purse seine. CWT monitoring from broodstock collections during the first several years of returns will provide insight to this tradeoff.

Coded wire tags are recovered from salmon carcasses during Chief Joseph Hatchery ladder surplus, CCT creel surveys, CCT purse seine, Okanogan weir trapping, and spawning ground surveys in the Okanogan Basin. All recovered CWTs are sent to the Chief Joseph Hatchery coded wire tag lab for extraction, reading, and data upload to the Regional Mark Processing Center operated by the Pacific States Marine Fisheries Commission (PSMFC) ${ }^{2}$. These data are used to develop estimates of total recruitment, rate of return to point of release (homing), contribution to fisheries, survival rates, mark rate, and other parameters, helping inform future management and production decisions within the CJHP.

[^1]Table 1. General mark and tag plan for Chief Joseph Hatchery summer/fall Chinook.

| Mark Group | Target max <br> smolt released | Life-stage <br> released | \% CWT | Adipose <br> Fin-Clip | PIT tag |
| :--- | :--- | :--- | :--- | :--- | :--- |
| Okanogan <br> Integrated | $1,100,000$ |  |  |  |  |
| Similkameen | 400,000 | Yearling | $100 \%$ | $100 \%$ |  |
| Omak Pond | 400,000 | Yearling | $100 \%$ | $100 \%$ | 5,000 |
|  | 300,000 | Sub- <br> yearling | $100 \%$ | $100 \%$ | 5,000 |
| Chief Joseph <br> Segregated | 500,000 | Yearling | $20 \%$ | $100 \%$ | 5,000 |
|  | 400,000 | Sub- <br> yearling | $25 \%$ | $100 \%$ | 5,000 |
| Natural-Origin | RST and <br> Confluence Seine | N/A | $0 \%$ | $0 \%$ | $\leq 25,000$ |

${ }^{1}$ The original plan was to use Riverside Pond for approximately $1 / 3$ of the summer Chinook yearling production, however, to date it has been only been used to acclimate the $10(\mathrm{j})$ spring Chinook because Tonasket Pond has not been rehabilitated for acclimation of spring Chinook.

In addition to the adipose fin-clip and CWT, a subset of hatchery-origin fish will be PITtagged to further assist with fish monitoring efforts in subsequent years. Table 1 represents the general plan at full production.

HATCHERY SPRING CHINOOK. -The general tag and mark plan for spring Chinook can be seen in Table 2.

Table 2. General marking and tagging plan for Okanogan spring Chinook as part of the Chief Joseph Hatchery Program.

| Mark Group | Smolt <br> released | Life-stage <br> released | \% CWT <br> (\#) | Adipose <br> Fin-Clip | PIT tag |
| :--- | :--- | :--- | :--- | :--- | :--- |
| Chief Joseph <br> Segregated | 700,000 | Yearling | $29 \%$ <br> $(200,000)$ | $100 \%$ | 5,000 |
| Reintroduction <br> (§10(j) fish from <br> Winthrop) |  |  |  |  |  |
| Tonasket or <br> Riverside Pond | 200,000 | Yearling | $100 \%$ |  | 5,000 |
| Natural-Origin | RST | Yearling | $0 \%$ | $0 \%$ | $\leq 5,000$ |

NATURAL-ORIGIN FISH TAGGING. -The RM\&E plan called for up to 25,000 PIT tags in juvenile natural-origin summer/fall Chinook parr/smolts. PIT tagging of natural-origin
summer/fall Chinook occurred at the rotary screw trap and the juvenile beach seine in 2016. Please see those sections for details.

## Genetic Sampling/Archiving

The CJHP collects and archives genetic samples for future analysis of allele frequency and genotyping of naturally spawned and hatchery Chinook populations. Genetic samples (fin clips) from outmigrant juvenile Chinook were collected during rotary screw trap operations. Samples were preserved in 200-proof molecular grade ethanol and are currently archived at USGS Snake River Field Station, Boise, ID. No genetic analyses are currently being conducted. Annual tissue collection targets are at least 200 samples for: (1) natural-origin sub-yearling Chinook handled at the rotary screw trap/beach seine; (2) natural-origin yearling ( $>130 \mathrm{~mm}$ ) Chinook handled at the rotary screw trap/beach seine and (3) natural- and hatchery-origin (100 each) Chinook encountered during carcass surveys on the spawning grounds.

The CJHP has also supported requests from Columbia River Inter-tribal Fish Commission (CRITFC) to provide genetic samples (caudal punches) from CJH summer- and spring-Chinook broodstock to aid in the development of a Columbia River Parentage Based Tagging (PBT) program. Samples were preserved on pre-labeled Whatman (GE Healthcare, Pittsburg, PA, USA) cellulose chromatography paper and shipped to CRITFC Lab in Hagerman, ID, USA. Genetic samples will continue to be collected from all hatchery broodstock at CJH.

## Rotary Screw Traps

One 2.4 m and one 1.5 m rotary screw trap (RSTs) were deployed from the Highway 20 bridge near the City of Okanogan (rkm. 40) (Figure 3). The RSTs were operated from April 4 to June 24, 2016. Trapping typically occurred continuously from Monday through Friday, and for 12 hours, from 2000-0800 Saturday and Sunday. Trapping operations were suspended on April 23-26, April 30, May 28-29, June 4-5, June 11-12, and June 17-19 due to high river discharge and debris load or staffing constraints. To continue trapping operations in varying river conditions, traps were operated in one of three trapping configurations: 2.4 m only, 1.5 m only, and both traps operational.


Figure 3. 2.4-m (left) and $1.5-\mathrm{m}$ (right) traps fishing in the Okanogan River. The boat is used by technicians to access the $2.4-\mathrm{m}$ trap. Photo by CCT.

During operation, the trap locations were adjusted in the river to achieve between 5-10 revolutions per minute. The traps were checked every two hours unless a substantial increase in flow ( $\geq 500$ cfs in a 24 -hour period) or debris load occurred, in which case they were checked and cleaned more frequently. All fish were enumerated, identified to species, and life stage, origin (adipose fin present or absent), and disposition (whether the fish was alive or dead), and a subsample of natural-origin Chinook was measured. The fork lengths of the first 10 unmarked Chinook of each 100 encountered in the live well were measured to the nearest mm and released during each trap check. Steelhead smolts were not measured in order to minimize handling and stress of ESA-listed species. Unmarked (adipose fin present) Chinook captured in the RST that were $\geq 65 \mathrm{~mm}$ total length received a 12 mm full duplex PIT tag. A tissue sample (fin clip) was collected from (1) all fish that received a PIT tag and (2) any yearling unmarked Chinook for future genetic analyses.

EFFICIENCY ESTIMATES. - An estimate of the daily number of juvenile out migrants passing the trap location requires an estimate of the proportion of fish caught by the traps. This was accomplished using mark-recapture methodologies developed by Rayton and Wagner (2006), maintaining continuity with the techniques employed at this RST operation in previous years. This mark-recapture procedure (hereafter referred to as an efficiency trial) was
conducted using both natural-origin sub yearling Chinook and hatchery-origin yearling Chinook. Only fish with a fork length of at least 45 mm were used in efficiency trials.

After collection from both the 2.4 m and 1.5 m rotary screw traps, fish were marked in 5 gal buckets with Bismarck Brown dye at a concentration of $0.06 \mathrm{~g} / \mathrm{gal}$, held for 10-15 minutes with aeration and transported in buckets via a truck for release. Fish were released at night (typically between 0000 and 0330) approximately 1.6 river km upstream by the Oak Street Bridge. Fish were distributed evenly on both sides of the river to allow for equal distribution across the channel. The probability of capture was assumed to be the same for hatchery-origin fish as it was for natural-origin fish.

Because of variable flow and debris conditions, at any given moment, one of several trapping configurations could have been employed, in which either one, both, or neither of the 2.4 and 1.5 m screw traps could be operating. In order to derive an ultimate out migrant estimate, efficiency estimates for all of these configurations were calculated.

Trap efficiency was calculated by the equation

$$
E_{t i}=\sum R_{t i} / \sum M_{i}
$$

where $E_{t i}$ is the trap efficiency for trapping configuration $t$ in sampling period $i, \sum R t i$ is the sum of marked fish that are recaptured in trap configuration $t$ during sampling period $i$, and $\sum M_{i}$ is the sum of marked fish released during the sampling period $i$.

Trap efficiencies were recorded for each individual trap as it operated, and for both traps operating in unison. Trap efficiencies for each individual trap were further refined by including results for each individual trap while both traps were in operation. For example, if 100 marked fish were released, and 1 was recaptured in each trap, each individual trap displays an efficiency of $1 \%$, and the efficiency of both traps operating simultaneously is $2 \%$. This relies on the assumption that the efficiency of each trap is unaffected by whether the other is operating or not.

RST ANALYSIS. - Hourly catch was expanded to an hourly outmigration estimate based on measured trap efficiency by using the Lincoln-Peterson mark-recapture model with a Chapman modifier, which can improve estimates when recapture rates are low (Seber 1982). This model relies on the following assumptions:
1.) All marked fish passed the screw trap or were recaptured during time period $i$
2.) The probability of capturing a marked or unmarked fish is equal
3.) All marked fish recaptured were correctly identified as a marked fish
4.) Marks were not lost or overlooked between time of release and recapture

Total juvenile Chinook emigration was calculated for each trap configuration using a pooled Peterson estimator with a Chapman modification, such that

$$
\widehat{N}=\left[\frac{\left(M_{p}+1\right)\left(C_{p}+1\right)}{\left(R_{p}+1\right)}\right]-1
$$

Where $\widehat{N}$ is total emigration estimate, $M_{p}$ is the total number of marked individuals during the trapping season, $C_{p}$ is the total number of fish caught during the trapping season, and $R_{p}$ is the total number of recaptured fish during the trapping season.

An approximately unbiased estimate of the variance of the population, $\widehat{V}[\widehat{N}]$, is calculated by the equation

$$
\widehat{V}[\widehat{N}]=\frac{\left(M_{p}+1\right)\left(C_{p}+1\right)\left(M_{p}-R_{p}\right)\left(M_{p}-R_{p}\right)}{\left(R_{p}+1\right)^{2}\left(R_{p}+2\right)}
$$

The precision of the population estimates was assessed by including $95 \%$ confidence intervals calculated by the equation

$$
\widehat{N} \pm 1.96 \sqrt{\widehat{V}[\widehat{N}]}
$$

Estimates and confidence intervals were calculated for all trapping configurations and then summed to generate an overall estimate for the trapping season. During periods when neither trap was operating, an estimate was calculated based on the average catch of an equal time period immediately prior and following the inoperable period. For example, no traps were operable on April 30, so catch for that day was estimated to be the average of total catch on April 29 and May 1.

Trapping efficiency and outmigration estimation was also examined using a smolt abundance estimator provided by WDFW and developed for its efforts in the Wenatchee River that incorporates stream flow and weights efficiency trials according to the number of released fish (Murdoch et al. 2012; Ryding 2000).

## Juvenile Beach Seine/PIT tag effort

Portions of the following text describing the methods were taken directly from a draft DPUD report (DPUD 2014).

Beach seining took place from May 12 to June 23 in the area near the confluence of the Okanogan and Columbia Rivers. Efforts focused on beaches along the North bank of the Columbia River, downstream of the mouth of the Okanogan ( $48^{\circ} 6^{\prime} 12.46^{\prime \prime} \mathrm{N}, 119^{\circ} 44^{\prime} 35.48 " \mathrm{~W}$ ) (Figure 4). In 2016, Gebber's Landing and Washburn Island were the only areas used for
collection. This location provided reasonable catch rates, limited bycatch, and provided suitable substrates (limited debris loads/underwater snags) for efficient sampling. Juvenile Chinook from this location were likely primarily fish originating from the Okanogan River; however, it is possible that offspring from mainstem Columbia River spawning could also be included, especially at the Washburn Island site.


Figure 4. Seining locations downstream (Gebber's Landing) and upstream (Washburn Island) of the confluence.

A single beach seine ( $30.49 \mathrm{~m} \times 3.05 \mathrm{~m}$ with a $28.32 \mathrm{~m}^{3}$ 'bag'; Christensen Net Works, Everson, WA) was used to capture fish. Netting was Delta woven 6.4 mm mesh with "fish-green" treatment. Weights ( $3-5 \mathrm{~kg}$ ) were attached to each end of the seine to help keep it open during retrieval.

To capture fish, one end of the seine was tied off to shore, while the other was towed out by boat until the seine was stretched perpendicular to shore. The boat would then pull the seine upstream and return to shore, causing the seine to form a semi-circle intersected by the shore line (Figure 5). The seine bridle was handed from the boat to a shore crew that would retrieve the seine. Juvenile Chinook were transferred to a 10 -gallon tub filled with river water and transferred to a nearby floating net pen. Handling/holding time in the tub was generally <15 minutes. Floating net pens were approximately $5 \mathrm{~m}^{3}$ and consisted of a PVC pipe frame covered with black $19.1-\mathrm{mm}$ and $3.2-\mathrm{mm}$ mesh. The mesh allowed for adequate water exchange, retained juvenile Chinook and prevented the entrance of predators. Noticeable bycatch, most commonly three-spine stickleback (Gasterosteus aculeatus) were released from the seine without enumeration. Any bycatch inadvertently transferred to the floating net pen were later sorted and released during tagging (untagged). On May 12, 19, and 25, fish captured in the beach seine were immediately tagged on the river shore and released after recovery from anesthesia.


Figure 5. Juvenile beach seine being retrieved by CCT staff near the confluence of the Okanogan and Columbia Rivers.

Juvenile Chinook were held 24 hours prior to tagging to assess capture/handling effects. Chinook $\geq 65 \mathrm{~mm}$ were tagged with a full duplex 12 mm PIT tag, and Chinook between 65 and 50 mm were tagged with a full duplex 9 mm PIT tag. After tagging, fish were returned to a floating net pen for 24 hours post-tagging to assess tag loss and tag application/handling mortality rates. Fish were then released to the Columbia River (Wells Pool) several hundred meters downstream of their capture location.

TAGGING PROCEDURES. -Tagging was conducted by CCT staff with support from USGS using a mobile tagging station (Biomark, Co., Boise, ID, USA). The tagging station consisted of an approximately $1 \mathrm{~m}^{2}$ aluminum work surface with a trough for holding fish during the tagging process as well as all the necessary electronics (computer, scale, tag reader, and antenna) needed for tagging. Water was pumped directly from the river using a $1 / 4$ horsepower pump and radiator system to keep water temperatures ambient with river temperatures. When tagging water temperatures were $>17{ }^{\circ} \mathrm{C}$, ice was added to the anesthetic solution to decrease the temperature. A solution of 4.0 g Tricaine methanesulfonate (MS-222) per 1 L of water was used to anesthetize fish prior to tagging. The applied concentration of MS-222 would sedate fish to the desired level of stage- 2 anesthesia in approximately 3 to 4 minutes. All fish were tagged within 10 minutes of the initial exposure. Recovery time was approximately 1 to 2 minutes.

Each tagging location had two net pens: one containing the fish to be tagged, and an empty pen for holding fish post-tagging. Fish to be tagged were collected from the respective net pens using a dip net and placed into an 18.9 L bucket of water. Up to 40 fish at a time were then transferred from the bucket using a smaller dip net and placed into the trough containing the anesthetic solution.

Fish were tagged with 12.5 mm 134.2 kHz ISO PIT tags using pre-loaded, 12-gauge hypodermic needles (BIO12.BPLT) fitted onto injection devices (MK-25). 12.5 mm PIT tags were used to maximize detection at downstream locations, particularly the Rocky Reach Juvenile Bypass and the Bonneville Dam Corner Collector. Detection efficiencies at both of the former sites would dramatically suffer when using the smaller PIT tags available. The tagging crew consisted of one fish sorter, one tagger and one data collector. The data collector interrogated the tag in each tagged fish, recorded its fork length with an electronic wand on a digitizer board, and noted any anomalies. Tagged fish were transferred to the recovery/holding pen via a PVC pipe with flowing water.

Data collected during tagging were stored using PITTAG3 (P3) software (Pacific States Marine Fisheries Commission). After completion of the tagging events, tag files were consolidated, uploaded to PTAGIS (www.ptagis.org), and shared with Douglas PUD.

FISH RELEASES. -Tagged fish were released the morning after they had been tagged. Prior to release, the net pen was opened and all observed mortalities and moribund fish were removed. Once the mortalities were removed the net pen was tilted to allow the fish to volitionally exit. PIT tags were recovered from dead/moribund fish, the associated tag codes
were marked as "Mortalities" in the tag files and the tag codes were deleted. Expelled tags were recovered from the mesh floor via a powerful magnet.

## Lower Okanogan Adult Fish Pilot Weir

The Okanogan adult fish pilot weir (herein referred to as the 'weir') was in its fifth year of design modifications and testing in 2016. Continued operation and improvements to the weir are a central part of CCT's strategy for the successful implementation of the CJHP summer/fall Chinook Salmon (Oncorhynchus tshawytscha) programs. Pilot weir test results are essential for updating key assumptions, operations and design of the weir.

Objectives for the pilot weir in 2016 included:

1. Install the weir in early July and operate until late September under allowable flow conditions ( $<3,000 \mathrm{cfs}$ ) and temperature ( $<22.5^{\circ} \mathrm{C}$ );
2. Document environmental effects of the weir through collection of physical and chemical data in the vicinity of the weir;
3. Test weir trapping operations including live Chinook capture, handling and release;
4. Direct observations and fish counts for estimating species composition, abundance, health, and timing to inform management decisions and future program operations;
5. Collect NOR broodstock at the weir and transport safely to the CJH;

The lower Okanogan fish weir was installed approximately 1.5 km downstream of Malott, WA ( $48^{\circ} 16^{\prime} 21.54 \mathrm{~N} ; 119^{\circ} 43^{\prime} 31.98 \mathrm{~W}$ ) in approximately the same location as previous years. Weir installation began on August 8th at a river flow of 1,530 cfs and was complete with the underwater video system on August 17th. An aluminum trap was installed near the center of the channel at the upstream end of the deep pool in the thalweg of the channel. The trap was 3 m wide, 6 m long and 3 m high (Figure 6). The wings of the weir stretched out from either side of the trap towards the river banks, angling downstream in a slight $V$ configuration. The wings consisted of steel tripods with aluminum rails that supported the 3 m long Acrylonitrile butadiene styrene (ABS) pickets. Each panel was zip-tied to the adjacent panel for strength and stability. Sand bags were placed between panels when needed to fill gaps that exceeded the target picket spacing. Picket spacing ranged from 2.5 to $6.4 \mathrm{~cm}(1$ to 2.5 inch$)$ in 1.2 cm (halfinch) increments (Figure 7). Pickets were manually forced into the river substrate upon deployment and then as needed to prevent fish passage under the weir.

The river-right wing consisted entirely of 2.5 cm picket spacing (Figure 7). A 3 m gap between the last panel and the right shoreline remained to allow for portage of small vessels around the weir. This was a very shallow gravelly area and under most flow conditions it did not appear to be a viable path for adult salmon passage. However, a block net was set up from the
last panel to the river-right shore to limit escapement via this route. The river left wing had variable picket spacing to accommodate non-Chinook fish passage through the pickets. The primary objective of the wider picket spacing was to allow Sockeye (O. nerka) to pass through the weir and reduce the number of Sockeye that would enter the trap. River left was selected for this spacing to better accommodate observation/data collection regarding successful passage of smaller fish through the panels. In past years CCT has observed jack and even adult Chinook passing through the 7.6 cm picket spacing panels. The 7.6 cm picket spacing panels were replaced with 6.4 cm picket spacing panels during deployment to reduce the escapement of smaller hatchery Chinook. This decision was made after consultation with the Technical Oversight Group (TOG) because the majority of Sockeye had already escaped into Canada before the weir was deployed.


Figure 6. Lower Okanogan adult fish pilot weir, 2016. Photo taken in early August during deployment.


Figure 7. Conceptual diagram of picket (ABS pipe) spacing within each panel (or set of 4 panels) at the Lower Okanogan adult fish pilot weir in 2016.

A negative pressure transport tube, known as a Whooshh ${ }^{\text {TM }}$ fish transport system, was installed to assist CCT with broodstock collection at the Okanogan weir (Figure 8). The 49 m tube was connected to an accelerator at the upstream side of the trap and at a mobile trailer fitted with a decelerator, tower, tube reel, blower housing, and accelerator entrance and exit mounts. A diesel generator provided remote power to run the pump that generated power for the pneumatic portions of the system.


Figure 8. Whooshh ${ }^{\text {TM }}$ fish transport system installed at the Okanogan weir. Photo taken from the gangway at the trap.

Physical and chemical data were collected in the vicinity of the weir including the water depth (ft.) inside the trap, water velocity (ft./sec) upstream, downstream and in the weir trap, dissolved Oxygen ( $\mathrm{mg} / \mathrm{L}$ ), total dissolved solids (TDS) (ppm), turbidity (NTU), temperature $\left({ }^{\circ} \mathrm{C}\right)$, discharge (cfs) and head differential (cm). Temperature and discharge were taken from the online data for the USGS gauge at Malott
(http://waterdata.usgs.gov/wa/nwis/uv?site no=12447200). When river temperature exceeded $22.5^{\circ} \mathrm{C}$, trapping operations ceased and weir pickets on panels adjacent to the trap on both sides were raised to allow for unrestricted fish passage.

Five minute tower observations were conducted at least three times a day, in the morning (0600-0800), early afternoon (1200-1400) and evening (1700-1900) and an estimate of the number fish observed was recorded. Ten minute bank observations were conducted about 0.8 river km downstream of the weir, around two pools, at least twice a day, in the morning and afternoon. An estimate of the number of fish observed below the weir was recorded. Algae and debris were cleared off of the weir at least once per day generally in the morning (0800-1000). Dead fish on the upstream side of the weir were enumerated, identified to species and the presence and extent of injuries were noted. The tail was cut off of each mortality before they were tossed downstream of the weir so that they would not be double counted during surveys.

Weir efficiency, a measure of the proportion of total spawning escapement encountered by the weir, was calculated by the equation;

$$
X=\frac{W_{T}}{T}
$$

where $X$ was weir efficiency, $W_{T}$ was the number of adult summer/fall Chinook encountered in the weir trap including released fish, and $T$ was the total summer/fall Chinook spawning escapement for the Okanogan River Basin.

Weir effectiveness was a measure of the proportion of the adult hatchery Okanogan summer/fall Chinook run encountered in the weir trap, becoming available for removal from the population as a form of adult fish management. It was calculated by the equation;

$$
Y=\frac{W_{H}}{W_{H}+H O S}
$$

where $Y$ is weir effectiveness, $W_{H}$ is the number of adult hatchery origin fish encountered in the weir trap, and HOS is the total number of hatchery origin spawners.

Trapping operations were conducted during daylight hours, generally 0400-2000, under allowable temperature conditions ( $\leq 22.5^{\circ} \mathrm{C}$ ) from August 22 to September 22. When fish entered the trap during an active trapping session, the downstream gate was closed and fish were identified and either released, surplussed or collected for brood. Sixteen natural origin Chinook were collected from the weir trap from September 9 to September 14, transported to a 2,500 gallon hatchery truck via the Whooshh ${ }^{\text {TM }}$ fish transport system. The fish were then transported approximately 32 km to Chief Joseph Hatchery where they were held in the broodstock raceways until spawning in October. Otoliths were removed from spawned adults and read at the WDFW otolith lab in Olympia. Results were used to assess whether or not Priest Rapids Hatchery fish were included in the CJH integrated brood.

Mark-recapture studies were performed at the weir trap to assess handling mortality at the weir as well as recovery bias of carcasses on the spawning grounds. All natural-origin Chinook that were trapped and destined for release upstream, were anesthetized with electronic anesthetic gloves, measured, and inserted with a floy tag. After the fish were tagged they were released over the crowder and into the upstream side of the trap where they recovered before they exited through the trap gates on their own volition.

## Spawning Ground Surveys

The objectives for spawning surveys were to:

1. Estimate total spawning escapement based on the number of Chinook redds per reach
2. Estimate the proportion of natural spawners composed of hatchery-origin recruits ( pHOS )
3. Estimate pre-spawn mortality and mean egg retention for wild- and hatchery-origin spawners
4. Determine the origin (rearing/release facility) of hatchery-origin spawners (HOS) in the Okanogan and estimate the spawner composition of out of population and out of ESU strays (immigration)
5. Estimate out-of-population stray rate for Okanogan hatchery Chinook and estimate genetic contribution to out-of-basin populations (emigration)
6. Determine age composition of returning adults through scale analysis
7. Monitor status and trends of demographic and phenotypic traits of wild- and hatchery-origin spawners (age-at-maturity, length-at-age, run timing, SAR)

## REDD SURVEYS

A primary metric used to monitor the status and trends of salmonid populations is spawning escapement. Estimates of spawning escapement can be calculated based on redd counts and expanded by sex-ratios (Matthews and Waples 1991, Gallagher et al. 2007). This requires intensive visual survey efforts conducted throughout the spawning area and over the course of the entire spawning period. Visual redd surveys were conducted to estimate the number of redds per survey reach from the mouth of the Okanogan River to Zosel Dam (river km 124); the Similkameen River from its confluence with the Okanogan River upstream to Enloe Dam (river km 14); and in the mainstem Columbia River from the mouth of the Okanogan River upstream to Chief Joseph Dam (Table 3). Weekly surveys were timed to coincide with spawning in the basin, generally beginning the last week of September or the first week of October and ending approximately the second week of November. Redds were counted using a combination of fixed-wing aerial flight surveys and inflatable raft float surveys.

Aerial surveys occurred once weekly throughout the spawning season, each covering the entire survey area. Aerial surveys were flown at low elevation and at moderate speeds to accommodate visual identification of redds. From the aircraft, a trained observer recorded the number and GPS coordinates of all new redds as the plane passed overhead. All data were recorded directly into a YUMA rugged computer tablet (Trimble Navigation, Ltd.). Aerial surveys were primarily used to document redds in areas inaccessible to rafts, or in areas of low redd densities, such that they did not warrant weekly float surveys. All data points were visualized in

ArcGIS (ESRI, Inc.), and quality controlled to ensure that redd counts were not duplicated during float surveys. Aerial surveys also served a secondary function of informing research crews where to focus weekly carcass recovery efforts (see below section on Carcass Surveys).

Float surveys occurred once daily, 5 days per week throughout the spawning season. Float surveys consisted of three 2-person teams using inflatable rafts to count redds while floating downstream. Each team was responsible for covering one-third of the river width, (1) left bank, (2) center, and (3) right bank. Each individual redd was counted and its position recorded directly into a YUMA rugged computer tablet (Trimble Navigation, Ltd.).

Table 3. Reach names and locations for the Okanogan and Similkameen for summer/fall Chinook Salmon spawning and carcass surveys.

| Stream | Code | Reach Description | River km |
| :---: | :---: | :---: | :---: |
| Okanogan | O1 | Mouth to Malott Bridge | $0.0-27.0$ |
|  | O2 | Malott Bridge to Okanogan Bridge | $27.0-41.8$ |
|  | O3 | Okanogan Bridge to Omak Bridge | $41.8-49.1$ |
|  | O4 | Omak Bridge to Riverside Bridge | $49.1-65.1$ |
|  | O5 | Riverside Bridge to Tonasket Bridge | $65.1-90.9$ |
|  | O6 | Tonasket Bridge to Zosel Dam | $90.9-124.0$ |
| Similkameen | S1 | Mouth to Oroville Bridge | $0.0-8.0$ |
|  | S2 | Oroville Bridge to Enloe Dam | $8.0-14.0$ |
|  | Cx | TBD | TBD |

All redds were classified as either a:

1. Test-redd (disturbed gravel, indicative of digging by Chinook, but abandoned or without presence of Chinook; generally, this classification is reserved for early season redd counts, before substantial post-spawn mortalities have occurred as indicated by egg-voidance analysis of recovered carcasses). Test-redds do not contribute to annual redd counts.
2. Redd (disturbed gravel, characteristic of successful Chinook redd construction and/or with presence of Chinook).
Redds per reach were calculated for each week as the combined number of new redds counted during aerial- and float-surveys for a given week. Post-season analysis consisted of summing the
combined aerial- and float-survey weekly redd totals to calculate annual redd totals per reach, and per total survey area. Estimated total spawning escapement was then calculated by multiplying the total redd count by the expansion factor for the current year ( 2.01 for 2016). The expansion factor $=1+$ the number of males per female as randomly collected for broodstock at Wells Dam (1.01:1.00 in 2016). Assumptions include:

$$
\begin{array}{ll}
\text { Assumption I - } & \begin{array}{l}
\text { Each redd was constructed by a single female Chinook, and each } \\
\text { female Chinook constructed only one redd }
\end{array} \\
\text { Assumption II - } & \begin{array}{l}
\text { The male: female ratio on the spawning grounds was the same for } \\
\text { wild- and hatchery-origin Chinook, and is equal to the male: female } \\
\text { ratio as randomly collected for broodstock at Wells Dam }
\end{array}
\end{array}
$$

Assumption III - Every redd was observable and correctly enumerated

## Escapement into Canada

Video systems operated by OBMEP and located in the fishways of Zosel Dam allow observation of salmonids passing over Zosel Dam and potentially into the British Columbia portion of the Okanagan River Basin. For detailed methods within a particular year please see the Okanogan Basin Monitoring and Evaluation Program (OBMEP) annual reports posted at (http://www.colvilletribes.com/obmep_publications.php).

Passage over Zosel Dam can occur via the fishways or through the open dam gates. OBMEP assumes that any gate level > 1 foot is high enough for fish to pass upstream through the open gate rather than through the fish ladders and video arrays. In high water years, Chinook have the opportunity to pass through the gates rather than through the fishways. The estimates of Chinook escapement past Zosel Dam do not account for fish moving through the gates rather than the fishways. 2014 PIT detections of Chinook in the fishways indicated that smaller fish were able to fall back through the small openings in the Zosel Dam gates and then reascend through the fishways. An ascension/fallback/ ascension rate (AFA) was calculated and applied to the total Zosel estimate in 2014. Fallback/reascension is likely an underestimate of actual fallback since not all fallback re-ascend. Actual fallback is unknown. Due to uncertainties regarding estimations of AFA as well as the limitations of the video observation methods, Chinook passage is reported unadjusted for all subsequent years. AFA was calculated as the ratio of the number of unique PIT tagged fish ( $\mathrm{N}_{\text {PIT }}$ ) ascending the fishways, divided by the total number of their ascensions:

$$
A F A=\frac{N_{P I T}}{\sum_{i-1}^{N_{P I T}} a_{i}}
$$

where,
$N_{P I T}=$ number of unique PIT-tagged fish ascending the ladder(s),
$\mathrm{a}_{\mathrm{i}}=$ number of ascents made by the $i$ th PIT-tagged fish $\left(i=1, \ldots,=1, \ldots, N_{P I T}\right)$.
The video count ( $C$ ) multiplied by the $A F A$ provided an estimate of the total passage abundance (N):

$$
N=C \cdot A F A
$$

There were times when the video system was inactive for routine maintenance and cleaning. To estimate missed fish observations during this period, an average was taken of passage events during the hour before and after the inactive period. Spring Chinook were also removed from the total estimate based on run timing at Zosel.

Data and discussion presented herein are intended to begin the process of understanding what is known, what is not known, and what the possibilities are for obtaining a reliable estimate of summer/fall Chinook spawners in the Canadian portion of the Okanogan River.

## CARCASS SURVEYS

Carcass surveys provide important biological samples for evaluation of hatchery- and natural-origin fish on the spawning grounds, including:

1) Spawner composition
a. pHOS
b. out of population hatchery strays (immigration)
c. spatial distribution of natural- and hatchery origin spawners
2) Fish size
3) Sex-ratio
4) Age structure (CWT and scale analysis)
5) Pre-spawn mortality (i.e. egg retention)

The target sample size for carcass recovery efforts is $20 \%$ of the spawning population within each reach (Hillman et al. 2014). Carcass recovery efforts occurred simultaneously with redd float surveys. Recovered carcasses were transported within inflatable rafts downstream until a suitable site was found for processing. If a carcass was too degraded to sample for biological data, it was returned to the river, unsampled. All adipose absent carcasses were assumed to be of hatchery-origin, and all carcasses displaying an intact adipose fin were assumed to be of natural-origin ${ }^{3}$. Biological data collected from carcasses included sex, fork length (FL) and post-orbital hypural length ( POH ) to the nearest cm, and estimated egg

[^2]retention for all females ( 0 to 5,000 max; visually estimated). All eggs that were not estimated to be within a carcass were assumed to have been successfully deposited. Any female carcass containing an estimated 5,000 eggs were considered a pre-spawn mortality. Forceps were used to remove five scale samples from all natural-origin Chinook. Scales were adhered to desiccant scale cards for preservation and identified by sample number and sample date. At the conclusion of spawning season, scales were sent to WDFW for post-hoc age analysis. Age analysis data were used to assess age-at-return (run-reconstruction), and combined with biological data to assess length-at-age. All Chinook were scanned for passive integrated transponder (PIT) tags and all PIT detections were recorded and later uploaded to PTAGIS. Carcasses were scanned with a T-wand (Northwest Marine Technology, Inc., Shaw Island, WA USA) for coded wire tags (CWT). If present, the snout portion was removed and individually bagged and labeled with species, origin, FL, river of recovery and date. After sampling each carcass, the caudal fin was removed before the carcass was returned to the river to avoid resampling on subsequent surveys. All data collected in the field were input directly into a YUMA rugged computer tablet (Trimble Navigation, Ltd.). Weekly carcass recovery totals were summed post-season to calculate annual carcass recovery totals per reach and per survey area.

Some key assumptions for carcass surveys included:

| Assumption I - | All carcasses had the same probability of being recovered on the <br> spawning grounds (despite differences in sex, origin, size or <br> spawning location) |
| :--- | :--- |
| Assumption II - | The diagnostic unit in which a carcass is recovered is the same as <br> the reach in which the fish spawned |
| Assumption III - | Sampled carcasses are representative of the overall spawning <br> composition within each reach |

## pHOS and PNI

pHOS was first calculated using the straightforward method of calculation for the population-level pHOS by simply dividing the number of hatchery-origin spawners by the total spawners, such that:

$$
p H O S=\frac{H O S_{O}}{\mathrm{HOS}_{O}+\mathrm{NOS}_{O}}
$$

where $\mathrm{HOS}_{O}$ is the total recovered hatchery-origin carcasses and $\mathrm{NOS}_{O}$ is the total recovered natural-origin carcasses. This simple algorithm does not account for assumed deficiencies in hatchery fish effectiveness (i.e. relative reproductive success) nor does it account for spatial variation in pHOS and unequal sampling effort across reaches. For example, reach S1 tends to have a higher pHOS than other reaches because the Similkameen acclimation site is located in the reach. Likewise, the probability of recovering carcasses in low density spawning reaches is
lower than in reaches with high density spawning. We have attempted to account for each of these factors.

Relative reproductive success has not been estimated for summer/fall Chinook in the Okanogan. One of the key assumptions in the In-Season Implementation Tool was that firstgeneration hatchery fish are less effective natural spawners than natural-origin fish. Currently, the hatchery fish effectiveness assumption for the Okanogan population is that first generation hatchery-origin spawners are $80 \%$ as effective as natural-origin fish as contributing genes to the next generation ${ }^{4}$ This assumption is based on research conducted by Reisenbichler and McIntyre (1977) and Williamson et al. (2010). Therefore, the pHOS calculation was amended in 2013 to account for the reduction in hatchery spawner effectiveness, such that:

$$
\text { Effective } p H O S=\frac{0.8 \operatorname{HOS}_{O}}{0.8 \operatorname{HOS}_{O}+\mathrm{NOS}_{O}}
$$

Further refinement of the pHOS calculation was needed to account for non-random sampling of carcasses and variable pHOS across reaches. This was done by weighting each reach's overall contribution to system-wide pHOS according to the overall proportion of summer/fall Chinook redds that occurred within that reach.

First, the proportion of redds that corresponded to each reach was calculated by the equation:

$$
r e d d_{p, r}=\frac{r e d d_{r}}{r e d d_{o}}
$$

where, redd $d_{r}$ is the number of documented redds that occur within reach $r, r e d d_{0}$ is the total number of redds documented in the U.S. portion in the Okanogan River Basin, and redd $p_{p, r}$ is the proportion of total redds that were documented in reach $r$.

Next, Effective pHOS was calculated separately for each sampled reach, $r$, so that:

$$
p H O S_{r}=\frac{0.8 H O S_{r}}{0.8 H O S_{r}+\text { NOS }_{r}}
$$

where $\mathrm{pHOS}_{r}$ is the Effective pHOS calculation for reach $r$, and $\mathrm{HOS}_{r}$ and $\mathrm{NOS}_{r}$ are the total recovered carcasses of hatchery- and natural-origin within that reach. Finally, Effective pHOS was corrected for the proportion of redds in each reach to determine an adjusted Effective pHOS, such that:

$$
\text { Effective } p H O S=\sum_{i=1}^{n} p \mathrm{HOS}_{r}\left(\text { redd }_{p, r}\right)
$$

[^3]where $n$ is the total number of sampled reaches that compose the Okanogan River Basin. These calculations assumed that sampled carcasses were representative of the overall spawning composition within each reach; that no carcasses were washed downstream into another reach; that all carcasses had an equal probability of recovery; and that all fish within origin types had equal fecundity. While it is unlikely that all of these assumptions were correct, the modified calculation results in a better representation of the actual census pHOS.

PNI was calculated as:

$$
P N I=\frac{p N O B}{\text { Effective } p H O S+p N O B}
$$

where $p N O B$ was the proportion of broodstock that were natural-origin Okanogan returns, and Effective pHOS was the reach weighted effective pHOS defined previously. To determine an Okanogan specific pNOB, we applied the results of a radio tracking study, which estimated that $90 \%$ of the natural-origin fish detected near the mouth of the Okanogan River in 2011 and 2012 ended up spawning in the Okanogan Basin (Mann and Snow 2013). Therefore, we assumed that $90 \%$ of the NOB collected in the purse seine (2010-2013) was of Okanogan origin.

In years prior to 2010 all of the broodstock for the Similkameen program were collected at Wells Dam. That program strived for $100 \%$ pNOB and did achieve $>95 \%$ pNOB in 7 of the last 8 years (Hillman et al. 2014). However, the Wells Dam broodstock collection efforts composited natural-origin fish from the Okanogan and Methow populations as well as fish originating from downstream populations ${ }^{5}$. We made a correction for non-Okanogan NOB for all years when Wells Dam was used for brood collection using the formula:

$$
\text { Adjusted Wells Dam pNOB }=\text { Wells Dam pNOB } *\left(\frac{\text { Okanogan NOS }}{\text { Okanogan NOS }+ \text { Methow NOS }}\right)
$$

where the Adjusted Wells Dam pNOB was estimated based on the proportion of natural-origin spawners (NOS) that were in the Okanogan compared to the Methow for that particular year. This correction was made for a portion of the broodstock in 2010 and 2011 and all of the broodstock previous to 2010. This correction did not account for stray NORs from downstream populations or NORs that would have remained in the Columbia River above Wells Dam. Although the radio tracking study provides an estimate of this for 2011 and 2012, there was uncertainty regarding the applicability of the radio tracking data for years prior.

## Origin of Hatchery Spawners

Snouts from adipose fin clipped fish were removed, individually labeled, frozen, and delivered to the Chief Joseph Hatchery coded wire tag lab for CWT extraction and reading. The

[^4]Regional Mark Information System (RMIS; http://www.rmis.org/rmis) was queried in July 2017 to assess the rearing facility of hatchery-origin Chinook recovered on the Okanogan spawning grounds, the in-to-basin stray rate, and the out-of-basin stray rates. RMIS data queries are described in detail in the 2013 CJHP Annual Report (Baldwin et al. 2016)

## Smolt-to-Smolt Survival and Travel Time

Survival and travel time were assessed using the Data Acquisition in Real Time (DART) website analysis tools. DART calculates a survival estimate using a Cormack Jolly Seber mark recapture model, for full details on the analysis methods please see the DART website (http://www.cbr.washington.edu/dart/query/pit sum tagfiles). Each CJH release group with PIT tags were queried for survival from release to Rocky Reach Dam Juvenile bypass (RRJ) and McNary Dam Juvenile bypass (MCN). Although some recaptures were obtained further downstream than McNary Dam, survival through the entire hydropower system to Bonneville Dam could not be generated because there were not enough recaptures downstream to estimate the recapture probability. Survival estimates and travel time for nearby hatcheries and the wild summer Chinook captured in the RST and beach seine were also analyzed for comparison purposes.

Survival estimates are 'apparent survival' because they were not adjusted for residuals, tag failure, tag loss (shedding), or other factors which could result in fish not dying but not being detected at a downstream location. Due to these factors, actual survival would be higher than the apparent survival estimates provided in this report.

Migration timing from release to the lower Okanogan River was determined using a query of the PTAGIS database (https://www.ptagis.org/data/quick-reports/small-scale-sitedetections) to determine the timing of PIT tag detections from releases of Summer Chinook at Omak Pond and Spring Chinook at Riverside Pond. No PIT tags were released from Similkameen Pond in 2016. The lower Okanogan River PIT tag interrogation site (OKL) is located at rkm. 25 and is within 2 km of the inundation effects of Wells Dam.

## Smolt-to-adult Return

The smolt to adult return rate (SAR) was estimated as:

$$
S A R=\frac{\text { expanded CWT recoveries }}{C W T \text { released }}
$$

where expanded CWT recoveries included estimated expanded recoveries on the spawning grounds, at hatcheries and in fisheries. Two expansions were applied. First the number of recoveries was expanded to account for the proportion of the release group that wasn't tagged. For example, with a $99 \%$ CWT mark rate the recoveries would be increased by $1 \%$. Second, the recoveries were expanded based on the proportion of the population that was sampled. For
example, if carcass surveys recovered $20 \%$ of the estimated spawners then the number of CWT recoveries was expanded by $80 \%$. The number of CWT fish released were simply the hatchery release data including all tag codes for CWT released fish (CWT + Ad Clip fish and CWT-only fish).

## Spring Chinook Presence and Distribution

Smolt releases of CJH 10(j) spring-Chinook did not occur in the Okanogan until April 2015. However, occasional occurrence of spring-Chinook was detected in the Okanogan, likely due to straying of individuals from other populations. Pre-reintroduction monitoring for spring Chinook consisted of environmental DNA (eDNA) (see Laramie et al 2015 for details of the eDNA sampling) and PIT tag sampling and analysis at tributary and mainstem Okanogan sites to determine the baseline distribution, prior to the $10(\mathrm{j})$ reintroduction. In order to target springChinook specifically, rather than summer/fall Chinook, eDNA samples were collected in tributary habitats, which are not typically utilized by summer/fall Chinook. Additionally, monitoring programs throughout the Columbia Basin are implanting PIT tags into both hatchery- and natural-origin spring Chinook as juveniles that might stray to the Okanogan as returning adults. The monitoring programs at Bonneville and Wells dams tags returning adult spring Chinook, which greatly increases the probability of encountering a spring Chinook with a PIT tag in the Okanogan. In 2016, the spatial distribution of spring Chinook was evaluated using a combination of eDNA and PIT tag data. We queried the PTAGIS database using an interrogation summary for all PIT detection sites in the Okanogan and Similkameen Rivers, including Canada. Once a list of tag codes was obtained, a second query was run to determine if any of the fish had a final detection outside the Okanogan. 2016 was the first year for jacks (3 year old male) to return from the releases of PIT tagged juveniles as part of the CJH program.

## Coded Wire Tag Lab Analysis

Coded wire tags (CWT) from broodstock, ladder surplus, purse seine harvest, creel and spawning ground surveys were extracted, read, and reported in the Chief Joseph Hatchery Lab from December 2016 to April 2017. The snouts were then interrogated for the presence of a CWT by using a V-reader or T-wand. After positive detection, the snout was cut bilaterally into symmetrical portions keeping the half that indicated detection and discarding the other half into the snout bag from which it came. This process was then repeated until only a small piece of tissue containing the CWT remains. The final piece of tissue was then smeared on a cutting mat exposing the CWT, then placed on its corresponding snout card and finally on to a cafeteria tray (groups of $\sim 25$ tags) to be read under a microscope.

Extracted tags were removed from the tray one-by-one to be cleaned, recorded and read. The CWT was cleaned by wetting a lint free cloth and rolling the tag between a finger and cloth
to remove all remaining tissue. The CWT was attached to a Northwest Marine Technologies (NMT) magnetic pencil and inserted into a jig to be read under a LCD microscope with the aid of an illuminator. Biological data was transcribed from the snout card to a final CWT datasheet. The CWT was attached to this datasheet with tape after the six digit code was read. Information from the datasheet was then transferred to an excel workbook which contains all applicable CWT code combinations.

CWTs were expanded based on their mark and sample rate to estimate total catch contribution for a specific fishery. For each fishery, every CWT recovered was grouped according to the hatchery from which it originated, and then divided by mark and sample rate to expand the number of fish caught. It was common to see many different tag codes from the same hatchery during a given brood year. These tags need to be summed together then multiplied by an adjustment factor to adjust for tags that are present, as well as lost and unreadable. This methodology is applied to all recovered CWTs in a given area, then summed together to represent the total number of fish harvested as well as their place of origin. The following equations are used to expand and adjust for CWT recoveries:

$$
C W T_{\text {Expansion }}=\frac{\text { Total Recovered }}{\text { Tag Rate }} \cdot\left(\frac{1}{\text { Sample Rate }}\right) \cdot(X Y Z)
$$

where Total Recovered was the total number of recovered tags in the fishery, Tag Rate was the proportion of fish that are coded wire tagged in the tag group (expressed as a decimal percentage), Sample Rate was the proportion of the total catch that are sampled for CWTs, and $X Y Z$ was the total adjustment factor calculated from recoveries without a decoded tag.

Coded wire tag recoveries were also adjusted for tags not adhering to the normal recovery process. That process was: the snout arrives at the lab, the tag is extracted from the snout, the tag is readable, and the result matches the tag's release information. Recovery results fall into one of two general categories: with and without decoded tags. They were assigned one of the following CWT result codes after completing the snout recovery to tag reading procedure:

## Recoveries with a decoded tag:

1 Code read from the wire tag matches a tag release and is consistent with release information.
9 The wire tag has no code (blank) or agency-only code.
7 Code read from the wire tag does not match a tag release, or it is not consistent with the release information.

Recoveries without a decoded tag:

4 Tag wire was present, but the code is unreadable.
3 Tag wire was present (dissected from the snout) but was lost prior to the reading process.
2 Tag was not present in snout.
8 Snout was determined in the field to be from a wire-tagged fish, but the snout was not processed in the Tag Recovery Lab.

Decoded recoveries were adjusted for CWT result code 8 (represented as $X$ in the expansion formula), CWT result code 4 (represented as $Y$ in the expansion formula), and CWT result code 3 (represented as $Z$ in the expansion formula). The formulas for these adjustment factors were as follows and are applied to CWT result codes 1,7 and 9 recoveries:
(1) The adjustment factor due to the presence of CWT result code 8 tags is

$$
X=\frac{\operatorname{Tags}_{1}+\mathrm{Tags}_{2}+\mathrm{Tags}_{3}+\mathrm{Tags}_{4}+\mathrm{Tags}_{7}+\text { Tags }_{8}+\mathrm{Tags}_{9}}{\operatorname{Tags}_{1}+\mathrm{Tags}_{2}+\mathrm{Tags}_{3}+\mathrm{Tags}_{4}+\mathrm{Tags}_{7}+\mathrm{Tags}_{9}}
$$

(2) The lost tag recoveries (code 3) are accounted for by the following formula:

$$
Y=\frac{\text { Tags }_{1}+\text { Tags }_{3}+\text { Tags }_{4}+\text { Tags }_{7}+\text { Tags }_{9}}{\operatorname{Tags}_{1}+\text { Tags }_{4}+\text { Tags }_{7}+\text { Tags } 9}
$$

(3) The unreadable tag recoveries (code 4) are accounted for by the following formula:

$$
\mathrm{Z}=\frac{\text { Tags }_{1}+\text { Tags }_{4}+\text { Tags }_{7}+\text { Tags }_{9}}{\text { Tags }_{1}+\text { Tags }_{7}+\text { Tags }_{9}}
$$

(4) Finally, any CWT result code 3, 4, or 8 recovery is assigned an expansion factor of zero, since its presence is accounted for in the decoded recoveries by the above adjustment equations. Decoded tags are not adjusted for CWT result code 2 "no tags".

## RESULTS

## Rotary Screw Traps

The rotary screw traps captured 13,321 Chinook juvenile out migrants, including 661 hatchery- and 12,660 natural-origin. Pulses of high catch rates coincided with periods of increased streamflow in mid- and late-April (Figure 9). The mean length of Chinook increased throughout the trapping season, but the number of natural-origin smolts that were large enough ( $>65 \mathrm{~mm}$ ) to PIT tag was small ( $\mathrm{n}=280$ ) (Figure 10). Unlike previous years, no fish were captured that were likely yearling Chinook. Dorsal fin clips were removed and archived on a portion of tagged fish, including all presumed yearlings, for genetic identification to determine if they were spring or summer/fall Chinook at a future date.

Following Chinook, the next most abundant species captured in the RST was mountain whitefish (

Table 4). Notably, only 16 sockeye were detected, which is far lower than in previous years. Twenty-seven adipose fin present ${ }^{6}$ steelhead and 43 adipose fin absent (hatchery-origin) steelhead were removed from the trap and released immediately into the river. There were seven juvenile steelhead mortalities (three adipose fin present and four adipose clipped) at the trap resulting in a $10 \%$ juvenile trapping and handling mortality rate. The encounter of 43 adipose clipped and 27 adipose present (assumed natural-origin) and mortality of three (3) assumed natural-origin steelhead are within the take limits identified in the authorizing ESA Section 10(a)(1)(A) Permit for the rotary screw trap operation (Permit 16122).


Figure 9. Daily natural-origin sub-yearling Chinook catch within an 8 foot and 5 foot rotary screw trap in the Okanogan River in 2016.

[^5]

Figure 10. Natural-origin sub-yearling Chinook size distribution ( $\mathrm{n}=3,694$ ) from the rotary screw traps on the Okanogan River in 2016. Boxes encompass the $25^{\text {th }}$ to $75^{\text {th }}$ percentiles of measured fish, the mid-line in the box is the median fish length.

Table 4. Number of juvenile fish trapped at the Okanogan River rotary screw traps in 2016.

| Species | Total Trapped |
| :---: | :---: |
| Pumpkinseed sunfish | 2 |
| Bluegill | 82 |
| Bridgelip Sucker | 2 |
| Common Carp | 16 |
| Longnose Dace | 2 |
| Mountain Whitefish | 452 |
| Northern Pikeminnow | 22 |
| Largemouth Bass | 41 |
| Sculpin (Cottus spp.) | 12 |
| Smallmouth Bass | 45 |
| Three Spine Stickleback | 1 |
| Peamouth | 1 |
| Redside shiner | 1 |
| Speckled dace | 1 |
| Bullhead (Ameiurus spp.) | 15 |
| Yellow Perch | 121 |
| Non-salmonid | 816 |
| Adipose Clipped Steelhead | 43 |
| Adipose Present Steelhead | 27 |
| Hatchery Chinook | 661 |
| Sockeye | 16 |
| Wild Chinook Subs | 12,660 |
| Wild Chinook Yearling | 0 |
| Salmonid | 13,402 |

Four efficiency trials were conducted with juvenile Chinook (three with natural-origin subyearlings and one with hatchery-origin yearlings) between 9,920-13,200 cfs (Table 5.). RST efficiency and Okanogan River flow were not correlated in 2016 ( $p=0.78$, adjusted $r^{2}=-0.43$ ). As such, the WDFW smolt abundance calculator was not used.

Table 5. Efficiency trials conducted on natural-origin Chinook sub-yearlings at the Okanogan rotary screw traps in April and May, 2016.

| Trap Date | River Flow @ <br> USGS Malott | Total Chinook <br> Marked and <br> Released | Age Class / <br> Origin | Total Chinook <br> Recaptured | Trap <br> Efficiency |
| :---: | :---: | :---: | :---: | :---: | :---: |
| $4 / 13$ | 10,400 | 1,410 | $1+/$ Hatchery | 3 | 0.002 |
| $5 / 5$ | 13,400 | 1,075 | $0+/$ Wild | 0 | 0.00 |
| $5 / 11$ | 12,200 | 1,482 | $0+/$ Wild | 4 | 0.003 |
| $5 / 17$ | 9,730 | 1,615 | $0+/$ Wild | 0 | 0.00 |
| Total |  | $\mathbf{1 , 7 3 3}$ | $\mathbf{1 , 7 3 3}$ | $\mathbf{1 0}$ | $\mathbf{0 . 0 0 1}$ |



Figure 11. Okanogan River CFS was not predictive of RST efficiency, and so was excluded as a variable from juvenile production estimates. Efficiency trials conducted with natural-origin subyearlings are marked in blue, trials conducted with hatchery-origin yearlings are marked in red.

An efficiency trial was conducted on April 13 using hatchery yearlings from the Omak Pond (Table 6.). In the past, such trials have been used to explore the possibility of using hatchery-origin yearlings as a surrogate for natural-origin subyearlings, but significant differences in capture efficiency ultimately led to the abandonment of this idea. Nevertheless, a trial with hatchery-origin yearlings was conducted in 2016. Three hatchery-origin yearlings out of 1,410 released were recaptured ( $.21 \%$ efficiency).

Since streamflow did not affect trapping efficiency, efficiency trials were pooled to calculate overall trap efficiency for both natural- and hatchery-origin fish (Table 7). Overall efficiency estimates for natural- and hatchery-origin fish were low as were total catches, leading to a relatively imprecise estimate of total emigration (Table 8).

Table 6. Efficiency trial conducted on hatchery-origin Chinook smolts at the Okanogan rotary screw traps in 2016

| Trap Date | River Flow <br> @ USGS <br> Malott | Total <br> Chinook <br> Marked | Total <br> Chinook <br> Released | Total <br> Chinook <br> Recaptured | Trap <br> Efficiency |
| :---: | :---: | :---: | :---: | :---: | :---: |
| $4 / 13$ | 10,400 | 1,410 | 1,410 | 3 | 0.02 |
| Total |  | 1,410 | $\mathbf{1 , 4 1 0}$ | 3 | $\mathbf{0 . 0 2}$ |

Table 7. Pooled efficiency trail results for all trap configurations. Whenever fish were released, each trap was operational. Efficiency was calculated based on recaptures for each individual trap, as well as the combined efficiency of both traps.

|  |  | Mark-Released | Recaptured | Efficiency |
| :---: | :---: | :---: | :---: | :---: |
| 2.4 m Trap | Hatchery Chinook | 1,410 | 1 | 0.071\% |
|  | Wild Chinook | 4,172 | 3 | 0.072\% |
| 1.5 m Trap | Hatchery Chinook | 1,410 | 2 | 0.14\% |
|  | Wild Chinook | 4,172 | 1 | 0.024\% |
| Combined Traps | Hatchery Chinook | 1,410 | 3 | 0.213\% |
|  | Wild Chinook | 4,172 | 4 | 0.096\% |

Table 8. Population estimates for hatchery- and natural-origin juvenile Chinook salmon in the Okanogan River Basin.

| Species | Population <br> Estimate | Lower 95\% <br> Confidence Interval | Upper 95\% <br> Confidence Interval |
| :--- | :--- | :--- | :--- |
| Hatchery-origin <br> Chinook* | 233,520 | 29,739 | 437,300 |
| Natural-origin <br> Chinook | $10,566,870$ | $2,118,348$ | $19,015,391$ |

* A total of 867,679 hatchery-origin Chinook were released into the Okanogan River system upriver from the screw trap site in 2016. 203,311 were released from the Riverside acclimation pond from April 15-21; 256,294 were released from the Similkameen hatchery from April 15 - April 22; 232,353 were released from the Omak acclimation pond on April 15-22; and 175,721 were released from Omak Pond from May 23-26.


## Juvenile Beach Seine and Pit Tagging

In 2016, 15,802 natural-origin juvenile salmonids were collected in 202 sets for a total catch per unit effort of 78 per seine haul (Table 9). Out of the 15,802 juvenile salmonids collected, 13,650 (86\%) sub-yearling Chinook were PIT tagged and released (Figure 12). Preand post-tag mortality was $2.0 \%$ and $3.9 \%$ respectively. Seventy-three shed tags were recovered from the net pens prior to release. Thirteen of the sheds were from post-tag mortalities, the other 60 were ejected from fish that were later released alive, but without a tag. All recovered tags were removed from the tagging file before upload to PTAGIS. Fish size increased through time (Figure 13), but the number of available fish to tag decreased (Table 9). By mid-June, Columbia River temperatures had risen to above $14^{\circ} \mathrm{C}$. We suspect that subyearling Chinook may have migrated downstream, or to deeper, cooler water making it difficult to collect them via beach seine, as has presumably happened in past years. Overall size distribution for tagged fish was skewed towards smaller size ranging from 47-122 mm in length (Figure 14.). Bycatch included hatchery-origin juvenile Chinook, three-spine stickleback, carp, mountain whitefish, pumpkinseed, redside shiner, smallmouth bass, dace species, and sculpin species.

Table 9. Summary of juvenile Chinook beach seining effort at Gebber's Landing (Geb.) and Washburn Island (W.I.) in 2016. CPUE represents total Chinook per set.

| Week start | Geb. <br> sets | Geb. collected | Geb. CPUE <br> (total/set) | W.I. <br> sets | W.I. <br> collected | W.I. CPUE <br> (total/sets) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $5 / 8 / 2016$ | 1 | 13 | 13.0 | 0 | 0 | - |
| $5 / 15 / 2016$ | 2 | 198 | 99.0 | 0 | 0 | - |
| $5 / 22 / 2016$ | 2 | 256 | 128.0 | 0 | 0 | - |
| $5 / 29 / 2016$ | 24 | 4729 | 197.0 | 0 | 0 | - |
| $6 / 5 / 2016$ | 25 | 7606 | 304.2 | 0 | 0 | - |
| $6 / 12 / 2016$ | 55 | 558 | 10.1 | 9 | 50 | 5.6 |
| $6 / 19 / 2016$ | 21 | 936 | 44.6 | 24 | 1411 | 58.8 |
| $6 / 26 / 2016$ | 25 | 44 | 1.8 | 4 | 1 | .25 |
| $7 / 3 / 2017$ | 10 | 0 | 0.0 | 0 | 0 | - |
| Total | $\mathbf{1 6 5}$ | $\mathbf{1 4 , 3 4 0}$ | $\mathbf{1 0 7}$ | $\mathbf{3 7}$ | $\mathbf{1 , 4 6 2}$ | $\mathbf{5 8}$ |
| Mean | $\mathbf{1 8}$ | $\mathbf{1 , 5 9 3}$ | $\mathbf{8 7}$ | $\mathbf{4}$ | $\mathbf{1 6 2}$ | $\mathbf{4 0}$ |



Figure 12 Total mortality and number of released natural-origin sub-yearling Chinook in 2016.


Figure 13. Size distribution of PIT tagged juvenile Chinook by release date from the beach seine effort in 2016. Boxes encompass the $25^{\text {th }}$ to $75^{\text {th }}$ percentiles of measured fish, the mid-line in the box is the median fish length.


Figure 14. Size distribution of natural origin sub-yearling Chinook tagged during the beach seining effort in 2016

1,378 PIT tagged juvenile Chinook were detected at the Rocky Reach juvenile bypass system, which was $10.1 \%$ of total fish tagged and released. Eighty ( $0.6 \%$ ), 31 ( $0.2 \%$ ) and 16 ( $0.1 \%$ ) were detected at the McNary, John Day and Bonneville Dams respectively. Detections for sub-yearlings occurred primarily in July at all downriver dams (Figure 15). There were no detections in the lower river dams from September through December.

b) McNary Passage

c) John Day Passage



Figure 15. Daily distribution of detections of PIT-tagged sub-yearling Chinook at Rocky Reach, McNary, John Day, and Bonneville Dams in 2016. Note differences in scale. The y axes denote the numbers of PIT-tagged fish encountered daily at each of the mainstem project arrays.

Travel time from release to Rocky Reach Dam was the slowest compared to travel time from release to the other lower river dams (Table 10). Larger fish travelled faster to Rocky Reach Dam (Figure 16). This is similar to what was reported in 2011-2013 by Douglas County PUD.

Table 10. Mean travel time (d) and rate (km/d) for PIT tagged sub-yearling Chinook released near Gebber's Landing and detected at Columbia River dam PIT arrays.

|  | Rocky Reach(762) |  | McNary (470) |  | John Day (347) |  | Bonneville(235) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Location (River KM) | Travel <br> Time <br> (d) | Rate $(\mathrm{km} / \mathrm{d})$ | Travel Time <br> (d) | Rate $(\mathrm{km} / \mathrm{d})$ | Travel <br> Time <br> (d) | Rate $(\mathrm{km} / \mathrm{d})$ | Travel <br> Time <br> (d) | Rate $(\mathrm{km} / \mathrm{d})$ |
| Release (856) | $\begin{aligned} & 28.8(\mathrm{SE} \\ & =0.33 ; \\ & \mathrm{n}=1380) \end{aligned}$ | $3.3$ | $\begin{array}{\|l} 36.1 \\ \text { (SE = } \\ 1.03 ; \\ \mathrm{n}=85) \end{array}$ | $10.7$ | $\begin{aligned} & 40.3(\mathrm{SE} \\ & =1.71 ; \\ & \mathrm{n}=31) \end{aligned}$ | $12.6$ | $\begin{aligned} & 36.2 \\ & \text { (SE= } \\ & 2.32 ; \\ & \mathrm{n}=17) \end{aligned}$ | $17.5$ |
| Rocky <br> Reach <br> (762) |  |  | $\begin{array}{\|l} 10.5 \\ (\mathrm{SE}= \\ 0.55 ; \\ \mathrm{n}=34) \end{array}$ | $27.8$ | $\begin{aligned} & 14.8 \\ & \mathrm{SE}= \\ & 1.73 ; \\ & \mathrm{n}=11) \end{aligned}$ | 28.0 |  |  |



Figure 16. Fish size (fork length) and travel time of tagged Chinook to Rocky Reach Dam.

## Lower Okanogan Adult Fish Pilot Weir

The Okanogan River (at Malott) discharge was below normal in 2016 and was below $1,800 \mathrm{cfs}$ for the trapping season. Staff were able to safely enter the river and begin installation on August 8 when discharge was 1,530 cfs (Figure 17). Discharge continued to drop throughout the installation period until August 31 when levels increased rapidly up to 1,760 cfs due to Okanagan water management objectives) before dropping again on September 6. After September 9, discharge stayed between 800-1,000 cfs for the rest of the season.

Migration of Sockeye and summer Chinook is generally affected by a thermal barrier that is caused by warm water temperatures ( $\geq \sim 22^{\circ} \mathrm{C}$ ) in the lower Okanogan River. The thermal barrier is dynamic within and between years, but generally it sets up in mid-July and breaks down in late August. In some years, the Okanogan River will temporarily cool off due to a combination of interrelated weather factors including rainstorms, cool weather, cloud cover or wildfire smoke. This 'break' in the thermal barrier can allow a portion of the fish holding in the Columbia River to enter the Okanogan and migrate up to thermal refuge in the Similkameen River or Lake Osoyoos. In 2016, temperatures were similar to, though occasionally higher than the median daily temperatures from the last 10 years (Figure 18).

Temperature was above $22.5^{\circ} \mathrm{C}$ on July 1 when flow was $3,530 \mathrm{cfs}$. Temperatures stayed above $22.5^{\circ} \mathrm{C}$ until July 4, when they began to drop and stayed below $22.5^{\circ} \mathrm{C}$ until July 24 . From July 24 to August 3 temperatures increased above $22.5^{\circ} \mathrm{C}$. Temperature varied again between $26.6^{\circ} \mathrm{C}$ and $20.6^{\circ} \mathrm{C}$ from August 3-30. After August 30 temperatures stayed below $22.5^{\circ} \mathrm{C}$ for the rest of the season.


Figure 17. Discharge of the Okanogan River between July 1 and October 31, 2016. This figure was copied directly from the USGS website (http://nwis.waterdata.usgs.gov/wa).


Figure 18. Temperature of the Okanogan River between July 1 and October 31, 2016. This figure was copied directly from the USGS website (http://nwis.waterdata.usgs.gov/wa).

Dissolved Oxygen varied from 5.7 to $10.1 \mathrm{mg} / \mathrm{L}$, total dissolved solids varied from $120-141 \mathrm{ppm}$ and turbidity varied from 0.7 and 4.6 NTUs (Table 11). The head differential ranged from $0-3 \mathrm{~cm}$ across the weir panels (Table 12). The maximum water velocity measured was 2.7 ft ./sec. (Table 13).

Table 11. Water quality data at or near the lower Okanogan weir in 2016. Temperature and discharge were taken from the USGS gauge at Malott.

| Date | Trap Depth (ft.) | Dissolved Oxygen (mg/L) | Total Dissolved Solids (ppm) | Turbidity (NTU) |
| :---: | :---: | :---: | :---: | :---: |
| 8/18 | 2.2 | 6.3 | 131 | 1.1 |
| 8/19 | 2.2 | 6.4 | 128 | 1.8 |
| 8/22 | 2.1 | 5.8 | 132 | 1.1 |
| 8/23 | 2.1 | 5.8 | 132 | 1.1 |
| 8/24 | 2.1 | 5.8 | 133 | 1.1 |
| 8/25 | 2.0 | 6.1 | 137 | 1.0 |
| 8/26 | 2.0 | 6.2 | 137 | 1.0 |
| 8/27 | 2.0 | 6.5 | 139 | 0.8 |
| 8/28 | 2.0 | 5.7 | 134 | 1.3 |
| 8/29 | 2.0 | 5.8 | 135 | 1.7 |
| 8/30 | 2.0 | 7.4 | 141 | 1.1 |
| 8/31 | 2.0 | 9.7 | 141 | 1.1 |
| 9/1 | 2.3 | 8.0 | 130 | 1.7 |
| 9/2 | 2.3 | 9.0 | 124 | 1.9 |
| 9/3 | 2.4 | 9.8 | 126 | 2.1 |
| 9/4 | 2.5 | 9.3 | 128 | 2.2 |
| 9/5 | 2.7 | 9.8 | 122 | 4.6 |
| 9/6 | 2.6 | 10.0 | 125 | 3.5 |
| 9/7 | 2.4 | 10.1 | 126 | 4.3 |
| 9/8 | 2.5 | 9.5 | 125 | 4.4 |
| 9/9 | 2.2 | 7.9 | 123 | 1.8 |
| 9/10 | 2.1 | 7.8 | 128 | 1.9 |
| 9/11 | 2.1 | 8.0 | 129 | 1.3 |
| 9/12 | 2.1 | 7.8 | 124 | 1.0 |
| 9/13 | 2.2 | 8.6 | 122 | 1.0 |
| 9/14 | 2.1 | 8.2 | 120 | 3.0 |
| 9/15 | 2.1 | 9.2 | 125 | 1.6 |
| 9/16 | 2.1 | 8.6 | 129 | 0.8 |
| 9/19 | 2.1 | 7.5 | 127 | 1.2 |
| 9/20 | 2.1 | 8.7 | 125 | 0.7 |
| 9/21 | 2.1 | 8.7 | 124 | 0.8 |
| 9/22 | 2.1 | 8.0 | 123 | 1.2 |
| Min | 2.0 | 5.7 | 120 | 0.7 |
| Max | 2.7 | 10.1 | 141 | 4.6 |

Table 12. Head differential across the different picket spacings. If differential exceeded 10 cm , pickets were cleaned immediately. Measurements are in cm . Daily mean gage height is included in feet. Gage height is copied directly from the USGS website (http://nwis.waterdata.usgs.gov/wa).

| Date | $\mathbf{1 . 0 "}$ <br> Picket <br> (cm) | $\mathbf{1 . 5 "}$ <br> Picket <br> Spacing <br> (cm) | $\mathbf{2 . 0 "}$ <br> Picket <br> Spacing <br> (cm) | $\mathbf{2 . 5 "}$ <br> Picket <br> Spacing <br> (cm) | 3.0" <br> Picket <br> Spacing <br> (cm) | Gage <br> Height <br> (ft.). |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $8 / 18$ | 0.5 | 1.0 | 1.0 | 0.5 | 0.0 | 3.1 |
| $8 / 19$ | 0.5 | 1.0 | 1.0 | 0.5 | 0.0 | 3.1 |
| $8 / 22$ | 1.0 | 1.0 | 1.0 | 1.0 | 0.0 | 3.1 |
| $8 / 23$ | 0.5 | 1.5 | 1.0 | 1.5 | 0.0 | 3.0 |
| $8 / 25$ | 1.0 | 1.5 | 1.0 | 1.0 | 0.0 | 3.0 |
| $8 / 30$ | 1.5 | 0.0 | 0.5 | 1.0 | 0.0 | 3.0 |
| $8 / 31$ | 1.0 | 1.5 | 0.5 | 0.5 | 0.0 | 3.0 |
| $9 / 1$ | 1.0 | 1.0 | 1.0 | 2.0 | 0.0 | 2.9 |
| $9 / 2$ | 3.0 | 1.0 | 1.0 | 1.0 | 0.0 | 2.9 |
| $9 / 3$ | 1.0 | 1.0 | 2.0 | 3.0 | 0.0 | 2.9 |
| $9 / 5$ | 1.0 | 1.0 | 1.0 | 2.0 | 0.0 | 2.9 |
| $9 / 6$ | 1.0 | 1.0 | 2.0 | 1.0 | 0.0 | 2.9 |
| $9 / 7$ | 2.0 | 2.0 | 2.0 | 2.0 | 0.0 | 2.9 |
| $9 / 8$ | 0.5 | 0.5 | 1.0 | 1.5 | 0.0 | 3.0 |
| $9 / 9$ | 1.5 | 1.5 | 1.5 | 1.0 | 0.0 | 3.3 |
| $9 / 10$ | 1.5 | 1.5 | 1.0 | 1.0 | 0.0 | 3.4 |
| $9 / 11$ | 1.0 | 1.5 | 1.0 | 1.0 | 0.0 | 3.5 |
| $9 / 12$ | 1.5 | 1.5 | 1.0 | 1.0 | 0.0 | 3.6 |
| $9 / 13$ | 1.5 | 1.5 | 1.0 | 1.0 | 0.0 | 3.9 |
| $9 / 14$ | 1.0 | 1.0 | 1.0 | 0.5 | 0.0 | 3.7 |
| $9 / 15$ | 2.0 | 2.0 | 2.0 | 1.0 | 0.0 | 3.7 |
| $9 / 19$ | 1.0 | 1.0 | 0.5 | 0.0 | 0.0 | 3.6 |
| $9 / 20$ | 1.0 | 1.0 | 0.5 | 1.0 | 0.0 | 3.2 |
| $9 / 21$ | 1.0 | 1.0 | 1.0 | 1.0 | 0.0 | 3.1 |
| $9 / 22$ | 1.0 | 1.0 | 1.0 | 1.0 | 0.0 | 3.1 |
| Min | 0.5 | 0.0 | 0.5 | 0.0 | 0.0 | 2.9 |
| Max | 3.0 | 2.0 | 2.0 | 3.0 | 0.0 | 3.9 |

Table 13. Water velocity upstream (US) and downstream (DS) of the weir and in the trap. Velocity should not exceed 3.5 ft . /sec. Measurements are in ft . /sec.

| Date | River <br> Left US | Center <br> US | River <br> Right US | River Left <br> DS | Center <br> DS | River <br> Right DS | Trap <br> Velocity |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $8 / 18$ | 2.3 | 1.5 | 1.3 | 1.3 | 2.0 | 2.6 | 1.3 |
| $8 / 19$ | 1.6 | 1.3 | 2.3 | 1.7 | 1.8 | 1.8 | 0.7 |
| $8 / 22$ | 1.7 | 1.2 | 1.4 | 1.0 | 1.7 | 1.2 | 0.5 |
| $8 / 23$ | 1.9 | 0.9 | 1.3 | 1.8 | 2.0 | 1.1 | 0.6 |
| $8 / 24$ | 1.9 | 0.9 | 1.2 | 1.0 | 1.1 | 0.9 | 0.9 |
| $8 / 25$ | 1.6 | 0.6 | 0.9 | 0.6 | 0.8 | 1.6 | 0.9 |
| $8 / 26$ | 1.8 | 0.8 | 1.1 | 0.8 | 0.5 | 1.6 | 1.9 |
| $8 / 27$ | 2.0 | 0.9 | 0.9 | 0.7 | 1.9 | 1.3 | 1.2 |
| $8 / 28$ | 2.0 | 0.7 | 0.6 | 0.6 | 1.6 | 1.4 | 0.3 |
| $8 / 29$ | 2.0 | 1.1 | 1.2 | 1.3 | 1.7 | 2.1 | 0.5 |
| $8 / 30$ | 1.3 | 1.3 | 1.5 | 2.0 | 1.7 | 1.3 | 0.2 |
| $8 / 31$ | 1.8 | 1.0 | 1.3 | 1.0 | 1.3 | 1.9 | 0.2 |
| $9 / 1$ | 1.1 | 1.4 | 1.5 | 1.5 | 1.8 | 2.5 | 0.3 |
| $9 / 2$ | 2.3 | 1.5 | 1.4 | 2.2 | 1.8 | 0.7 | 1.0 |
| $9 / 3$ | 2.3 | 1.0 | 1.3 | 1.3 | 1.7 | 1.8 | 0.5 |
| $9 / 5$ | 2.5 | 1.5 | 1.9 | 1.3 | 2.1 | 2.3 | 1.3 |
| $9 / 6$ | 2.3 | 1.4 | 1.3 | 1.4 | 1.9 | 1.9 | 0.2 |
| $9 / 8$ | 1.7 | 1.4 | 1.6 | 1.5 | 1.7 | 2.5 | 0.1 |
| $9 / 9$ | 2.2 | 1.0 | 1.6 | 1.6 | 1.1 | 1.0 | 1.4 |
| $9 / 10$ | 1.8 | 1.1 | 1.4 | 1.9 | 1.1 | 0.2 | 1.2 |
| $9 / 11$ | 2.7 | 1.2 | 1.4 | 0.6 | 1.4 | 2.0 | 0.8 |
| $9 / 12$ | 2.0 | 1.3 | 1.2 | 1.0 | 1.3 | 1.9 | 1.6 |
| $9 / 13$ | 1.7 | 1.1 | 1.6 | 1.2 | 1.4 | 1.0 | 1.3 |
| $9 / 14$ | 1.9 | 1.0 | 1.0 | 1.1 | 1.5 | 1.8 | 0.7 |
| $9 / 15$ | 1.9 | 0.7 | 1.0 | 0.9 | 1.3 | 1.6 | 1.6 |
| $9 / 16$ | 2.0 | 1.0 | 1.1 | 1.0 | 1.4 | 2.0 | 1.4 |
| $9 / 19$ | 1.8 | 1.2 | 1.6 | 0.7 | 1.8 | 1.6 | 1.3 |
| $9 / 20$ | 1.8 | 0.7 | 1.1 | 0.1 | 1.2 | 1.2 | 1.2 |
| $9 / 21$ | 1.9 | 1.2 | 1.3 | 0.5 | 0.9 | 1.3 | 0.9 |
| $9 / 22$ | 1.9 | 1.0 | 1.0 | 0.4 | 1.2 | 1.0 | 0.8 |
| Min | 1.1 | 0.6 | 0.6 | 0.1 | 0.5 | 0.2 | 0.1 |
| Max | 2.7 | 1.5 | 2.3 | 2.2 | 2.1 | 2.6 | 1.9 |
|  |  |  |  |  |  |  |  |

Fifty four dead fish were removed from the weir between August 17 and September 22 (Table 14). Chinook Salmon were the most commonly encountered species (74\%). There were no steelhead mortalities removed from the weir in 2016. All mortalities were impinged on the upstream side of weir indicating that they had most likely died upstream and floated down onto the weir. The majority of the Chinook carcasses were observed a week before the majority of Chinook were encountered in the trap (Figure 19). There were also no observations of fish caught between pickets in a head upstream direction, which would have indicated that a fish got stuck and died while trying to push through the pickets.

Table 14. Date and species of fish mortalities observed at the lower Okanogan fish weir in 2016. All fish mortalities were considered "wash downs" and collected on the upstream panels of the weir.

| Date | $\begin{gathered} \text { Black } \\ \text { Bullhead } \end{gathered}$ | Chinook | Mountain Whitefish | Smallmouth Bass | Sockeye | Unknown Sucker |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 08/17 |  | 1 |  |  |  |  |
| 08/22 |  | 2 |  |  |  |  |
| 08/23 |  | 2 |  |  |  |  |
| 08/24 |  | 3 |  |  |  |  |
| 08/25 |  | 6 |  |  |  |  |
| 08/26 |  | 6 |  |  |  |  |
| 08/27 |  | 1 |  |  |  |  |
| 08/28 |  | 1 |  |  |  |  |
| 08/29 |  | 5 |  |  |  | 1 |
| 09/01 |  | 1 |  |  |  |  |
| 09/02 |  | 1 |  |  |  |  |
| 09/04 |  | 1 |  |  |  |  |
| 09/05 | 1 |  | 2 | 1 |  | 1 |
| 09/06 |  | 1 |  |  |  |  |
| 09/07 |  | 3 |  |  | 1 |  |
| 09/08 |  |  |  |  | 1 |  |
| 09/10 |  | 1 | 1 |  |  |  |
| 09/12 |  | 2 |  |  |  |  |
| 09/13 |  |  |  |  | 1 | 1 |
| 09/15 |  |  |  | 1 | 1 |  |
| 09/20 |  | 1 |  |  |  |  |
| 09/22 |  | 2 |  |  |  |  |
| Total | 1 | 40 | 3 | 2 | 4 | 4 |



Figure 19. Total number of Chinook trapped and total number of Chinook carcasses collected off the weir panels. The majority of the Chinook carcasses occurred a week before most Chinook were encountered in the trap.

Tower observations showed that most fish were equally distributed across the river, milling in the river right, left and center sections (looking downstream). Estimates were highest during the last week of August and the first week of September when river temperatures were below $22.5^{\circ} \mathrm{C}$. Bank observations showed that the number fish observed holding in the lower pool, 0.8 km below the weir, was highest about four days after the thermal barrier breakdown (Figure 20. ). Trapping operations were conducted after August 22nd when river temperature was $\leq 22.5^{\circ} \mathrm{C}$. The total fish trapped at the weir in 2016 was 190 with $89 \%$ of them being Chinook Salmon (Figure 21). Seventy percent of the Chinook trapped were released back into the river (Figure 22). Two steelhead were trapped between $9 / 10-9 / 20$ and released in good condition within 30 minutes of observation. The TOG was notified when steelhead were trapped, including the total number, origin and condition after release. To reduce handling of fish, trap attendants opened the gate of the crowder and the upstream gate of the trap to allow for complete passage. Fish that were passed upstream were classified as having a vigorous condition, swimming away unharmed.

Prior to collecting broodstock we conducted several tests using hatchery origin fish brought to the weir via the hatchery truck. Eight adults and two jacks were manually carried to the weir in a boot (inner-tube) which took approximately 90 seconds each. These fish were sent back to the hatchery truck via the Whooshh ${ }^{\text {TM }}$ transport tube in approximately eight seconds. This process was repeated three times over a several hour period. There were no immediate mortalities, although two fish were dead upon arrival at the hatchery that
evening (nearly 10 hours after their initial capture in the hatchery ladder). This test was considered a success and the mortalities were not a concern because the fish had been handled much more extensively than the natural origin brood would be.

Sixteen natural origin Chinook were transported to the hatchery and held in the broodstock ponds concurrently with the fish taken for broodstock from the purse seine and hatchery ladder. Adult Chinook were transported from the weir trap to the hatchery brood truck via the Whooshh ${ }^{\text {TM }}$ fish transport system. No immediate mortalities were observed related to the Whooshh ${ }^{\text {TM }}$ transport system or during transport to the hatchery. There were zero mortalities of the Whooshh ${ }^{\text {TM }}$ transported fish by September 15 th (1-6 days). On September $15^{\text {th }}, 70$ natural-origin brood fish collected via a beach seine on the Similkameen River were added to the raceway with the weir collected fish so we could not evaluate longer term pre-spawn mortality separately. The mortality of the combined group was through the entire pre-spawn holding period was $26.7 \%$. The overall pre-spawn mortality of all summer Chinook broodstock was 13.4\%; however the majority of these fish were collected earlier and handled differently, making direct comparisons problematic.



Figure 20. Estimate of Chinook observed from the bank at the lower pool, 0.8 km downstream of the weir.


Figure 21. Total number of fish trapped at the Okanogan weir in 2016.


Figure 22. Final destination of Chinook adults captured in the weir trap during trapping operations in 2016.

In 2016, 0.014 of total spawning escapement was detected in the trap (i.e., weir efficiency) (Table 15). The potential weir effectiveness (if we had been removing all of the HOR encountered) was 0.016.

Table 15. The number of hatchery and natural origin Chinook Salmon encountered at the lower Okanogan weir in 2016. Weir efficiency and effectiveness were metrics for evaluating the potential for the weir to contribute to the CJHP population management goals in the future.

| Survey <br> Year | Chinook Adults Encountered in the Weir Trap |  | Chinook Spawning <br> Escapement Estimates ${ }^{\text {c,d }}$ |  | Weir Metrics |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Natural Origin (NOR) | Hatchery Origin (HOR) | Natural Origin (NOS) | Hatchery Origin (HOS) | Weir Efficiency ${ }^{\text {a }}$ | Weir Effectiveness ${ }^{\text {b }}$ |
| 2013 | 73 | 18 | 5,627 | 2,567 | 0.010 | 0.006 |
| 2014 | 2,006 | 318 | 10,402 | 1,762 | 0.147 | 0.138 |
| 2015 | 35 | 19 | 10,350 | 3,398 | 0.004 | 0.005 |
| 2016 | 135 | 34 | 8,661 | 1,944 | 0.014 | 0.016 |

${ }^{\text {a }}$ Estimates for weir efficiency are adjusted for prespawn mortality and include Chinook adults that are harvested, released, and collected for brood.
${ }^{\mathrm{b}}$ Estimates for weir effectiveness are adjusted for prespawn mortality and include Chinook adults that are harvested or removed for pHOS management.
c Estimates do not include Chinook Zosel Dam counts.
${ }^{\text {d }}$ NOS and HOS estimates determined by 'reach-weighted' pHOS calculations

## Redd Surveys

In 2016, 5,276 summer/fall Chinook redds were counted in the Okanogan and Similkameen rivers using a combination of ground and aerial surveys (Table 16. Figure 23). The number of redds counted in 2016 was higher than the long-term or more recent 5-year average (Table 16). The majority of Chinook redds were located in 06 (44.3\%) and S1 (31.2\%). The overall redd distribution across the reaches was similar to previous years with the majority of spawning taking place in the upper Okanogan reaches ( 05 and 06) and lower Similkameen (S1) (Table 17, Figure 24).

Estimated spawning escapement was 10,605 ( 5,276 redds $\times 2.01$ fish per redd) (Table 18). During the survey period 1989 through 2016, the summer/fall Chinook spawning escapement within the U.S. portion of the Okanogan River Basin averaged 5,872 and ranged from 473 to 13,857 (Table 18).

The majority of summer/fall Chinook redds were counted during spawning ground surveys between October 3-30 (Table 19). No spawning ground surveys were conducted after November 6.

Table 16. Total number of redds counted in the Okanogan River Basin, 1989-2016 and the averages for the total time series and the most recent 5-year period.

| Survey <br> Year | Number of summer Chinook redds |  |  |
| :---: | :---: | :---: | :---: |
|  | Okanogan River | Similkameen River | Total Count |
| 1989 | 151 | 370 | 521 |
| 1990 | 99 | 147 | 246 |
| 1991 | 64 | 91 | 155 |
| 1992 | 53 | 57 | 110 |
| 1993 | 162 | 288 | 450 |
| 1994 | 375* | 777 | 1,152 |
| 1995 | 267* | 616 | 883 |
| 1996 | 116 | 419 | 535 |
| 1997 | 158 | 486 | 644 |
| 1998 | 88 | 276 | 364 |
| 1999 | 369 | 1,275 | 1,644 |
| 2000 | 549 | 993 | 1,542 |
| 2001 | 1,108 | 1,540 | 2,648 |
| 2002 | 2,667 | 3,358 | 6,025 |
| 2003 | 1,035 | 378 | 1,413 |
| 2004 | 1,327 | 1,660 | 2,987 |
| 2005 | 1,611 | 1,423 | 3,034 |
| 2006 | 2,592 | 1,666 | 4,258 |
| 2007 | 1,301 | 707 | 2,008 |
| 2008 | 1,146 | 1,000 | 2,146 |
| 2009 | 1,672 | 1,298 | 2,970 |
| 2010 | 1,011 | 1,107 | 2,118 |
| 2011 | 1,714 | 1,409 | 3,123 |
| 2012 | 1,613 | 1,066 | 2,679 |
| 2013 | 2,267 | 1,280 | 3,547 |
| 2014 | 2,231 | 2,022 | 4,253 |
| 2015 | 2,379 | 1,897 | 4,276 |
| 2016 | 3,486 | 1,790 | 5,276 |
| Average | 1,191 | 1,050 | 2,179 |
| $\begin{gathered} 5-y r \\ \text { Average } \end{gathered}$ | 2,395 | 1,611 | 4,006 |



Figure 23. Distribution of summer/fall Chinook redds in 2016. Individual redds are identified by red circles. Horizontal coordinate information are referenced to the North American Datum of 1983 (NAD 83).

Table 17. Annual and average abundance of summer/fall Chinook redds in each reach of the Okanogan (01-06) and Similkameen (S1-S2) Rivers from 2006-2016.

| Return <br> Year | Okanogan |  |  |  |  |  |  |  | O1 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\mathbf{O - 2}$ | $\mathbf{O - 3}$ | $\mathbf{O - 4}$ | $\mathbf{O - 5}$ | O-6 | S-1 | S-2 |  |  |
|  | 10 | 56 | 175 | 145 | 840 | 1366 | 1277 | 405 | 4274 |
| 2007 | 3 | 16 | 116 | 63 | 549 | 554 | 624 | 86 | 2011 |
| 2008 | 4 | 51 | 59 | 96 | 374 | 561 | 801 | 199 | 2145 |
| 2009 | 3 | 32 | 91 | 138 | 619 | 787 | 1091 | 207 | 2968 |
| 2010 | 9 | 58 | 67 | 89 | 357 | 431 | 895 | 212 | 2118 |
| 2011 | 3 | 20 | 101 | 55 | 593 | 942 | 1217 | 192 | 3123 |
| 2012 | 12 | 54 | 159 | 68 | 555 | 765 | 914 | 152 | 2679 |
| 2013 | 3 | 2 | 158 | 46 | 397 | 1661 | 1254 | 26 | 3547 |
| 2014 | 11 | 57 | 191 | 111 | 851 | 1010 | 1737 | 285 | 4253 |
| 2015 | 36 | 113 | 284 | 79 | 1008 | 859 | 1611 | 286 | 4276 |
| 2016 | 2 | 57 | 52 | 130 | 907 | 2338 | 1645 | 145 | 5276 |
| Average | $\mathbf{9}$ | $\mathbf{4 7}$ | $\mathbf{1 3 2}$ | $\mathbf{9 3}$ | $\mathbf{6 4 1}$ | $\mathbf{1 0 2 5}$ | $\mathbf{1 1 8 8}$ | $\mathbf{2 0 0}$ | $\mathbf{3 3 3 4}$ |



Figure 24. Proportion of redds in each reach of the Okanogan and Similkameen Rivers from 2006 to 2016.

Table 18. Spawning escapements for summer/fall Chinook in the Okanogan and Similkameen Rivers for return years 1989-2016.

| Return Year | Fish/Redd Ratio | Spawning Escapement |  |  |
| :---: | :---: | :---: | :---: | :---: |
|  |  | Okanogan | Similkameen | Total |
| 1989* | 3.3 | 498 | 1,221 | 1,719 |
| 1990* | 3.4 | 337 | 500 | 837 |
| 1991* | 3.7 | 237 | 337 | 574 |
| 1992* | 4.3 | 228 | 245 | 473 |
| 1993* | 3.3 | 535 | 950 | 1,485 |
| 1994* | 3.5 | 1,313 | 2,720 | 4,033 |
| 1995* | 3.4 | 908 | 2,094 | 3,002 |
| 1996* | 3.4 | 394 | 1,425 | 1,819 |
| 1997* | 3.4 | 537 | 1,652 | 2,189 |
| 1998 | 3 | 264 | 828 | 1,092 |
| 1999 | 2.2 | 812 | 2,805 | 3,617 |
| 2000 | 2.4 | 1,318 | 2,383 | 3,701 |
| 2001 | 4.1 | 4,543 | 6,314 | 10,857 |
| 2002 | 2.3 | 6,134 | 7,723 | 13,857 |
| 2003 | 2.4 | 2,505 | 915 | 3,420 |
| 2004 | 2.3 | 2,986 | 3,735 | 6,721 |
| 2005 | 2.9 | 4,720 | 4,169 | 8,889 |
| 2006 | 2 | 5,236 | 3,365 | 8,601 |
| 2007 | 2.2 | 2,862 | 1,555 | 4,417 |
| 2008 | 3.3 | 3,725 | 3,250 | 6,975 |
| 2009 | 2.5 | 4,247 | 3,297 | 7,544 |
| 2010 | 2.8 | 2,841 | 3,111 | 5,952 |
| 2011 | 3.1 | 5,313 | 4,368 | 9,681 |
| 2012 | 3.1 | 4,952 | 3,273 | 8,225 |
| 2013 | 2.3 | 5,237 | 2,957 | 8,194 |
| 2014 | 2.9 | 6,381 | 5,783 | 12,164 |
| 2015 | 3.2 | 7,660 | 6,108 | 13,769 |
| 2016 | 2.01 | 7,007 | 3,598 | 10,605 |
| Average | 2.95 | 2990 | 2881 | 5872 |
| 5-Year Average | 2.70 | 6247 | 4344 | 10591 |

[^6]Table 19. Number and timing of summer Chinook redd counts in reaches of the Okanogan and Similkameen Rivers in 2016.

| Reach | River mile | $\begin{gathered} \text { Sept } \\ 19-25 \end{gathered}$ | $\begin{aligned} & \text { Sept } \\ & 26- \\ & \text { Oct } 2 \end{aligned}$ | $\begin{gathered} \text { Oct 3- } \\ 9 \end{gathered}$ | $\begin{gathered} \text { Oct } \\ \mathbf{1 0 - 1 6} \end{gathered}$ | $\begin{aligned} & \hline \text { Oct } \\ & 17- \\ & 23 \end{aligned}$ | $\begin{gathered} \text { Oct } \\ 24- \\ 30 \end{gathered}$ | $\begin{gathered} \hline \text { Oct } \\ \text { 31- } \\ \text { Nov } 6 \end{gathered}$ | Redd Count | Percent |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Okanogan River |  |  |  |  |  |  |  |  |  |
| 01 | 0.0-16.9 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 2 | 0\% |
| 02 | $\begin{aligned} & \hline 16.9- \\ & 26.1 \\ & \hline \end{aligned}$ | 0 | 0 | 2 | 22 | 20 | 12 | 1 | 57 | 2\% |
| 03 | $\begin{gathered} \hline 26.1- \\ 30.7 \end{gathered}$ | 0 | 0 | 6 | 28 | 16 | 1 | 1 | 52 | 1\% |
| 04 | $\begin{gathered} 30.7- \\ 40.7 \\ \hline \end{gathered}$ | 0 | 0 | 10 | 57 | 34 | 26 | 3 | 130 | 4\% |
| 05 | $\begin{gathered} 40.7- \\ 56.8 \\ \hline \end{gathered}$ | 0 | 0 | 127 | 447 | 240 | 91 | 2 | 907 | 26\% |
| 06 | $\begin{gathered} 56.8- \\ 77.4 \end{gathered}$ | 0 | 9 | 646 | 643 | 498 | 537 | 5 | 2338 | 67\% |
| Total |  | 0 | 9 | 791 | 1197 | 810 | 667 | 12 | 3486 | 100\% |
|  | Similkameen River |  |  |  |  |  |  |  |  |  |
| S1 | 0.0-1.8 | 2 | 15 | 518 | 434 | 447 | 229 | 0 | 1645 | 92\% |
| S2 | 1.8-5.7 | 0 | 0 | 46 | 31 | 51 | 17 | 0 | 145 | 8\% |
| Total |  | 2 | 15 | 564 | 465 | 498 | 246 | 0 | 1790 | 100\% |

## Escapement into Canada

In 2016 there were 1,823 adult summer/fall Chinook counted in the fishways of Zosel Dam (Table 20). While not the highest count on record, 2016 continued a trend showing increasing escapements above Zosel. $13 \%$ of the Chinook observed at Zosel Dam had a clipped adipose fin (i.e., hatchery-origin).

Table 20. Count of run escapement of adult summer/fall Chinook at Zosel Dam using video monitoring in the fishways.

| Chinook Passage at Zosel Dam |  |  |
| :---: | :---: | :---: |
| Year | Video <br> Count | \% <br> Hatchery |
| 2006 | 481 | $1 \%$ |
| 2007 | 455 | $40 \%$ |
| 2008 | 267 | $29 \%$ |
| 2009 | 256 | $17 \%$ |
| 2010 | 359 | $29 \%$ |
| 2011 | 1415 | $36 \%$ |
| 2012 | 826 | $24 \%$ |
| 2013 | 2275 | $14 \%$ |
| $2014{ }^{\text {a }}$ | 1188 | $10 \%$ |
| 2015 | 1206 | $7 \%$ |
| 2016 | 1823 | $13 \%$ |
| Average | 959 | $20 \%$ |

${ }^{a} 2014$ data were adjusted for fallback/re ascension, down camera time, and differentiation of spring Chinook from summer/fall Chinook.

## Carcass Surveys

In 2016, 2,755 carcasses were recovered including 2,171 natural-origin and 584 hatchery-origin ${ }^{7}$. The overall carcass recovery rate was $26 \%$ of the total spawning escapement. Genetic samples (tissue punches) were collected from a portion of the summer/fall Chinook carcasses in 2016. Samples are archived at the USGS Snake River Field Station Genetics Lab in Boise, ID. The majority of carcasses ( $n=2,289$; 83\%) were collected from reaches 06 and S1 (Figure 25, also see Appendix C). The proportion of natural-origin carcasses recovered in 2016 was slightly lower in reach 05 compared to the 10 -year average (Figure 25, panel A). The proportion of hatchery-origin carcasses recovered in 2016 was lower in 05 and 06 and higher in S 1 compared to the 10-year averages (Figure 25, panel B).

[^7]


Figure 25. Distribution of natural-origin (A) and hatchery-origin (B) summer/fall Chinook carcasses recovered in the Okanogan (01-06) and Similkameen (S1-S2) Rivers in 2016 and the 10-year average (2007-2016).

In the Okanogan basin, 59 of the 1,804 sampled female carcasses were estimated to have all their eggs. Therefore pre-spawn mortality (for fish that survived to the spawn period) was estimated to be $3.4 \%$ for natural-origin females and $2.8 \%$ for hatchery-origin females (Table 21). Overall egg retention of all fish sampled (including fish that had expelled a portion of their eggs) was $4.1 \%$.

Table 21. Egg retention and pre-spawn mortality of sampled summer/fall Chinook carcasses in the Okanogan Basin.

| Year | Origin | Total <br> carcasses <br> sampled | Female <br> carcasses <br> sampled | Potential <br> egg <br> depositio <br> n | Natural- <br> origin | 613 | 326 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Eggs <br> retained <br> yatcher <br> y-origin | 297 | 237 | retention <br> rate | bPre-spawn <br> mortality <br> rate |  |  |
|  | Total | 910 | 563 | $2,815,000$ | 6,152 | $0.40 \%$ | $0.00 \%$ |
| 2014 | Natural- <br> origin | 2123 | 1136 | $5,680,000$ | 373,708 | $6.60 \%$ | $1.40 \%$ |
|  | Hatcher <br> y-origin | 329 | 166 | 830,000 | 81,105 | $9.80 \%$ | $1.80 \%$ |
|  | Total | 2452 | 1302 | $6,510,000$ | 454,813 | $7.00 \%$ | $1.50 \%$ |
| 2015 | Natural- <br> origin | 2554 | 981 | $4,905,000$ | 609,869 | $12.40 \%$ | $10.90 \%$ |
|  | Hatcher <br> y-origin | 738 | 340 | $1,700,000$ | 96,354 | $5.70 \%$ | $5.00 \%$ |
|  | Total | 3292 | 1321 | $6,605,000$ | 706,223 | $10.70 \%$ | $9.40 \%$ |
| 2016 | Natural- <br> origin | 2171 | 1370 | $6,850,000$ | 300,046 | $4.38 \%$ | $3.43 \%$ |
|  | Hatcher <br> y-origin | 584 | 434 | $2,170,000$ | 66,254 | $3.05 \%$ | $2.76 \%$ |
|  | Total | 2755 | 1804 | $9,020,000$ | 366300 | $4.06 \%$ | $3.27 \%$ |

${ }^{\text {a Assuming fecundity of 5,000 eggs per female, egg retention rate is calculated as: (\# eggs estimated remaining in sampled }}$ female carcasses) / (\# female carcasses sampled * 5,000 eggs each)
${ }^{\text {b }}$ A pre-spawn mortality is determined when a female retains the assumed 5,000 eggs on the spawning grounds.

## PHOS AND PNI

There was a decrease in the proportion of hatchery-origin spawners ( pHOS ) across most reaches in the Okanogan and Similkameen rivers in 2016 compared to the 10-year average with the exception of reach 02 and 04 (Figure 26), which were based on the recovery of only 12 and 11 carcasses, respectively. No carcasses were recovered in reach 01 and only 2 redds were observed in this reach. Hatchery-origin spawners comprised $18 \%$ of the spawn escapement estimate in the U.S. portion of the Okanogan, which was the second lowest pHOS observed since 1992 (Table 22). After corrections for hatchery fish effectiveness assumptions ( 0.80 relative reproductive success rate for hatchery-origin spawners) the effective pHOS for 2016 was 0.15 , which was considerably less than the five-year average ( 0.23 ) (Table 23). The five-year average is currently meeting the biological objective for $\mathrm{pHOS}(<0.3)$, and continues to trend toward the program objective of reduced pHOS (Figure 27).

The proportion of natural-origin broodstock (pNOB) in 2016 was $100 \%$ and the pNOB for Okanogan origin fish was $90 \%$ (Table 23). The resulting PNI for 2016 was 0.87 , with a 5 -year average PNI of 0.82 , both meeting the Biological Objective ( $>0.67$ )(Figure 28).


Figure 26. Okanogan (01-06) and Similkameen River (S1-S2) summer/fall Chinook unadjusted pHOS by reach for 2016 and 10-year average (2007-2016). Reaches with fewer than 10 carcasses recovered were not shown. Error bars represent standard error of the mean.

Table 22. Natural- (NOS) and hatchery- (HOS) origin spawner abundance and composition for the Okanogan River Basin, brood years 1989-2016.

| Brood Year | Spawners |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
|  | NOS | HOS | pHOS | Effective pHOS^ |
| 1989 | 1,719 | 0 | 0 | 0 |
| 1990 | 837 | 0 | 0 | 0 |
| 1991 | 574 | 0 | 0 | 0 |
| 1992 | 473 | 0 | 0 | 0 |
| 1993 | 915 | 570 | 0.38 | 0.33 |
| 1994 | 1,323 | 2,710 | 0.67 | 0.62 |
| 1995 | 979 | 2,023 | 0.67 | 0.62 |
| 1996 | 568 | 1,251 | 0.69 | 0.64 |
| 1997 | 862 | 1,327 | 0.61 | 0.55 |
| 1998 | 600 | 492 | 0.45 | 0.40 |
| 1999 | 1,274 | 2,343 | 0.65 | 0.60 |
| 2000 | 1,174 | 2,527 | 0.68 | 0.63 |
| 2001 | 4,306 | 6,551 | 0.60 | 0.55 |
| 2002 | 4,346 | 9,511 | 0.69 | 0.64 |
| 2003 | 1,933 | 1,487 | 0.43 | 0.38 |
| 2004 | 5,309 | 1,412 | 0.21 | 0.18 |
| 2005 | 6,441 | 2,448 | 0.28 | 0.23 |
| 2006* | 5,507 | 3,094 | 0.21 | 0.18 |
| 2007* | 2,983 | 1,434 | 0.37 | 0.32 |
| 2008* | 2,998 | 3,977 | 0.60 | 0.54 |
| 2009* | 4,204 | 3,340 | 0.46 | 0.40 |
| 2010* | 3,189 | 2,763 | 0.47 | 0.41 |
| 2011* | 4,642 | 5,039 | 0.52 | 0.47 |
| 2012* | 4,840 | 3,385 | 0.45 | 0.40 |
| 2013*, ${ }^{\text {a }}$ | 5,520 | 2,674 | 0.31 | 0.27 |
| 2014* | 10,402 | 1,762 | 0.14 | 0.12 |
| 2015* | 10,350 | 3,398 | 0.25 | 0.21 |
| 2016* | 8,661 | 1,944 | 0.18 | 0.15 |
| Average | 3,466 | 2,405 | 0.39 | 0.35 |
| 5-year <br> Average | 7,981 | 2,607 | 0.27 | 0.23 |

a 2013 data have been updated to reflect age and origin data acquired from scale reading since the publication of the 2013 annual report.

* Indicates that pHOS has been reach-weighted by \# of redds per reach.
${ }^{\wedge}$ Effective pHOS assumes 0.80 HOS effectiveness


Figure 27. Annual and 5-year average proportion of hatchery-origin spawners ( pHOS ) in the Okanogan and Similkameen River (combined) from 1998-2016. pHOS values represent the effective pHOS.

Table 23. Okanogan River summer Chinook spawn escapement and broodstock composition, and calculated pHOS and PNI for Brood Years 1989-2016.

| Brood Year | Spawners |  |  | Broodstock |  |  |  |  | PNI | Okan. PNI |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | NOS | HOS | Effective pHOS | NOB | Okan. <br> NOB | $\mathrm{HOB}$ | pNOB | Okan. <br> pNOB |  |  |
| 1989 | 1,719 | 0 | 0.00 | 1,297 |  | 312 | 0.81 |  | 1.00 |  |
| 1990 | 837 | 0 | 0.00 | 828 |  | 206 | 0.80 |  | 1.00 |  |
| 1991 | 574 | 0 | 0.00 | 924 |  | 314 | 0.75 |  | 1.00 |  |
| 1992 | 473 | 0 | 0.00 | 297 |  | 406 | 0.42 |  | 1.00 |  |
| 1993 | 915 | 570 | 0.33 | 681 |  | 388 | 0.64 |  | 0.66 |  |
| 1994 | 1,323 | 2,710 | 0.62 | 341 |  | 244 | 0.58 |  | 0.48 |  |
| 1995 | 979 | 2,023 | 0.62 | 173 |  | 240 | 0.42 |  | 0.40 |  |
| 1996 | 568 | 1,251 | 0.64 | 287 |  | 155 | 0.65 |  | 0.50 |  |
| 1997 | 862 | 1,327 | $0.55$ | 197 |  | 265 | $0.43$ |  | 0.44 |  |
| 1998 | 600 | 492 | 0.40 | 153 | 77 | 211 | 0.42 | 0.21 | 0.51 | 0.35 |
| 1999 | 1,274 | 2,343 | 0.60 | 224 | 112 | 289 | 0.44 | 0.22 | 0.42 | 0.27 |
| 2000 | 1,174 | 2,527 | 0.63 | 164 | 82 | 337 | 0.33 | 0.16 | 0.34 | 0.21 |
| 2001 | 4,306 | 6,551 | 0.55 | 12 | 46 | 345 | 0.03 | 0.13 | 0.06 | 0.19 |
| 2002 | 4,346 | 9,511 | 0.64 | 247 | 124 | 241 | 0.51 | 0.25 | 0.44 | 0.29 |
| 2003 | 1,933 | 1,487 | 0.38 | 381 | 191 | 101 | 0.79 | 0.40 | 0.67 | 0.51 |
| 2004 | 5,309 | 1,412 | 0.18 | 506 | 253 | 16 | 0.97 | 0.48 | 0.85 | 0.73 |
| 2005 | 6,441 | 2,448 | 0.23 | 391 | 196 | 9 | 0.98 | 0.49 | 0.81 | 0.68 |
| 2006 | 5,507 | 3,094 | 0.18 | 500 | 250 | 10 | 0.98 | 0.49 | 0.85 | 0.73 |
| 2007 | 2,983 | 1,434 | 0.32 | 456 | 228 | 17 | 0.96 | 0.48 | 0.75 | 0.60 |
| 2008 | 2,998 | 3,977 | 0.54 | 359 | 202 | 86 | 0.81 | 0.45 | 0.60 | 0.46 |
| 2009 | 4,204 | 3,340 | 0.40 | 503 | 254 | 4 | 0.99 | 0.50 | 0.71 | 0.55 |
| 2010 | 3,189 | 2,763 | 0.41 | 484 | 242 | 8 | 0.98 | 0.49 | 0.70 | 0.54 |
| 2011 | 4,642 | 5,039 | 0.47 | 467 | 332 | 26 | 0.95 | 0.67 | 0.67 | 0.59 |
| 2012 | 4,840 | 3,385 | 0.40 | 107 | 96 | 0 | 1.00 | 0.90 | 0.72 | 0.69 |
| 2013 | 5,627 | 2,567 | 0.27 | 353 | 318 | 0 | 1.00 | 0.90 | 0.79 | 0.77 |
| 2014 | 10,402 | 1,762 | 0.12 | 499 | 449 | 5 | 0.99 | 0.89 | 0.89 | 0.88 |
| 2015 | 10,350 | 3,398 | 0.21 | 421 | 379 | 9 | 0.98 | 0.88 | 0.82 | 0.81 |
| 2016 | 8,661 | 1,944 | 0.15 | 584 | 526 | 0 | 1.00 | 0.90 | 0.87 | 0.86 |
| Average | 3,466 | 2,406 | 0.35 | 423 | 229 | 152 | 0.74 | 0.52 | 0.68 | 0.56 |
| 5-Year <br> Average | 7,976 | 2,611 | 0.23 | 393 | 354 | 3 | 0.99 | 0.89 | 0.82 | 0.80 |

pHOS values are effective from 1989-2006 and Effective, Reach-weighted pHOS from 2006-2016


Figure 28. Annual and 5-year average proportionate natural influence (PNI) in the Okanogan and Similkameen Rivers (combined) from 1998 to 2016.

## Age Structure

Attempts were made to age all carcasses recovered on the spawning grounds, either by reading scales on natural-origin fish or by extracting and reading coded wire tags for hatchery-origin fish. Historically, most natural-origin summer Chinook migrate as subyearlings, while the majority of hatchery-origin releases in the Okanogan river basin have been released as yearlings. To account for this difference, the number of winters a fish spent in the marine environment - salt age - is the format of reported data.

In 2016, both male and female natural- and hatchery-origin fish were more apt to return as 3-salt fish than the 10-year average - the pool of spawners in brood year 2016 were older than average (Figure 29).



## Hatchery-origin Female Age Structure




Figure 29. The salt ages of carcasses collected on the spawning grounds of the Okanogan and Similkameen Rivers in 2016 and 10-year averages (2007-2016).

## Hatchery-ORIGIN Stray Rates

Strays within the Okanogan.-The majority ( $95 \%$ ) of hatchery-origin spawners recovered on the spawning grounds in 2016 were from Similkameen (91\%) and Okanogan (4\%) River releases (Table 24). This was very similar to the average ( $95 \%$ ) of recent years (2006-2016). Strays from outside the Okanogan but within the Upper Columbia summer/fall Chinook ESU consisted of fish from Dryden Pond, Carlton Pond, Entiat NFH, Chelan Falls Hatchery, and Wells Hatchery (4\%). Strays from outside the ESU were from releases into the Yakima River (1\%). Stray hatchery fish from outside the Okanogan comprised less than $1 \%$ of total (hatchery plus natural-origin) Okanogan spawner
composition (i.e., stray pHOS) (Table 25). This was less than the recent (2006-2016) average of $1.9 \%$ and well under the biological target of $<5 \%$.

Strays outside the Okanogan.- The most recent brood year that could be fully assessed (through age 5) for stray rate of Okanogan fish to spawning areas outside the Okanogan was 2011. The 2011 brood year had a stray rate of $1.1 \%$, which was similar to the long term and recent five year averages (Table 26). RMIS queries revealed an estimate of 14 Okanogan hatchery-origin CWT codes from spawning ground recoveries in nontarget spawning areas in 2016 (Table 26). Okanogan basin hatchery program strays comprise $\leq 1 \%$ to other basin population's spawner composition (in 2016 as well as long term average). (Table 27).

Table 24. Estimated number (and percent of annual total) of hatchery-origin spawners from different release basins recovered on the Okanogan/Similkameen spawning grounds, based on CWT recoveries and expansions, for return years 20062016.

| Return Year | Release Site |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Summer Chinook Run |  |  |  |  |  |  | Spring and Fall Chinook Run |  |  |
|  | Homing Fish |  | Straying Fish |  |  |  |  |  |  |  |
|  | Okanogan River Basin |  | Within ESU Stray |  |  |  |  | Out of ESU Stray |  |  |
|  | Okanogan River ${ }^{\text {a }}$ | Similkameen River ${ }^{\text {b }}$ | Methow River ${ }^{\text {c }}$ | Wenatchee River ${ }^{\text {d }}$ | Entiat <br> River ${ }^{\text {e }}$ | Chelan River ${ }^{f}$ | Mainstem Columbia Rivers | Mainstem Columbia Riverh | Snake River ${ }^{\text {i }}$ | Other ${ }^{\text {i }}$ |
| 2006 | 0 (0\%) | 709 (87\%) | 12 (2\%) | 12 (2\%) | 0 (0\%) | 0 (0\%) | 81 (10\%) | 0 (0\%) | 0 (0\%) | 0 (0\%) |
| 2007 | 0 (0\%) | 1121 (95\%) | 17 (1\%) | 5 (0\%) | 0 (0\%) | 0 (0\%) | 42 (4\%) | 0 (0\%) | 0 (0\%) | 0 (0\%) |
| 2008 | 0 (0\%) | 3224 (95\%) | 11 (0\%) | 24 (1\%) | 0 (0\%) | 4 (0\%) | 133 (4\%) | 3 (0\%) | 0 (0\%) | 0 (0\%) |
| 2009 | 0 (0\%) | 2733 (95\%) | 14 (0\%) | 14 (0\%) | 0 (0\%) | 9 (0\%) | 99 (3\%) | 0 (0\%) | 5 (0\%) | 4 (0\%) |
| 2010 | 4 (0\%) | 2165 (89\%) | 44 (2\%) | 35 (1\%) | 0 (0\%) | 110 (5\%) | 75 (3\%) | 0 (0\%) | 4 (0\%) | 0 (0\%) |
| 2011 | 219 (5\%) | 4196 (93\%) | 44 (1\%) | 5 (0\%) | 0 (0\%) | 34 (1\%) | 22 (0\%) | 0 (0\%) | 6 (0\%) | 0 (0\%) |
| 2012 | 379 (13\%) | 2397 (83\%) | 29 (1\%) | 23 (1\%) | 0 (0\%) | 17 (1\%) | 52 (2\%) | 0 (0\%) | 0 (0\%) | 0 (0\%) |
| 2013 | 254 (14\%) | 1437 (81\%) | 10 (1\%) | 54 (3\%) | 0 (0\%) | 0 (0\%) | 10 (1\%) | 0 (0\%) | 0 (0\%) | 0 (0\%) |
| 2014 | 55 (5\%) | 1023 (90\%) | 16 (1\%) | 0 (0\%) | 6 (1\%) | 12 (1\%) | 29 (3\%) | 0 (0\%) | 0 (0\%) | 0 (0\%) |
| 2015 | 38 (1\%) | 2562 (91\%) | 70 (3\%) | 17 (1\%) | 19 (1\%) | 33 (1\%) | 33 (1\%) | 4 (0\%) | 4 (0\%) | 21 (1\%) |
| 2016 | 81(4\%) | 1963 (91\%) | 42 (2\%) | 7 (0\%) | 3 (0\%) | 31 (1\%) | 14 (1\%) | 0 (0\%) | 0 (0\%) | 17(1\%) |
| Avg. | 94 (4\%) | 2139 (90\%) | 28 (1\%) | 18 (1\%) | 3 (0\%) | 23 (1\%) | 54 (3\%) | 1 (0\%) | $\begin{gathered} 2 \\ (0 \%) \\ \hline \end{gathered}$ | 4 (0\%) |

${ }^{\text {a }}$ Includes releases from Omak and Bonaparte Pond. Three spring Chinook recovered in 2008 from an Omak Creek release were excluded from analysis.
${ }^{\mathrm{b}}$ Includes releases from Similkameen Pond
${ }^{\text {c }}$ Includes releases from Carlton Acclimation Pond
${ }^{d}$ Includes releases from Dryden Pond and Eastbank Hatchery
${ }^{e}$ Includes releases from Entiat NFH
${ }^{f}$ Includes releases from Chelan PUD Hatchery, Chelan River NFH, and Chelan Hatchery
g Includes releases of summer Chinook from Wells Hatchery, Turtle Rock Hatchery, and Grant County PUD Hatchery
${ }^{\mathrm{h}}$ Includes releases of fall Chinook from Hanford Reach
${ }^{i}$ Includes Releases from NPT Hatchery
${ }^{j}$ Includes releases from Marion Yakama Tribal, Cle Elum Hatchery, and Prosser Hatchery

Table 25. Estimated percent of spawner composition of hatchery-origin spawners from different release basins recovered on the Okanogan/Similkameen spawning grounds, based on CWT recoveries and expansions, for return years 2006-2016.

| Return Year | Release Site |  |  |  |  |  |  |  |  |  | HOS Stray Contribution to Total Spawning Escapement | pHOS |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Summer Chinook Run |  |  |  |  |  |  | Other Chinook RunOut of ESU Stray |  |  |  |  |
|  | Okanogan River Basin |  | Within ESU Stray |  |  |  |  |  |  |  |  |  |
|  | Okanogan River ${ }^{\text {a }}$ | Similkameen River ${ }^{\text {b }}$ | Methow Riverc | Wenatchee River ${ }^{\text {d }}$ | Entiat <br> Rivere | Chelan <br> Riverf | Mainstem Columbia Riverg | Mainstem Columbia River ${ }^{\text {h }}$ | Snake River ${ }^{\text {i }}$ | Other ${ }^{\text {j }}$ |  |  |
| 2006 | 0.0\% | 15.6\% | 0.3\% | 0.3\% | 0.0\% | 0.0\% | 1.8\% | 0.0\% | 0.0\% | 0.0\% | 2.3\% | 0.18 |
| 2007 | 0.0\% | 30.0\% | 0.5\% | 0.1\% | 0.0\% | 0.0\% | 1.1\% | 0.0\% | 0.0\% | 0.0\% | 1.7\% | 0.32 |
| 2008 | 0.0\% | 51.5\% | 0.2\% | 0.4\% | 0.0\% | 0.1\% | 2.1\% | 0.1\% | 0.0\% | 0.0\% | 2.8\% | 0.54 |
| 2009 | 0.0\% | 38.4\% | 0.2\% | 0.2\% | 0.0\% | 0.1\% | 1.4\% | 0.0\% | 0.1\% | 0.1\% | 2.0\% | 0.40 |
| 2010 | 0.6\% | 40.71\% | 0.84\% | 0.67\% | 0.0\% | 2.1\% | 1.43\% | 0.0\% | 0.08\% | 0.0\% | 4.54\% | 0.41 |
| 2011 | 2.52\% | 48.26\% | 0.51\% | 0.06\% | 0.0\% | 0.39\% | 0.25\% | 0.0\% | 0.07\% | 0.0\% | 1.27\% | 0.47 |
| 2012 | 5.38\% | 34.05\% | 0.41\% | 0.33\% | 0.0\% | 0.24\% | 0.74\% | 0.0\% | 0.0\% | 0.0\% | 1.72\% | 0.40 |
| 2013 | 3.4\% | 19.5\% | 0.1\% | 0.7\% | 0.0\% | 0.0\% | 0.1\% | 0.0\% | 0.0\% | 0.0\% | 1.29\% | 0.24 |
| 2014 | 0.7\% | 12.99\% | 0.21\% | 0.0\% | 0.0\% | 0.15\% | 0.37\% | 0.0\% | 0.0\% | 0.0\% | 0.72\% | 0.11 |
| 2015 | 0.34\% | 22.69\% | 0.58\% | 0.15\% | 0.16\% | 0.27\% | 0.27\% | 0.03\% | 0.03\% | 0.17\% | 1.67\% | 0.21 |
| 2016 | 0.32\% | 17.37\% | 0.26\% | 0.09\% | 0.0\% | 0.12\% | 0.04\% | 0.0\% | 0.0\% | 0.09\% | 0.59\% | 0.19 |
| Avg. | 1.2\% | 30.1\% | 0.4\% | 0.3\% | 0.0\% | 0.3\% | 0.9\% | 0.0\% | 0.0\% | 0.0\% | 1.9\% | 31.5\% |

a Includes releases from Bonaparte Pond. Three spring Chinook recovered in 2008 from an Omak Creek release were excluded from analysis.
b Includes releases from Similkameen Pond
${ }^{\mathrm{c}}$ Includes releases from Carlton Acclimation Pond
${ }^{d}$ Includes releases from Dryden Pond and Eastbank Hatchery
${ }^{e}$ Includes releases from Entiat NFH
${ }^{f}$ Includes releases from Chelan PUD Hatchery, Chelan River NFH, and Chelan Hatchery
g Includes releases of summer Chinook from Wells Hatchery, Turtle Rock Hatchery, and Grant County PUD Hatchery
${ }^{h}$ Includes releases of fall Chinook from Hanford Reach
${ }^{\text {i }}$ Includes Releases from NPT Hatchery
j Includes releases from Marion Yakama Tribal, Cle Elum Hatchery, and Prosser Hatchery

Table 26. Number and percent (\%) of hatchery-origin Okanogan summer/fall Chinook that were recovered at target spawning areas or were captured at en route hatcheries (Wells and Chief Joseph Hatchery), and number and percent that strayed to non-target spawning areas and non-target hatcheries, brood years 1989-2010. As fish continue to return through time and the RMIS database is continually updated, reported data from recent brood years may change.

| Brood Year | Homing |  |  |  | Straying |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Target Stream |  | En Route Hatchery |  | Non-target Streams |  | Non-target Hatchery |  |
|  | Number | \% | Number | \% | Number | \% | Number | \% |
| 1989 | 3,132 | 69.7\% | 1,328 | 29.6\% | 2 | 0.0\% | 31 | 0.7\% |
| 1990 | 729 | 71.4\% | 291 | 28.5\% | 0 | 0.0\% | 1 | 0.1\% |
| 1991 | 1,125 | 71.3\% | 453 | 28.7\% | 0 | 0.0\% | 0 | 0.0\% |
| 1992 | 1,264 | 68.5\% | 572 | 31.0\% | 8 | 0.4\% | 1 | 0.1\% |
| 1993 | 54 | 62.1\% | 32 | 36.8\% | 0 | 0.0\% | 1 | 1.1\% |
| 1994 | 924 | 80.8\% | 203 | 17.7\% | 16 | 1.4\% | 1 | 0.1\% |
| 1995 | 1,883 | 85.4\% | 271 | 12.3\% | 52 | 2.4\% | 0 | 0.0\% |
| 1996 | 27 | 100.0\% | 0 | 0.0\% | 0 | 0.0\% | 0 | 0.0\% |
| 1997 | 11,659 | 97.1\% | 309 | 2.6\% | 35 | 0.3\% | 2 | 0.0\% |
| 1998 | 2,784 | 95.4\% | 102 | 3.5\% | 31 | 1.1\% | 2 | 0.1\% |
| 1999 | 828 | 96.7\% | 18 | 2.1\% | 10 | 1.2\% | 0 | 0.0\% |
| 2000 | 2,091 | 93.8\% | 29 | 1.3\% | 94 | 4.2\% | 15 | 0.7\% |
| 2001 | 105 | 98.1\% | 2 | 1.9\% | 0 | 0.0\% | 0 | 0.0\% |
| 2002 | 702 | 96.2\% | 17 | 2.3\% | 11 | 1.5\% | 0 | 0.0\% |
| 2003 | 1,580 | 96.2\% | 47 | 2.9\% | 16 | 1.0\% | 0 | 0.0\% |
| 2004 | 4,947 | 94.4\% | 206 | 3.9\% | 85 | 1.6\% | 2 | 0.0\% |
| 2005 | 606 | 93.2\% | 22 | 3.4\% | 22 | 3.4\% | 0 | 0.0\% |
| 2006 | 5,210 | 97.6\% | 60 | 1.1\% | 68 | 1.3\% | 0 | 0.0\% |
| 2007 | 1,330 | 97.9\% | 19 | 1.4\% | 10 | 0.7\% | 0 | 0.0\% |
| 2008 | 3,673 | 96.5\% | 111 | 2.9\% | 19 | 0.5\% | 4 | 0.1\% |
| 2009 | 1,149 | 80.8\% | 256 | 18.0\% | 14 | 1.0\% | 2 | 0.1\% |
| 2010 | 1,058 | 61.4\% | 646 | 37.5\% | 9 | 0.5\% | 10 | 0.6\% |
| 2011 | 4,449 | 79.9\% | 873 | 18.9\% | 10 | 0.6\% | 25 | 0.5\% |
| Total | 51,309 | 86.3\% | 5,867 | 12.5\% | 512 | 1.0\% | 97 | 0.2\% |

Table 27. Number and percent (\%) of spawning escapements that consisted of hatcheryorigin Okanogan summer/fall Chinook within other non-target basins, return years 19942016.

| Return Year | Wenatchee |  | Methow |  | Chelan |  | Entiat |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Number | \% | Number | \% | Number | \% | Number | \% |
| 1994 | 0 | 0.0\% | 0 | 0.0\% | - | - | - | - |
| 1995 | 0 | 0.0\% | 0 | 0.0\% | - | - | - | - |
| 1996 | 0 | 0.0\% | 0 | 0.0\% | - | - | - | - |
| 1997 | 0 | 0.0\% | 0 | 0.0\% | - | - | - | - |
| 1998 | 0 | 0.0\% | 0 | 0.0\% | 0 | 0.0\% | 0 | 0.0\% |
| 1999 | 0 | 0.0\% | 0 | 0.0\% | 0 | 0.0\% | 0 | 0.0\% |
| 2000 | 0 | 0.0\% | 6 | 0.5\% | 30 | 6.4\% | 0 | 0.0\% |
| 2001 | 12 | 0.1\% | 0 | 0.0\% | 10 | 1.0\% | 0 | 0.0\% |
| 2002 | 0 | 0.0\% | 3 | 0.1\% | 4 | 0.7\% | 5 | 1.0\% |
| 2003 | 0 | 0.0\% | 8 | 0.2\% | 22 | 5.3\% | 14 | 2.0\% |
| 2004 | 0 | 0.0\% | 0 | 0.0\% | 5 | 1.2\% | 0 | 0.0\% |
| 2005 | 5 | 0.1\% | 27 | 1.1\% | 36 | 6.9\% | 7 | 1.9\% |
| 2006 | 0 | 0.0\% | 5 | 0.2\% | 4 | 1.0\% | 2 | 0.4\% |
| 2007 | 0 | 0.0\% | 3 | 0.2\% | 4 | 2.1\% | 0 | 0.0\% |
| 2008 | 0 | 0.0\% | 9 | 0.5\% | 46 | 9.3\% | 4 | 1.3\% |
| 2009 | 15 | 0.2\% | 3 | 0.2\% | 11 | 1.8\% | 18 | 7.1\% |
| 2010 | 6 | 0.1\% | 0 | 0.0\% | 33 | 3.0\% | 0 | 0.0\% |
| 2011 | 0 | 0.0\% | 0 | 0.0\% | 45 | 3.5\% | 0 | 0.0\% |
| 2012 | 7 | 0.1\% | 5 | 0.2\% | 18 | 1.4\% | 0 | 0.0\% |
| 2013 | 0 | 0.0\% | 0 | 0.0\% | 0 | 0.0\% | 0 | 0.0\% |
| 2014 | 0 | 0.0\% | 4 | 0.2\% | 11 | 1.0\% | 0 | 0.0\% |
| 2015 | 4 | 0.1\% | 5 | 0.1\% | 4 | 0.3\% | 0 | 0.0\% |
| 2016 | 0 | 0.0\% | 4 | 0.2\% | 4 | 0.5\% | 0 | 0.0\% |
| Total | 49 | 0.0\% | 82 | 0.2\% | 283 | 2.4\% | 50 | 0.7\% |
| 5-year <br> Total | 11 | 0.0\% | 18 | 0.1\% | 37 | 0.6\% | - | 0.0\% |

## Homing Fidelity

Based on 370 coded-wire tags recovered during spawning grounds surveys in fall of 2016, an estimated 67 and 1858 (CWT-expanded) spawners originated from Omak Pond and Similkameen Pond acclimation sites, respectively. Of those spawners, the majority of the fish originating from the Omak Pond acclimation site spawned in the Okanogan River (Table 28). Those fish tended to spawn in habitat upstream of the Omak Pond site, with the
majority CWT's recovered in the reach 05. Only a single CWT was recovered below reach 04, from a fish originally acclimated at Omak Pond, which was recovered in reach 02 (Figure 30). Only an estimated 18 fish that were acclimated at Omak Pond chose to spawn in the Similkameen River. These returns would consist only of $\leq 3$ year olds for fish released from Omak Pond. Future return years will include additional age classes and should provide greater insight into homing fidelity for fish acclimated at Omak Pond. Fish acclimated at the Similkameen Pond site spawned primarily in reach S1, with substantial numbers of fish also spawning in habitat within reaches 05 and 06 (Figure 30). Overall, Similkameen pond-origin fish tended to home to the Similkameen River (Table 28).

Table 28. Spawning distribution by river, for fish acclimated at Omak Pond and Similkameen Pond acclimation sites.

|  | Acclimation site (origin) |  |
| :---: | :---: | :---: |
| Spawning location | Omak Pond | Similkameen Pond |
| Okanogan River | $73 \%$ | $45 \%$ |
| Similkameen River | $27 \%$ | $55 \%$ |



Figure 30. Spatial distribution of CJH integrated program Chinook spawners originally reared at the Similkameen Pond and Omak Pond acclimation sites.

## Smolt-to-Smolt Survival and Travel Time

2016 was the second year of PIT tagged releases from the CJH. Apparent survival of yearlings to RRJ was generally between 75-80\% with the exception of integrated yearlings from Omak Pond that was $57 \%$ (Tables 28-29). Apparent survival of yearlings to MCN was generally $50-60 \%$ regardless of species or release location (Tables 28-29). Apparent survival to RRJ for sub-yearling summer Chinook was 35-44\% for fish released from CJH facilities, $51 \%$ for Wells Fish Hatchery releases and $25 \%$ for wild subyearlings captured in a beach seine near the mouth of the Okanogan (Table 30).

Statistical tests were not conducted to evaluate if the CJH releases were significantly different than nearby hatcheries or previous years. The guidance from the Annual Program Review was to wait until a multi-year assessment could be conducted with at least 4 or 5 years of data to more accurately evaluate patterns between years and programs.

Table 29. Apparent survival estimates for PIT tagged Spring Chinook Salmon smolts released from Chief Joseph hatchery (CJH), Winthrop National Fish Hatchery (WNFH) and Leavenworth National Fish Hatchery (LNFH) in 2016.

|  |  |  |  |  | Survival Standard | Capture Prob. | Capture Prob |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Spring Chinook Release Group | \# PIT tags |  |  |  |  |  |  |
|  | Released | Recap. | Reach | Survival | Error (SE) |  | (SE) |
| Yearlings released | 4967 | 1040 | Release to RRJ | 0.74 | 0.03 | 0.28 | 0.01 |
| at CJH |  | 703 | Release to MCN | 0.48 | 0.03 | 0.29 | 0.02 |
| Yearlings released | 4959 | 1054 | Release to RRJ | 0.81 | 0.03 | 0.26 | 0.01 |
| at Riverside (10j) |  | 673 | Release to MCN | 0.63 | 0.04 | 0.22 | 0.02 |
| Yearlings released | 7361 | 3896 | Release to RRJ | 0.75 | 0.02 | 0.30 | 0.01 |
| at WNFH |  | 2170 | Release to MCN | 0.58 | 0.02 | 0.22 | 0.01 |
| Yearlings released | 19 |  |  |  |  |  |  |
| at LNFH | 19679 | 2596 | Release to MCN | 0.49 | 0.02 | 0.27 | 0.01 |

Table 30. Apparent survival estimates for PIT tagged Summer Chinook Salmon released in 2016 from Chief Joseph Hatchery (CJH) and other nearby hatcheries.

|  |  |  |  |  | Survival <br> Standard <br> Error (SE) | Capture Prob. | Capture Prob. (SE) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Summer Chinook Release Group | \# PIT tags |  |  |  |  |  |  |
|  | Released | Recap. | Reach | Survival |  |  |  |
| Yearlings released at CJH | 4951 | 1069 | Release to RRJ | 0.78 | 0.04 | 0.28 | 0.01 |
|  |  | 521 | Release to MCN | 0.53 | 0.04 | 0.20 | 0.02 |
| Yearlings released at Omak Pond | 4193 | 459 | Release to RRJ | 0.57 | 0.04 | 0.19 | 0.02 |
|  |  | 295 | Release to MCN | 0.44 | 0.05 | 0.16 | 0.02 |
| Yearlings released at Carlton Pond | 4992 | 977 | Release to RRJ | 0.81 | 0.04 | 0.24 | 0.01 |
|  |  | 505 | Release to MCN | 0.63 | 0.06 | 0.16 | 0.02 |
| Yearlings released at Dryden Pond | 20786 |  |  |  |  |  |  |
|  |  | 2381 | Release to MCN | 0.79 | 0.03 | 0.15 | 0.01 |
| Subyearlings released at CJH | 4983 | 425 | Release to RRJ | 0.44 | 0.08 | 0.19 | 0.04 |
|  |  | 70 | Release to MCN | 0.14 | 0.05 | 0.10 | 0.04 |
| Subyearlings released at Omak | 4979 | 465 | Release to RRJ | 0.35 | 0.05 | 0.27 | 0.04 |
|  |  | 64 | Release to MCN | 0.14 | 0.06 | 0.09 | 0.04 |
| Wells Fish Hatchery Subyearlings | 5983 | 686 | Release to RRJ | 0.51 | 0.05 | 0.22 | 0.02 |
|  |  | 158 | Release to MCN | 0.25 | 0.05 | 0.11 | 0.02 |

Table 31. PIT tag survival estimates for juvenile wild Summer Chinook Salmon captured in a beach seine in Wells Pool, primarily near the mouth of the Okanogan River. 2015a and 2016a were only the fish captured in the beach seine and 2015b and 2016b were the beach seined fish pooled with those tagged at the rotary screw trap in the Okanogan River as well as in a side-channel at Conservancy Island.

| Wild Summer Chinook Release Group |  |  |  |  | Survival <br> Standard <br> Error (SE) | Capture Prob. | Capture <br> Prob. (SE) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | \# PIT tags |  | Reach |  |  |  |  |
|  | Released | Recap. | Release to: | Survival |  |  |  |
| 2011 | 13,221 | 1,200 | RRJ | 0.45 | 0.02 | 0.20 | 0.01 |
|  |  | 920 | MCN | 0.30 | 0.02 | 0.23 | 0.02 |
| 2012 | 15,311 | 912 | RRJ | 0.54 | 0.04 | 0.11 | 0.01 |
|  |  | 795 | MCN | 0.40 | 0.03 | 0.13 | 0.01 |
| 2013 | 17,760 | 1,988 | RRJ | 0.44 | 0.02 | 0.26 | 0.01 |
|  |  | 747 | MCN | 0.39 | 0.04 | 0.11 | 0.01 |
| 2014 | 8,226 | 845 | RRJ | 0.35 | 0.03 | 0.29 | 0.02 |
|  |  | 240 | MCN | 0.19 | 0.04 | 0.16 | 0.03 |
| 2015a | 5,823 | 519 | RRJ | 0.26 | 0.06 | 0.342 | 0.077 |
|  |  | 13 | MCN | NE | NE | NE | NE |
| 2015b | 7,787 | 569 | RRJ | 0.25 | 0.05 | 0.288 | 0.0628 |
|  |  | 19 | MCN | NE | NE | NE | NE |
| 2016a | 13,651 | 1,378 | RRJ | 0.25 | 0.03 | 0.40 | 0.04 |
|  |  | 80 | MCN | NE | NE | NE | NE |
| 2016b | 14,674 | 1,411 | RRJ | 0.24 | 0.03 | 0.40 | 0.04 |
|  |  | 81 | MCN | NE | NE | NE | NE |

NE = No Estimate due to small sample size and low recapture probability
The travel time of fish released from CJH facilities to RRJ in 2016 varied from 18 days ( $6.4 \mathbf{k m} /$ day) for subyearlings released from the hatchery to 12 days ( $13.7 \mathrm{~km} /$ day) for the 10 j Spring Chinook released from Riverside Pond (Table 32). As in 2015, sub-yearling summer Chinook from the Omak Pond had one of the fastest migration speeds ( $11.2 \mathrm{~km} /$ day) of all release groups, outpacing all of the yearling summer Chinook groups (Table 32Table 32. ). Direct comparisons of migration speed may not be applicable because not all fish are released at the same time and location and therefore do not experience the same water conditions (e.g., temperature, velocity). Most notably, the sub-yearlings were released approximately 1-1.5 months later than the
yearlings and the wild juveniles were not tagged until June and July. Consequently, arrival timing at RRJ was substantially earlier for yearlings than sub-yearlings (


Figure 31). Summer Chinook arrived at Rocky Reach over a 5-6 week period; late April to early June for yearlings and early June to mid- July for sub-yearlings. The majority of Spring Chinook from CJH and Riverside Pond arrived at RRJ during the first two weeks in May (Figure 32). The migration speed increased substantially in reaches downstream of Rocky Reach Dam for all release groups, although downstream reach estimates could not be estimated for all groups due to small sample sizes ( $\mathrm{n}<10$ )(Table 31.).

Table 32. Travel time and migration speed for various Chinook release groups in 2016

| Release Group | First Day of Release 2016 | Last <br> Day of Release 2016 | Forced or Volitional | Release to RRJ |  | RRJ to MCN | MCN to BON |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | Mean <br> Travel Time (d) | Travel <br> Rate (km/day) | $\begin{gathered} \text { Travel } \\ \text { Rate } \\ (\mathrm{km} / \text { day }) \end{gathered}$ | $\begin{gathered} \text { Travel } \\ \text { Rate } \\ \text { (km/day) } \end{gathered}$ |
| CJH Summer subs | 16-May | 16-May | F | 18.1 | 6.4 | 35.9 | a |
| Omak Pond subs | 23-May | 24-May | F | 13.4 | 11.2 | 32.0 | a |


| Wells FH subs | 14-May | 17-May | F | 22.0 | 3.1 | 28.4 | 57.2 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Wild subs | 12-May | 23-Jun | NA | 28.4 | 3.4 | 28.3 | a |
| CJH Summer | 15-Apr | 22-Apr | V | 14.7 | 7.9 | 27.4 | 59.0 |
| yearlings <br> Omak Pond yearlings <br> Carlton yearlings | 15-Apr | 22-Apr | V | 16.2 | 9.3 | 28.5 | 58.5 |
| 18-Apr | 18-Apr | F | 14.5 | 8.6 | 21.9 | 56.6 |  |
|  | 18-Apr | 27-Apr | V | NA | NA | $17.0^{\mathrm{b}}$ | 60.4 |
| CJH Spring Chk | 15-Apr | 22-Apr | V | 13.9 | 8.3 | 28.3 | 57.6 |
| RivP Spr Chk (10j) | 15-Apr | 21-Apr | V |  |  |  |  |
| Winthrop Spring Chk | 11-Apr | 11-Apr | F | 11.9 | 13.7 | 30.2 | 64.0 |
| Leavenworth Spr Chk | 21-Apr | 21-Apr | F | NA | 11.1 | 25.4 | 57.5 |

${ }^{\text {a }}$ sample size too small $(<10)$ to calculate an estimate
${ }^{\text {b }}$ Release to McNary, not Rocky Reach to McNary


Figure 31. Arrival timing at Rocky Reach Juvenile bypass (RRJ) of PIT tagged summer Chinook released from the Chief Joseph Hatchery and Omak Pond in 2016.


Figure 32. Arrival timing at Rocky Reach Juvenile bypass (RRJ) of PIT tagged Spring Chinook released from the Chief Joseph Hatchery (CJH) and Riverside Pond (RivP) in 2016.

One hundred ninety four (194) hatchery smolts with a PIT tag were detected at OKL between April 19 and September 17, 2016. Omak Pond sub-yearling Summer Chinook moved out the fastest, with $96 \%$ of the PIT detections within three days of release (Table 33. The majority ( $71 \%$ ) of yearling summer Chinook from Omak Pond were detected at OKL within 8 days of release, but it was approximately three weeks before $95 \%$ of the fish had passed (Table 32). For yearling Spring Chinook released from Riverside Pond, 95\% of the PIT detections occurred within eight days of release. One subyearling was not detected at OKL until September 17, it was unclear if this fish was precocial and had residualized or if it was expressing a fall migrant life history. Two yearlings (not included in Table 32) migrated to below Bonneville Dam in the spring and returned and were detected at OKL in September and therefore should be considered a minijack, not a residual.

Table 33. Detections of PIT tags at the lower Okanogan PIT array (OKL) from release groups of hatchery smolts on the Okanogan River in 2016.

| Release Group | Detection Date | Number Detected | Percent <br> Detected | Cumulative \% Detected |
| :---: | :---: | :---: | :---: | :---: |
| Omak Pond <br> Subyearlings <br> Release Date <br> 23 May 2016 <br> Forced | 5/11/2016 | 1 | 4\% | 4\% |
|  | 5/23/2016 | 16 | 62\% | 65\% |
|  | 5/24/2016 | 6 | 23\% | 88\% |
|  | 5/25/2016 | 1 | 4\% | 92\% |
|  | 5/26/2016 | 1 | 4\% | 96\% |
|  | 9/17/2016 | 1 | 4\% | 100\% |
| Omak Pond Yearling Summer <br> Chinook <br> Release Period <br> 15-22 April, 2016 <br> Volitional | 4/19/2016 | 1 | 2\% | 2\% |
|  | 4/20/2016 | 5 | 8\% | 10\% |
|  | 4/21/2016 | 13 | 21\% | 31\% |
|  | 4/22/2016 | 16 | 26\% | 56\% |
|  | 4/23/2016 | 9 | 15\% | 71\% |
|  | 4/24/2016 | 4 | 6\% | 77\% |
|  | 4/25/2016 | 3 | 5\% | 82\% |
|  | 4/26/2016 | 2 | 3\% | 85\% |
|  | 5/1/2016 | 1 | 2\% | 87\% |
|  | 5/2/2016 | 1 | 2\% | 89\% |
|  | 5/3/2016 | 1 | 2\% | 90\% |
|  | 5/5/2016 | 1 | 2\% | 92\% |
|  | 5/6/2016 | 2 | 3\% | 95\% |
|  | 5/9/2016 | 1 | 2\% | 97\% |
|  | 5/11/2016 | 2 | 3\% | 100\% |
| Riverside Pond Spring Chinook Yearlings <br> Release Period <br> 15-21 April, 2016 Volitional | 4/19/2016 | 2 | 2\% | 2\% |
|  | 4/20/2016 | 31 | 30\% | 32\% |
|  | 4/21/2016 | 44 | 42\% | 74\% |
|  | 4/22/2016 | 18 | 17\% | 91\% |
|  | 4/23/2016 | 4 | 4\% | 95\% |
|  | 4/25/2016 | 1 | 1\% | 96\% |
|  | 4/29/2016 | 1 | 1\% | 97\% |
|  | 5/4/2016 | 2 | 2\% | 99\% |
|  | 5/5/2016 | 1 | 1\% | 100\% |

## Smolt-to-Adult Return (SAR)

The most recent brood year that could be fully assessed (through age 5) for SAR was 2010. Based on expanded CWTs, the 2010 brood year had a SAR of $1.3 \%$, which was above the long-term average, but below the 5-year averages. However, this number may change as more adult captures from BY 2010 are uploaded to the RMIS database, and this table changes in the coming years to reflect those data (Table 34).

Table 34. Smolt-to-adult return rate (SARs) for Okanogan/Similkameen summer/fall Chinook, brood years 1989-2010.

| Brood Year | Number of tagged smolts released ${ }^{\mathbf{a}}$ | Estimated adult <br> captures $^{\mathbf{b}}$ | SAR |
| :---: | :---: | :---: | :---: |
| 1989 | 202,125 | 4,293 | $2.1 \%$ |
| 1990 | 367,207 | 972 | $0.3 \%$ |
| 1991 | 360,380 | 975 | $0.3 \%$ |
| 1992 | 537,190 | 2,282 | $0.4 \%$ |
| 1993 | 379,139 | 117 | $0.0 \%$ |
| 1994 | 212,818 | 1,528 | $0.7 \%$ |
| 1995 | 574,197 | 2,851 | $0.5 \%$ |
| 1996 | 487,776 | 31 | $0.0 \%$ |
| 1997 | 572,531 | 18,600 | $3.2 \%$ |
| 1998 | 287,948 | 7,687 | $2.7 \%$ |
| 1999 | 610,868 | 2,776 | $0.5 \%$ |
| 2000 | 528,639 | 6,762 | $1.3 \%$ |
| 2001 | 26,315 | 424 | $1.6 \%$ |
| 2002 | 245,997 | 1,975 | $0.8 \%$ |
| 2003 | 574,908 | 3,489 | $0.6 \%$ |
| 2004 | 676,222 | 12,896 | $1.9 \%$ |
| 2005 | 273,512 | 1,660 | $0.6 \%$ |
| 2006 | 597,276 | 13,626 | $2.3 \%$ |
| 2007 | 610,379 | 4,758 | $0.8 \%$ |
| 2008 | 604,064 | 14,932 | $2.5 \%$ |
| 2009 | 673,372 | 8,547 | $2.2 \%$ |
| 2010 | 650,137 | 8,504 | $1.3 \%$ |
| Total | $\mathbf{1 0 , 0 5 3 , 0 0 0}$ | $\mathbf{1 1 1 , 1 8 1}$ | $\mathbf{1 . 2 \%}$ |
| $5 \mathbf{y e a r} \boldsymbol{T o t a l}$ | $\mathbf{3 , 1 3 5 , 2 2 8}$ | $\mathbf{1 , 8 6 3}$ |  |
|  |  |  |  |

[^8]
## Spring-Chinook Presence and Distribution

## ENVIRONMENTAL DNA

CJHP collaborates with USGS to conduct Environmental DNA (eDNA) sampling and as one approach to monitor status and trends in spring-Chinook spatial distribution throughout the Okanogan basin in response to the reintroduction of the experimental population. Monitoring began prior to the reintroduction in an attempt to assess the premanagement action spatial distribution of spring-Chinook, allowing CJHP to assess the status and progress of the 10 (j)reintroduction efforts. While spring-Chinook were listed as extirpated within the Okanogan ESU, analysis of eDNA samples from tributaries (i.e. stream-type Chinook habitat) revealed that the basin likely does have a limited distribution of spring-Chinook. Additionally, PIT tag detections confirm the presence of occasional strays from out of basin (see PIT Tag Detections Section below).

As a proof of concept, sampling was initiated in 2012 with five mainstem Okanogan River sites and 11 Okanogan tributary sites as well as 32 sites throughout the Methow basin (See Laramie et al. 2015a and CJHP 2013 Annual Report). Sampling was conducted in June and August 2012 at all sites. In 2013, sampling was conducted only in the Okanogan basin, at eight additional tributary sites not visited during the proof of concept study. These sites were sampled in June and were located in tributary streams with potential for springChinook recolonization. In 2014, all previously sampled sites in the Okanogan basin were re-visited and sampled (U.S. sites on 12-13 July, 2014, and Canada sites on 2 October 2014). All sampling was conducted following the methods and protocols described in Laramie et al 2015b, and available as PNAMP Method ID\# 5476
(www.monitoringresources.org/Document/Method/Details/5476). See Appendix C for results from 2012 thru 2014 eDNA analyses. Several tributaries have produced consistent annual detection of Chinook eDNA, including Shingle Creek, Vaseux Creek, Salmon Creek and Omak Creek. No sampling was conducted in 2015. In 2016, 15 sites were re-sampled to monitor status and trends in spatial distribution during the early stages of the reintroduction effort (Table 35).

Table 35. eDNA results for sampling conducted in Okanogan basin tributaries from 20122016.

| Site | 2012 | 2013 | 2014 | 2015 | 2016 | Indicates Presence | + |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Loup Loup Creek |  | - | + |  | + | Indicates not-detected | - |
| Omak Creek (above falls) | - |  |  |  | + |  |  |
| Omak Creek (near mouth) | + |  | + |  | + |  |  |
| Salmon Creek | + |  | + |  | + |  |  |
| North Fork Salmon Creek | - |  | - |  |  |  |  |
| West Fork Salmon Creek | - |  | - |  |  |  |  |
| Wanacut Creek |  | - |  |  | - |  |  |
| Tunk Creek |  | - |  |  | + |  |  |
| Aeneas Creek |  | - | - |  | - |  |  |
| Bonaparte Creek | + |  | - |  | - |  |  |
| Siwash Creek |  | + |  |  |  |  |  |
| Tonasket Creek |  | + |  |  | - |  |  |
| Antoine Creek |  | - | + |  | + |  |  |
| Similkameen River |  | + | + |  |  |  |  |
| Ninemile Creek | - |  | - |  | + |  |  |
| Inkaneep Creek | + |  | - |  | - |  |  |
| Vaseux Creek | + |  | + |  | + |  |  |
| Shuttleworth Creek | - |  | - |  | - |  |  |
| Shingle Creek | + |  | + |  | - |  |  |
| Okanogan River (near mouth) | + |  | + |  |  |  |  |
| Okanogan River (above Salmon Cr.) | + |  | + |  |  |  |  |
| Okanogan River (above Siwash Cr.) | + |  | + |  |  |  |  |
| Okanogan River (Inkaneep Cr.) | + |  | + |  |  |  |  |
| Okanogan River (at Shuttleworth Cr.) | - |  | + |  |  |  |  |

## PIT TAG DETECTIONS

PTAGIS contained 65 unique records of adult and jack spring-Chinook detected in the Okanogan basin in 2014 (Table 36). One tag was filtered out of the summary because it appeared to be a fish harvested in Icicle Creek and discarded in Omak Creek. Of the 64 remaining unique detections in the Okanogan and Similkameen, the majority ( $n=45 ; 70 \%$ ) were hatchery fish that had been tagged at Wells Dam as an adult then detected in the Okanogan, primarily at the lower Okanogan array. Twelve of the fish (19\%) detected in the Okanogan were re-detected in the Methow basin at a later date.

Adult spring Chinook were detected at the Lower Okanogan array between April 10 and September 30, 2016 with a median run timing of June 5. Eight fish were detected at Zosel Dam between June 9 and September 22, with four of those fish being detected at a later date in a Canadian tributary (Shingle and Shuttleworth creeks) , the array at Vertical Drop Structure \#3 on the Okanagan upstream of Lake Osoyoos, and at Skaha Dam (Table 36). None of the fish tagged and classified as spring Chinook appeared to be mis-classified summer Chinook (there were 3 such fish in 2014). We did not evaluate fish tagged as summer Chinook or Chinook with undetermined race to assess if they might be spring Chinook.

Thirteen fish from the Riverside Pond release in 2015 (BY 2013) were detected in the Okanogan. These fish were all 3 year old jacks. Five of them (38\%) were detected in Omak Creek, one in Tunk Creek and two at Zosel Dam (Table 36).

There were 12 PIT detections in the tributaries of the Okanogan in 2016. Omak Creek had eight detections (67\%), Loup Loup Creek had two (17\%) and Salmon and Tunk creeks had one each (8\%).

Six of the 64 fish (9\%) had been tagged and released as juveniles somewhere outside the Okanogan, including four that were released from hatchery facilities in the Methow, one natural origin fish from the Methow and one natural origin fish from the Entiat (Table 36).

Table 36. Final PIT tag detections in the Okanogan for spring-Chinook in 2016. OKC/VDS3 is at vertical drop structure 3 in British Columbia upstream of Lake Osoyoos and OKL is the detection array in the lower Okanogan River near Malott, Washington.

| Origin | Release <br> Location | Final Detection Location(s) |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Okan. <br> then Methow | Okan then Other (besides Methow) | OKL | Salmon Ck. | Omak Ck. | Other U.S. Trib | Zosel Dam | OKC / <br> VDS-3 | B.C. Trib |
| Hatchery Spring Chinook | Methow/ <br> Winthrop Juvenile |  |  | 4 |  |  |  |  |  |  |
|  | Wells Dam Adult | 12 |  | 22 | 1 | 3 | 2 | 2 | 1 | 2 |
|  | Bon Dam Adult |  |  |  |  |  |  |  |  |  |
|  | CJH |  |  |  |  |  |  |  |  |  |
|  | $\operatorname{RivP}(10 \mathrm{j})$ |  |  | 5 |  | 5 | 1 | 2 |  |  |
|  | Other |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Natural <br> Spring <br> Chinook | Wells Dam Adult |  |  |  |  |  |  |  |  |  |
|  | Twisp R. Juv. |  |  |  |  |  |  |  |  |  |
|  | Entiat R. Juv. |  |  |  |  |  |  |  |  | 1 |
|  | Rock Isl. <br> Dam Juv. |  |  |  |  |  |  |  |  |  |
|  | Methow R. Juv. |  |  | 1 |  |  |  |  |  |  |
|  | Okan R. Juv. |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Summer <br> Chinook <br> (mis- <br> classified) | NA |  |  |  |  |  |  |  |  |  |
|  | NA |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
|  | Total | 12 | 0 | 32 | 1 | 8 | 3 | 4 | 1 | 3 |

*note: : one tag (3D9.1C2DF32F76) was excluded from this summary because it appeared to be a fish harvested in Icicle Creek and discarded in Omak Creek.

## DISCUSSION

## Rotary Screw Traps (RST)

The pooled trap efficiency of approximately $0.1 \%$ is lower than in previous years (Rayton and Arterburn 2008, Johnson and Rayton 2007; https://static1.squarespace.com/static/56f45574d51cd42551248613/t/57c06a21e58c62 290279a3d7/1472227873603/2006 Screw Trap Report Final.pdf; https://static1.squarespace.com/static/56f45574d51cd42551248613/t/57c06a12e58c62 290279a376/1472227860447/2007RstReportFinal.pdf),and remains insufficient to precisely estimate juvenile production for the basin. Additionally, the $95 \%$ confidence interval for hatchery-origin population did not capture the total known number of hatchery-origin fish released upstream of the RST $(867,679)$. This indicates that, due to the difficulties in accurately estimating trap efficiency and juvenile production, the results of screw trapping activities in 2016 are unlikely to provide an accurate estimate of juvenile production.

NOAA Fisheries suggested a goal for precision of juvenile outmigration monitoring was to achieve a coefficient of variation (CV) of 15\% or less (Crawford and Rumsey 2009). It is not clear that this level of precision is attainable in any large river system using conventional sampling methods such as a rotary screw trap (see Scofield and Griffith, 2014). Still, improving trap efficiency and narrowing juvenile emigration estimates remains the goal of CJHP such that informed management decisions can be made. Environmental factors such as river discharge, configuration, and trap size influenced the efficiencies of these trials. In order to mitigate these confounding variables, we will attempt to conduct more frequent efficiency trials with large release groups ( $n \geq 500$ ).

Again, no relationship between Okanogan River flow and trapping efficiency was observed, and the flow regression model used by other agencies in other river systems (Murdoch et al. 2012) was not applied to estimate outmigration. The CJHP will continue to assess methods to improve capture techniques to increase the precision of juvenile production estimates.

Differing efficiency rates for trials involving yearling and sub-yearling fish indicate that using hatchery releases of yearling fish, as a surrogate to measure natural production would be inappropriate. However, in future years when wild spring Chinook yearlings are present, this possibility could be reexamined. This should be especially relevant once integrated, $\S 10(\mathrm{j})$ spring Chinook, first released from the Riverside Acclimation pond in April 2015, begin to return and presumably spawn.

Finally, Pacific lamprey (Entosphenus tridentatus) were captured in the RST in both 2006 and 2007, but were not observed from 2008 to 2016. The status of this fish, an
important cultural and ecological resource in the Okanogan River Basin is not examined in this report, but its disappearance from the RST is notable.

## Juvenile Beach Seine

The CJHP took over the beach seining effort in 2014, adopting methods used by Douglas County PUD and Biomark in 2011-2013. Given the low catch rate of taggable summer/fall Chinook from the RST, beach seining appeared to be a more reliable opportunity to capture large numbers of taggable summer/fall Chinook juveniles. Again in 2016, PIT tags deployed at the beach seine far outnumbered tags deployed at the RST.

Mortality related to capture, handling and tagging was lower than in past years ( $5.3 \%$, compared to $13 \%$ in 2015). Maintaining water temperatures below $18{ }^{\circ} \mathrm{C}$, reducing MS-222 concentrations in the anaesthetizing solution, and further limiting handling time during tagging and capture likely decreased our post tagging mortality. Although mortality rate has decreased, the hope for future years is to continue to reduce it.

Fish size increased through the tagging period, but the number of fish captured and CPUE began to decrease in mid-June, earlier than in previous years. Similarly, dates of detection at downstream PIT arrays occurred earlier in 2016 than in previous years. This may be explained in part by the size of tagged fish, differences in flow regime, or other factors that can be examined in future years when a more robust, multiyear dataset has been developed.

Although capture locations in 2016 were limited to areas near the confluence of the Okanogan and Columbia Rivers, fish were captured in areas that could also be used by juveniles originating from Methow and Columbia River spawning areas. Therefore, future analyses of returning adults will need to take this into account by recognizing that some fish may not be destined for the Okanogan. In the future, stable isotope analysis may be used to determine which proportion of tagged fish have an origin in the Okanogan basin, and which come from the mainstem Columbia or other river systems.

## Lower Okanogan Adult Fish Pilot Weir

Discharge conditions on the Okanogan River in 2016 were higher than in 2015, delaying installation and operation of the weir until early August, which was a couple weeks later than 2015. Temperatures on the Okanogan River were fairly normal, compared to the 10 year median. Because temperatures stayed below $22.5^{\circ} \mathrm{C}$ during daylight hours throughout the season, trapping operations were not suspended. Tower and bank fish observations were generally higher after the thermal barrier broke on August 22. During this time, fish observations 0.8 km below the weir, at the lower pool, were higher than observations at the weir. When river temperature was lower and gauge height was less than 3 feet, Chinook were more likely to mill in deeper pools. Continued monitoring of Chinook passage through the weir with respect to temperatures should continue in order to better refine weir operations and future expectations for weir effectiveness. The number of Chinook handled at the weir $(\mathrm{n}=169)$ was more than in $2015(\mathrm{n}=54)$ when wild fires prevented operation of the weir during and immediately after the thermal barrier break and but less than $2014(n=2,324)$. Configuration and operation of the weir was similar in 2014 and 2016 and it is unclear why more fish did not enter the trap in 2016, as they did in 2014, given the number of fish observed in the pools downstream of the weir. We intend to modify the trap entrance in 2017 to attempt to improve trapping efficiency.

None of the water quality parameters monitored were at a level that would cause concern regarding an environmental effect of the weir on water quality. The number (54) of dead fish at the weir was lower in 2016 than previous years. Mortality was highest a week before the peak trapping period, indicating that trap operation and handling were not the cause of mortality. The behavioral observations and lack of fish impinged between pickets (head upstream) were good indicators that this weir configuration and picket spacing were not a major cause of direct mortality. In an attempt to assess immediate indirect mortality, we marked and released 117 adult natural-origin Chinook at the weir trap. Two of them were collected as mortalities at the weir, which is less than $1.7 \%$ of all trapped, marked and released fish. Additionally 4 of them were collected as carcasses on the spawning grounds (3.4\%) and 2 were collected as broodstock for the integrated program (1.7\%). We recommend additional mark-recapture studies to continue to assess indirect mortality at the weir in future years.

There were few observations of Sockeye at the weir during daylight hours and only three were trapped in 2016. Most sockeye passed the weir before trapping began on August 22. We did observe a few sockeye ( $\sim 20$ ) pass through the weir trap during night hours (2000-0400) during the first couple weeks of trapping operations. It is likely that more sockeye moved through the weir panels at night when observations did not occur. An estimated 164 jack and small adult Chinook escaped through the 2.5 " weir panels that were
intended to allow Sockeye passage. We recommend testing a weir configuration that does not include the 2.5 inch weir panels to increase the efficiency of Chinook trapping without causing too many Sockeye to also use the trap. In 2016, there were very few Sockeye observations during daylight hours, but in past years we did have observations of sockeye passing through the 2.0 " picket spacing. We will continue to document passage of Sockeye and Chinook through all picket spacings.

There was no way to know exactly how many fish escaped past the weir before it was installed or how many fish swam through, around or jumped over the wings after it was installed (jumping over the wings has never been observed). The potential weir effectiveness measure of $1.39 \%$ was low because, after reviewing PIT detection at the Okanogan Instream Lower array, we suspect that about 40-50\% of the fish had migrated past the weir before deployment in August. There were two thermal barrier breakdowns that occurred before the weir was fully functional. The first breakdown occurred in midJuly and a second, smaller one occurred a week before weir deployment. Although the second breakdown only occurred over a few days before the temperature increased above $22.5^{\circ} \mathrm{C}$ again. Unfortunately river discharge was too high ( $\geq 2,500 \mathrm{cfs}$ ) during the thermal barrier breakdown to install the weir under its current design, thus limiting the weir as an effective tool for pHOS management in 2016. Fortunately, this did not hinder fish management objectives in 2016 because pHOS was already low and only $20 \%$ of the Chinook trapped were hatchery origin. In the future, with larger returns of hatchery fish due to CJH releases we anticipate a much higher pHOS at the weir resulting in higher weir effectiveness. Continuing these evaluations in future years will be critical to determining the long-term viability of the weir as a fish management tool for summer Chinook.

The broodstock collection protocol at the weir was to get $15 \%(n=85)$ of the integrated program from the later arriving fish (in September, post thermal barrier). The weir failed to meet its broodstock goal, collecting only 16 fish, or $19 \%$ of the broodstock collection protocol, through the trap post thermal barrier breakdown period. The Whooshh ${ }^{\text {TM }}$ transport tube worked well and initial mortality was low, indicating that further use and testing of this system should be continued in the future.

Although the weir was not very successful at trapping Chinook in 2016, CCT F\&W staff were able to safely and successfully deploy, operate, and monitor the weir and add to the multi-year evaluation of the weir as a fish management tool for the CJH program. The weir's importance to the Okanogan summer/fall Chinook population will increase in the coming years with larger hatchery returns resulting from the increased production at CJH. Experiencing a broad range of environmental conditions spanning the extremely high summer flows of 2012 to the very low and warm flows in 2015 is important for understanding the range of challenges and resulting weir effectiveness that can be expected through time.

RECOMMENDED WEIR AND TRAP CHANGES FOR 2017.— In December 2016, the CJHP Science Program staff convened a post- season review group to discuss operations and recommendations for improvements/changes. The entire season was reviewed and subsequently, data were reviewed with results appearing in the text of this document. A summary of the 2016 weir operations was presented at the $7^{\text {th }}$ Annual Chief Joseph Hatchery Program's Annual Review. This presentation is posted on the programs website at: https://www.cct-fnw.com/annual-program-review/

The following list of changes has been built into this Plan and the CCT/GPUD/BPA Funding Agreement 430-3128 - Amendment No. 3. We envision both pre and postseason weir meetings being called in the same manner as 2015 and 2016, to occur this year. The following recommendations are derived from the 2016 post- season analysis and the subsequent findings from CCT's research, monitoring and evaluation activities:

1. Trap entrance chute (similar to Washougal design)
2. Whooshh ${ }^{\text {TM }}$ system air hose support structures
3. Trap fyke restraint brackets
4. Consider alternative power source locations
5. Consider alternative trap locations
6. Add two more sections of trap walkway
7. Additional side apron support tripods
8. Trap access ladder for personnel
9. Recessed video and lighting housings
10. Trap leveling attachments
11. Crowder extension
12. Install weir panels adjacent to trap box and direct water velocity through the trap

## Redd Surveys

Summer Chinook spawning activity was high again in 2016, with the highest redd count observed in the Okanogan River Basin since 2002 and above average redd counts in five out of the eight survey reaches (Table 17).

The redd count in reach 0-6 was the highest on record, and spawning in reach S-1 was the second highest on record. These two adjacent reaches continue to provide the primary spawning habitat for summer Chinook in the Okanogan/Similkameen basin. One objective of the CJHP is to increase the spatial distribution of spawning into the lower reaches of the Okanogan, where historically, a low proportion of the spawning activity has occurred. 2016 redd counts showed substantial increase in spawning in the lower Okanogan reaches (0-1 thru 0-3), but 2016 counts were a return to the lower average. As we begin to see returns from CJHP Chinook reared at the Omak pond acclimation site, we anticipate that those fish will provide increased spawning in lower reaches through natal homing. Continued monitoring of redd and carcass distribution will be critical to evaluation of this metric.

Spawn timing was similar to previous years, with the intensive spawning beginning the first week of October, coinciding with mean water temperatures dropping below $15^{\circ} \mathrm{C}$. Few redds were constructed in November (Table 19), suggesting that, if extant, a latearriving fall-run population of Chinook is contributing minimally to the population at large. Although aerial surveys contribute a relatively small portion of the observed redds compared to ground or float surveys, they remain an important tool for documenting spawning, or lack of, in areas not accessible by ground crews.

The fish per redd expansion was based on the sex ratio of fish passing Wells Dam. This method has been used since at least 1998 (Hillman et al. 2014) and is still being applied to both the Methow and Okanogan populations. However, there is uncertainty that the combined sex ratio of hatchery and wild summer Chinook at Wells Dam is representative of the Okanogan population because it includes Methow returns as well as mainstem released hatchery fish and downstream hatchery and wild fish. If the Okanogan has a different ratio of precocial males (jacks) than that of the Wells count, then the Okanogan abundance estimate would be biased. We suggest exploring other approaches to estimating the number of fish per redd in the Okanogan and Similkameen Rivers.

## EsCAPEMENT INTO CANADA

Escapement of summer/fall Chinook into Canada had been largely overlooked until recently, when the video counts of Chinook passing over Zosel Dam increased to a level where OBMEP staff brought the results to the attention of CJHP staff. Spawning escapement in Canada is still unknown, as the video counts represent run escapement and the relationship between run escapement and spawn escapement is not clear. However, a
substantial number of Chinook have been counted passing Zosel Dam, 1,823 in 2016 (Table 20). No formal Chinook spawning grounds surveys are currently being conducted, but surveys for Sockeye ( 0. nerka) occur annually. Biologists in Canada have observed small numbers (i.e., substantially fewer than the Zosel Dam video counts) of Chinook spawners building redds in the Canadian portion of the Okanogan River (R. Bussanich, ONA, pers. comm., 2014). This discrepancy has at least three possible explanations that need to be further explored in the coming years.

1) Chinook can migrate downstream through Zosel Dam without being detected in the fishways video monitoring system.
2) Chinook are making it to spawning areas in the Canadian Okanogan and not being detected by Canadian spawning ground surveys. These surveys currently target sockeye, but the spawn timing and potential spawning areas are similar.
3) High pre-spawn mortality kills fish between passage at Zosel Dam and potential spawning grounds somewhere in Canada.
Some possible solutions to exploring these explanations include:
a) Evaluate PIT tag results for fish that might ascend through the fishways multiple times (this will not account for fish that fall back once and do not re-ascend).
b) Conduct surveys in the Canadian portions of the Okanagan River targeting Chinook spawning grounds (i.e. larger substrate) during peak spawning (mid-October).
c) Conduct carcass surveys above Zosel Dam, throughout Lake Osoyoos and the Canadian Okanogan looking for pre-spawn mortality.
d) Capture and radio tag fish in the Zosel fishways.

Until a definitive method is developed for estimating spawn escapement in Canada, the CJHP will continue to monitor and report run escapement via video monitoring. However, we will not add run escapement past Zosel Dam to spawn escapement in the U.S. because this could overestimate total spawners if explanation 1 or 3 (see list above) are true.

## Carcass Surveys

2,755 carcasses were recovered out of an estimated 10,605 spawners, which exceeded the target carcass recovery rate of $20 \%$. Zhou (2002) reported fish length as a significant factor in carcass recovery probability, with larger fish recovered at a higher rate than smaller fish. This is especially important as it relates to precocious males, or jacks, which are expected to occur with higher frequencies in hatchery-origin Chinook. Failing to assess and correct for biases and population discrepancies could lead to potential underestimation of hatchery-origin Chinook survival (resulting in inflated hatchery production) or over-estimation of wild-origin Chinook survival (masking potentially negative effects of the hatchery program) (Murdoch et al. 2010). We are considering methods (e.g. mark-recapture) to assess and quantify potential size bias in our carcass recovery efforts.

Surveys in September revealed few carcasses attributable to pre-spawn mortality (PSM) and October surveys found few PSM as well ( $\sim 3 \%$ ). It is likely that the majority of PSM occurs earlier in the season while water temperatures are higher and are a greater risk to fish attempting to travel to or hold near the spawning grounds. If this were true, the current design of our redd/carcass surveys would provide an underrepresentation of actual PSM. Therefore, egg retention and pre-spawn mortality results should be interpreted cautiously. The carcasses of fish that died prior to the onset of spawning and before sampling began may have been carried downstream of recovery floats, consumed by scavengers, or covered with sediment, making them unavailable for sampling or harder to detect and collect. This could result in an underestimation of pre-spawn mortality. The protocol assumes that each female may contain 5,000 eggs and were only considered prespawn mortality if they retained all 5,000 eggs. A static fecundity assumption may not be the best approach because younger and smaller females will likely have fewer eggs. Additionally, the current assumption used by the CJH during in-hatchery spawning activities for average fecundity is 4,600 eggs. We expanded the assessment to include an evaluation of fish that retained greater than 1,000 eggs as an attempt to capture some of the variability in fecundity and situations where fish died before depositing a biologically important portion of their eggs. We are not sure that 1,000 eggs are biologically important, but clearly there should be some amount of egg retention that matters besides $100 \%$. We suggest continued review and modification of the egg retention estimation methods/protocol in the future.

## PHOS AND PNI

The biological target for CJHP is to maintain a 5- year average pHOS <0.3. 2015 was the first year since the CJHP began monitoring the population that the 5-year average ( 0.30 ) met this objective. $2016 \mathrm{pHOS}(0.15)$ further reduced the 5 -year average to 0.23 . The program met the biological target for PNI ( $>0.67$ ) for the second year in a row. The 5-year
mean (0.87) met the objective, and PNI continues to improve. In the future, we suggest that continued aggressive removal of hatchery fish through selective fisheries and adult management at the weir and hatchery ladder given the uncertainty regarding the adequacy of the objectives to meet long-term population conservation goals. Exceeding the targets whenever possible also provides a buffer for years when goals may not be achieved due to low run size or challenging environmental conditions.

## Origin of Hatchery Spawners

Hatchery-origin fish recovered on the spawning grounds in the Okanogan Basin were predominantly (95\%) from the Okanogan Basin releases. Stray hatchery-origin fish from outside the Okanogan made up only 1\% of total estimated spawners. Likewise, Okanogan Basin hatchery-origin fish strayed to other areas at a low rate ( $0.7 \%$ ) and were a small percentage of the spawner composition in other Upper Columbia tributaries. Eight fish released from Similkameen Pond were detected on spawning grounds surveys at the Methow and Chelan Rivers in 2016 (Table 27). Stray rates and hatchery spawner composition were within the target levels for the program both within and outside the Okanogan Basin. Fish released within the Okanogan River basin have consistently homed to their natal stream, and 2016 was not an exception.

## Smolt to Smolt Survival and Travel Time

The survival results for each release group provide a useful index of annual survival for comparison between release groups and, in the future, between years. Statistical tests were not conducted to determine if observed differences were statistically valid because we believe this should be done with a multi-year data set. Targets for post release survival have not been established, but it was encouraging to see that the 2016 estimates of CJH programs were similar to nearby programs. In the future, with $4-5$ years of smolt migration data the program should develop a statistical framework for evaluating smolt-tosmolt survival and establish targets that could be used to help adaptively manage the release strategies, if it is determined that survival or travel time are not adequate to meet program goals. Similar to 2015, it appeared that all hatchery fish migrated out of the system relatively quickly in 2016, with no detections of yearling migrants in July and very few detections of sub-yearlings in August. Unfortunately it is not possible to evaluate juvenile outmigration (or movement within the Columbia River) in the winter months because juvenile bypass facilities do not operate year round.

The vast majority (82-95\%) of PIT tagged hatchery smolts released from Omak Pond and Riverside Pond migrated to the lower Okanogan River within one week of release. There was only one detection of a juvenile hatchery fish at OKL after May 26. This assessment suggests that the program was successful at releasing actively migrating smolts. This analysis did not attempt to account for detection probability at OKL. It is
likely that the detection rate was different throughout the time period when smolts were detected. However, detection rates at large river arrays generally increase with decreased flow, so late arriving fish would have a better chance of being detected at OKL than fish outmigrating during high flows from April to June. Therefore, it is not likely that a meaningful number of late migrating smolts or residual hatchery fish would have crossed OKL when compared to what was detected during peak migration. Although the OKL PIT detection site is 25 km from the confluence with the Columbia River, it is very close ( $\sim 2 \mathrm{~km}$ ) to the inundated zone of Wells Pool. Therefore we can assume that smolts crossing OKL do represent fish leaving the Okanogan River system, or at least they are entering a more reservoir-like environment where interspecific competition for food and space is likely to be less than in the river.

## Smolt-to-AdUlt Return

SAR for the most recent full brood returns (2010) was slightly below the 5-year and long-term averages. It is likely that the SAR estimate is biased low because some recovery efforts were not expanded within RMIS, and also because some fish likely have yet to return. We had no way to obtain information necessary to do these expansions or to even speculate as the magnitude of the potential error introduced because of it. In the future, we suggest also using PIT tags as an independent, additional estimate of SAR.

## Spring-Chinook Presence and Distribution

Environmental DNA (eDNA) surveys have been an important tool for monitoring the early stages of the spring Chinook reintroduction effort. CJHP has developed an annual eDNA survey strategy that allows for basin-wide spatiotemporal distribution assessments. eDNA surveys have confirmed a wide distribution of spring Chinook in the Okanogan River basin, including 11 tributaries in U.S. and Canada. This effort has been successful at identifying and prioritizing tributaries for future spawning ground surveys. Implementing eDNA sampling at a finer scale, within those tributaries that have indicated spring Chinook presence would help to locate spawning areas and/or reaches that would be most appropriate for more intensive survey efforts, such as visual redd surveys. Additionally, eDNA surveys conducted in winter or early spring could help to confirm successful spawning in a tributary, as a positive detection during this time of year would likely be the result of juvenile presence.

PIT tag have been another important tool for monitoring the progress of reintroduction efforts. Future analysis should include an estimate of run escapement to the Okanogan using fish PIT-tagged at Wells Dam as the mark group and those redetected in the Okanogan as the recapture group. Similar to previous years, a portion (19\%) of the Spring Chinook detected in the Okanogan had a final destination in the Methow River basin, so it will be important to exclude those fish from escapement estimates.

## Adaptive Management and Lessons Learned

## The Annual Program Review (APR)

Each year the CJHP hosts a workshop to review and present findings from the previous year and plan for the upcoming fish production and science monitoring cycle. The APR was convened in March 2017 with the purpose of reviewing data collection efforts and results from 2016 and developing the hatchery implementation and monitoring plan for 2017 (Figure 33). This effort is focused on using adaptive management to guide the program. After a series of presentations highlighting the data collection activities and results, the group (CJHP staff and invited guests from Federal, State, PUD, and other organizations) used the In-Season Implementation Tool (ISIT) during the "Analysis" step (Figure 34). The group reviewed the ISIT input parameters for key assumptions, status and trends and decision rules to be sure that the best available information was included in the model. ISIT then used the pre-season Upper Columbia summer/fall Chinook Salmon forecast to provide an estimate of how the program could be implemented with respect to broodstock collection, harvest, weir and hatchery ladder operations to achieve biological targets for 2017. APR materials with more details than what is provided within this report can be found at https://www.cct-fnw.com/annual-program-review/.

## Key Management Questions

Answering key management questions is an essential function of the CJHP and is central to the analysis and reporting steps in both the APR and this annual report. Management questions inform the development of the RM\&E activities, the CJHPs Key Management Questions (KMQs) are:

1. What is the current status and recent historical trend of the naturally-spawning population in terms of Viable Salmonid Population (VSP) parameters? ${ }^{8}$
2. What is the current status and recent historical trends for hatchery returns and harvest?
3. Is the hatchery program meeting target in-hatchery performance standards?
4. Are the hatchery post-release targets met for survival, catch contribution and straying?
5. Are targets for total catch contribution and selectivity for HORs met?
6. Are there negative effects of the hatchery on the natural population?
7. Are assumptions about natural production potential valid?

[^9]8. How should the program be operated in the coming year?

## Annual Planning Workflow



Figure 33. The Chief Joseph Hatchery's annual planning process and work flow.


Figure 34. The Chief Joseph Hatchery's analytical work flow.

## 2017 Run Size Forecast and Biological Targets

Run-size forecasts and updates are an early indicator for the biological targets for the coming season, through the Decision Rules outlined in the ISIT. The preseason forecast is based on brood year escapement and juvenile survival indicators and is generated through the Technical Advisory Committee (TAC) to the U.S. v. Oregon fish management agreement. As the season nears, this information is supplemented with return data from downstream dam counts. The pre-season forecast for Upper Columbia summer Chinook Salmon was 74,100 . The pre-season forecast, and subsequent run updates from early dam counts, were used to predict the NOR and HOR run size for the Okanogan population. Hatchery broodstock and selective harvest targets are determined based on these estimates and the objectives for $\mathrm{pHOS}(<0.30)$ and $\mathrm{PNI}(>0.67)$. A regression analysis conducted within ISIT in preparation for the APR predicted that the pre-season forecast of 74,100 upper Columbia would yield 7,627 NORs and 6,413 HORs (Figure 35). The harvest and broodstock collection goals were established from this prediction. With a NOR run size over 5,000 the broodstock collection recommendation for the integrated program was full production ( 616 NOB) with $100 \%$ pNOB (Figure 35). Likewise, the segregated program should achieve full production with 503 HOB. The model predicted that 1,889 HORs would be captured in the terminal (above Wells Dam) fisheries and that 116 HORs could be removed at the weir. These efforts could result in 6,345 NOS and 2,146 HOS for a pHOS of $21 \%$ and a PNI of 0.82 . Under this modeling scenario the biological targets would be met in 2016. As run size updates become available (through TAC) the ISIT outputs will be double checked until the final in-season check point on July 15, 2017. At that time the run size at Wells Dam will be input into ISIT and the final plan for broodstock and harvest will be updated. If the July 15 update includes more hatchery and natural fish than predicted, then harvest and removal of surplus fish at the weir and the hatchery ladder will be implemented by CCT and WDFW (through their mark-selective sport fishery).


Figure 35. The in-season updates management worksheet used to set biological targets for the upcoming year (2017) in the In-Season Implementation Tool.

## 2017 Key Assumptions

The CJHP reviews the key assumptions (working hypothesis) each year at the APR workshop. These assumptions directly affect the decision rules used to guide in-season management decisions. The program documents the changes and uses this information for future review and analysis (Figure 36).

KEY ASSUMPTIONS-AHA


Figure 36. The key assumptions worksheet used in the 2016 In-Season Implementation Tool for the CJHP planning at the Annual Program Review

## 2016 Status and Trends

The recent performance of the population is a primary driver for determining how the hatchery program should be operated in the future. This was accomplished by updating and reviewing the status and trend information within five categories: (1) natural production, (2) hatchery production, (3) harvest, (4) migration, and (5) habitat (Figure 37).

| Retum year |  | FPC Reported Dam Count at Wells thru 7/15 (excludes jacks) | \%of <br> final <br> count | $\begin{gathered} \hline \text { PUD Counts at Wells } \\ \text { Dam } \\ \hline \end{gathered}$ |  | Okanogan Origin Fish to Wells Dam |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | NORAll Origins (excludes jacks) |  | $\begin{gathered} \text { HoR All } \\ \text { Origins } \\ \text { (excludes } \\ \text { jacks) } \end{gathered}$ | $\begin{aligned} & \text { Okan, } \\ & \text { NORs } \end{aligned}$ | Okan. HORs |
| 1998 | 3 |  | 1,060 | 0.25 | 970 | 5,519 |  |  |
| 1999 | 4 | 999 | 0.11 | 2,708 | 4,580 | 1,426 | 2,668 |
| 2000 | 5 | 2,266 | 0.26 | 2,726 | 7,398 | 1,111 | 2,257 |
| 2001 | 6 | 9,766 | 0.24 | 10,266 | 19,195 | 4,543 | 6,984 |
| 2002 | 7 | 23,221 | 0.34 | 24,138 | 42,035 | 5,060 | 11,757 |
| 2003 | 8 | 20,564 | 0.40 | 9,194 | 7,373 | 2,434 | 2,937 |
| 2004 | 9 | 14,762 | 0.40 | 23,227 | 13,989 | 7,716 | 2,598 |
| 2005 | 10 | 14,449 | 0.42 | 18,911 | 15,164 | 8,259 | 3,401 |
| 2006 | 11 | 12,563 | 0.43 | 20,262 | 8,730 | 8,348 | 4,113 |
| 2007 | 12 | 5,532 | 0.37 | 7,088 | 7,789 | 4,466 | 2,899 |
| 2008 | 13 | 8,838 | 0.35 | 11,244 | 13,779 | 4,311 | 6,368 |
| 2009 | 14 | 13,753 | 0.46 | 15,184 | 14,187 | 5,561 | 5,673 |
| 2010 | 15 | 12,264 | 0.41 | 5,671 | 7,167 | 4,541 | 5,394 |
| 2011 | 16 | 3,912 | 0.12 | 12,139 | 19,164 | 5,116 | 6,419 |
| 2012 | 17 | 10,082 | 0.24 | 14,424 | 27,716 | 6,271 | 7,168 |
| 2013 | 18 | 25,571 | 0.38 | 34,965 | 30,179 | 8,305 | 8,636 |
| 2014 | 19 | 26,010 | 0.39 | 36,060 | 21,015 | 12,797 | 7,555 |
| 2015 | 20 | 25,153 | 0.38 | 46,030 | 31,625 | 13,567 | 14,332 |
| 2016 | 21 | 21,479 | 0.32 | 28,467 | 21,542 | 10,03 | 10,572 |
| 2 2vil | 22 | 10,124 | 0.20 |  |  |  |  |



Figure 37. The status and trends worksheet in the In-Season Implementation Tool for CJHP planning at the Annual Program Review.

## 2016 Decision Rules

The decision rules determine the targeted size of the hatchery program and the management of natural escapement abundance and composition. The purpose of the Decision Rules is to assure that the CJHP manages the hatchery, terminal fisheries and weir to meet the guidelines for abundance, spawner composition, and distribution of the natural spawning escapement (Figure 38).

|  | BIOLOGICAL TARGETS AND "PHASE TRIGGERS" |  | Population Designation: Primary |  |  | (from Decision Rules) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Applied Scenario |  |  |  | Phase 1 | Phase 2 | Phase 3 |
|  |  |  |  | Recolonization | Local Adapt. | Recovered |
|  | Biological Triggers for Phase Change Rules | Transition 1 | Baseline | Transition 1 | Transition 2 | Long term |
|  | Year | 2020 | 2013 | 2020 | 2025 |  |
|  | Move up one phase if NORs greater than: | 5,250 | 1,000 | 5,250 | 7,000 |  |
|  | Move down one phase if NORs less than: | 800 | - | 800 | 3,000 | 6,000 |
|  | Based on N -Year Running Average, where $\mathrm{N}=$ | 5 | [Enter integer between 3 and 10, inclusive] |  |  |  |
| Management Control Variables for "Sliding Scale" Rules |  | Transition 1 | Baseline | Transition 1 | Transition 2 | Long term |
| Integrated <br> Program | Minimum NOR escapement | 800 | 800 | 800 | 800 | 800 |
|  | Smallest viable hatchery program | 100,000 | 100,000 | 100,000 | 100,000 | 100,000 |
|  | Max \% of NORs used for Broodstock | 30\% | 30\% | 30\% | 30\% | 30\% |
|  | Maximum Yearling Releases | 800,000 | 250,000 | 800,000 | 750,000 | 750,000 |
|  | Maximum Sub-yearling Releases | 300,000 | 1 | 300,000 | 1 | 1 |
|  | Broodstock Required | 616 | 140 | 616 | 421 | 421 |
|  | pNOB [Lo] Trigger (NOR run) | 2,000 | 1,100 | 2,000 | 2,000 | 3,000 |
|  | pNOB above Trigger | 100\% | 50\% | 100\% | 100\% | 100\% |
|  | pNOB below Trigger | 30\% | 30\% | 30\% | 30\% | 100\% |
| Segregated <br> Program | Maximum Yearling Releases Maximum Sub-yearling Releases Backfill w/ HORs (Y, N) | 500,000 | - | 500,000 | 600,000 | 600,000 |
|  |  | 400,000 |  | 400,000 |  | - |
|  |  | N | N | N | N | N |
| Other Control Variables | Maximum Weir Efficiency | 5\% | 5\% | 5\% | 5\% | 5\% |
|  | Term. Harvest Rate Integrated HORs | 29\% | 29\% | 29\% | 29\% | 29\% |
|  | Term. Harvest Rate Segregated HORs | 90\% | 90\% | 90\% | 90\% | 90\% |
|  | pNOB Trigger Range (NOR run) | 1,000 | sets range for "sliding scale pNOB" --applied to all phases |  |  |  |

Modeled outcomes versus Biological Targets

*Median, minimum and maximum values from 2017 to 2041 based on a single model run.
Figure 38. Screen shot of the decision rules in the In-Season Implementation Tool for CJHP planning at the Annual Program Review.

## Data Gaps and Research Needs

In a partnership with USGS, WDFW and the ONA, the CJHP is working to identify data gaps and applied research needs within the Okanogan Basin that would better inform hatchery management, increase available data for resource management decision making, and benefit overall salmonid recovery in the greater Columbia River basin. If funded in the future, the tasks identified could directly inform CJHP and other natural resource managers and aid in the decision making process. Some of the data gaps and applied research needs that have been identified include:

1. Refined estimates (extent, fate, timing and location) of summer/fall Chinook using the mainstem Columbia River above Wells Dam for spawning (i.e. straying), rather than returning to their natal Okanogan River using radio or acoustic telemetry.
2. Extent, fate, timing and location of spawning Chinook in the Canadian portion of the Okanogan Basin.
3. Development and testing of a panel of microsatellites and/or single nucleotide polymorphisms (SNPs) for genotyping genetic stocks of Chinook salmon in the Okanogan Basin and upper-Columbia River, upstream of Wells dam, to identify and differentiate Okanogan summer- vs. fall- vs. spring-Chinook, as well as hatchery $\times$ hatchery, hatchery $\times$ wild, and wild $\times$ wild crosses of these various life-history types.
4. Utilization of advancements in thermal imaging/LiDAR or other remote sensing technologies combined with in-stream temperature loggers and ArcGIS/R Statistical Program (STARS \& FLoWs toolsets \& SSN package) to map current thermal refugia in the Okanogan basin and model potential changes resulting from climate change scenarios.
5. Development and/or adaptation of existing methods for better estimation of fine sediment loads per reach length in the Okanogan River to quantify effects on Chinook salmon spawning redds and productivity.
6. Design for testing fish tagging rate assumptions. PIT, radio and genetic tagging emphasis.
7. Post-release mortality for various capture techniques including the purse seine, hatchery ladder, sport fishing, the weir, etc.
8. Abundance of Priest Rapids Hatchery fish at the Okanogan weir and CJH ladder.
9. Use of otolith microchemistry to determine origin and rearing locations of subyearling Chinook captured at various beach seining locations.

## References

Baldwin, C. M., A.M. Pearl, M. B. Laramie, J. P. Rohrback, P. E. Phillips and K. S. Wolf 2016. The Chief Joseph Hatchery Program 2013 Annual Report. BPA Project No. 2003-023-00, 148 pages.

Pearl, A.M., M. B. Laramie, C.M. Baldwin, J. P. Rohrback, P. E. Phillips 2016. The Chief Joseph Hatchery Program 2014 Annual Report. BPA Project No. 2003-023-00, 188 pages.

Busack, C. A., and K.P. Currens. 1995. Genetic Risks and Hazards in Hatchery Operations: Fundamental Concepts and Issues. American Fisheries Society Symposium 15: 7180.

CPUD (Chelan County Public Utility District). 2002a. Anadromous fish agreement and habitat conservation plan: Rock Island Hydroelectric Project, FERC License No. 943. March 26, 2002. Retrieved from http://www.chelanpud.org/documents/RI_HCP.pdf.

CPUD (Chelan County Public Utility District). 2002b. Anadromous fish agreement and habitat conservation plan: Rocky Reach Hydroelectric Project, FERC License No. 2145. March 26, 2002. Retrieved from http://www.chelanpud.org/documents/RR_HCP.pdf.

CJHP (Chief Joseph Hatchery Program). 2009. Monitoring and evaluation plan for summer/fall Chinook salmon. Retrieved from http://www.colvilletribes.com/cjhp_summer_chinook_m_e_plan_11_12_09.php.

CJHP (Chief Joseph Hatchery Program). 2012. Chief Joseph Hatchery Implementation Plan 2013. Prepared by Wolf, K.S. \& Pearl, A. M. Colville Confederated Tribes Fish \& Wildlife. Nespelem, WA.

CCT (Colville Confederated Tribes). 2008a. Hatchery and genetic management plan for the Chief Joseph Hatchery Program, Okanogan Summer/Fall Chinook. Submitted to the National Marine Fisheries Service 1 July 2008. Colville Confederated Tribes, Nespelem, Washington.

CCT (Colville Confederated Tribes). 2008b. Hatchery and genetic management plan for the Chief Joseph Hatchery Program, Okanogan Basin Spring Chinook. Submitted to the National Marine Fisheries Service 1 July 2008. Colville Confederated Tribes, Nespelem, Washington.

CCT (Colville Confederated Tribes). 2013. Hatchery and Genetic Management Plan for the Chief Joseph Hatchery Program, Okanogan Basin Spring Chinook, Non-essential Experimental Population. Submitted to the National Marine Fisheries Service 21 December 2012, updated 13 May 2013. Colville Confederated Tribes, Nespelem, Washington.

Crawford, B. A. and S. Rumsey. (2009). Guidance for monitoring recovery of salmon and steelhead listed under the Federal Endangered Species Act (Idaho, Oregon and Washington). (Draft). National Marine Fisheries Service- Northwest Region.

DPUD (Douglas County Public Utility District). 2002. Habitat Conservation Plans, Section 10 ESA Permits. Retrieved from http://www.westcoast.fisheries.noaa.gov/habitat/conservation_plans/rocky_reach_ wells_and_rock_island_mid-columbia_hcp.html

DPUD (Douglas County Public Utility District). 2014 Draft. Wells project sub-yearling Chinook life-history study, 2011-2013 report. Wells Hydroelectric Project. FERC No. 2149. East Wenatchee, Washington.

Ford, M. J. 2002. Selection in captivity during supportive breeding may reduce fitness in the wild. Conservation Biology, 16(3), 815-825.

Gallagher, S. P., P. K. J. Hahn, and D. H. Johnson. 2007. Redd counts. Pages 197-234 in D. H. Johnson, B. M. Shier, J. S. O'Neal, J. A. Knutzen, X. Augerot, T. A. O'Neal, and T. N. Pearsons. Salmonid field protocols handbook; techniques for assessing status and trends in salmon and trout populations. American Fisheries Society, Bethesda, Maryland.

HSRG (Hatchery Science Review Group). 2009. Report to Congress on Columbia River basin hatchery reform. February, 2009. Retrieved from http://www.hatcheryreform.us/hrp downloads/reports/columbia river/report to congress/hsrg report 12.pdf.

HSRG (Hatchery Scientific Review Group). 2014. On the Science of Hatcheries: An updated perspective on the role of hatcheries in salmon and steelhead management in the Pacific Northwest. A. Appleby, H.L. Blankenship, D. Campton, K. Currens, T. Evelyn, D. Fast, T. Flagg, J. Gislason, P. Kline, C. Mahnken, B. Missildine, L. Mobrand, G. Nandor, P. Paquet, S. Patterson, L. Seeb, S. Smith, and K. Warheit. June 2014; revised October 2014. Retrieved from http://hatcheryreform.us.

Hillman, T., M. Miller, C. Moran, M. Tonseth, M. Hughes, A. Murdoch, L. Keller, C. Willard, B. Ishida, C. Kamphaus, T. Pearsons, and P. Graf. 2014. Monitoring and evaluation of the Chelan and Grant County PUDs hatchery programs: 2013 annual report. Report to the HCP and PRCC Hatchery Committees, Wenatchee, WA. http://www.bioanalysts.net/FileShares/Uploaded\ Files/2013\ Annual\ R eport\%20with\%20Appendices.pdf

Hillman, T., M. Miller, M. Johnson, C. Moran, M. Tonseth, A. Murdoch, C. Willard, L. Keller, B. Ishida, C. Kamphaus, T. Pearsons, and P. Graf. 2015. Monitoring and evaluation of the Chelan and Grant County PUDs hatchery programs: 2014 annual report. Report to the HCP and PRCC Hatchery Committees, Wenatchee and Ephrata, WA. 1 April 2015 Draft version.

Laramie, M.B., Pilliod, D.S., and C.S.Goldberg. 2015b. Characterizing the distribution of an endangered salmonid using environmental DNA analysis. Biol. Conserv. 183: 29-37. doi: 10.1016/j.biocon.2014.11.025.

Laramie, M.B., Pilliod, D.S., Goldberg, C.S., and K.M. Strickler., 2015b, eDNA Sampling Protocol - Filtering Water to Capture DNA from Aquatic Organisms: U.S. Geological Survey, p. 15, http://dx.doi.org/10.3133/tm2A13.

MacDonald, M. 1894. Investigations in the Columbia River Basin in regard to the salmon fisheries. Report of the Commissioner of Fish and Fisheries. Government Printing Office. Washington D.C.

Mann R., and C. Snow. 2013. Population structure, movement patterns, and prespawn mortality for natural origin summer/fall Chinook salmon above Wells Dam. Report prepared for NOAA Fisheries and PSMFC, Award Number: NA10NMF4360439, PSMFC Job Number: 936A.11.10. Twisp, Washington. 83 pages.
Matthews, G. M., and R. S. Waples. 1991. Status review for Snake River spring and summer Chinook salmon. NOAA Technical Memorandum NMFS-F/NWC-200.

McClure, M. M., F. M. Utter, C. Baldwin, R. W. Carmichael, P. F. Hassemer,P. J. Howell, P. Spruell, T. D. Cooney, H. A. Schaller, and C. E. Petrosky. 2008. Evolutionary effects of alternative artificial propagation programs: implications for viability of endangered anadromous salmonids. Evolutionary Applications 1: 356-375.
McElhany, P., M.H. Ruckelshaus, M.J. Ford, T.C. Wainwright, and E.P. Bjorkstedt. 2000. Viable salmonid populations and the recovery of evolutionarily significant units. U.S. Dept. of Commerce. NOAA Tech. Memo. NMFS-NWFSC-42,156 p.

Miller, B.F., J.L. Miller, S.T. Schaller, and J.E. Arterburn. 2013. Okanogan Basin Monitoring and Evaluation Program, 2012 Annual Report. Colville Confederated Tribes Fish and Wildlife Department, Nespelem, WA. Project No. 2003-022-00.

Miller, B.F., J.D. Enns, J.L. Miller, S.T. Schaller, D.T. Hathaway, L. George and J.E. Arterburn. 2015. Okanogan Basin Monitoring and Evaluation Program, 2014 Annual Report. Colville Confederated Tribes Fish and Wildlife Department, Nespelem, WA. Project No. 2003-022-00.

Mobrand, L., and 9 coauthors. 2005. Hatchery reform in Washington State: principles and emerging issues. Fisheries 30: 11-23.

Murdoch, A. R., T.L. Miller, B.L. Truscott, C. Snow, C. Frady, K. Ryding, J.E. Arterburn, and D. Hathaway. 2012. Upper Columbia spring Chinook salmon and steelhead juvenile and adult abundance, productivity, and spatial structure monitoring. BPA Project Number 2010-034-00. Bonneville Power Administration, Portland, OR.

Murdoch, A. R., T. N. Pearsons, and T. W. Maitland. 2010. Estimating the spawning escapement of hatchery- and natural-origin spring Chinook salmon using redd and carcass data. North American Journal of Fisheries Management. 30:361-375. DOI: 10.1577/M09-071.1.

NMFS (National Marine Fisheries Service). 2008. Biological Opinion on National Marine Fisheries Service Endangered Species Act (ESA) Section 7 Consultation and MagnusonStevens Act (MSA) Essential Fish Habitat Consultation. Construction and Operation of Chief Joseph Hatchery by the Confederated Tribes of the Colville Reservation. The Salmon Recovery Division, Northwest Region, NMFS Consultation Number F/NWR/2006/07534.

National Oceanic and Atmospheric Administration (NOAA). (1994). Annual climatological summary. U.S. Department of Commerce National Oceanic and Atmospheric Administration. Asheville, NC: 1969-1994.

Paquet, P. J. and 15 co-authors. 2011. Hatcheries, conservation, and sustainable fisheriesachieving multiple goals: Results of the Hatchery Scientific Review Group's Columbia River Basin review. Fisheries 36:547-561.

Peven, C. and 18 co-authors. 2010. Proceedings and findings of the summer/Fall Chinook salmon summits. 2009 Status Report. Wenatchee, WA.

PNAMP (Pacific Northwest Aquatic Monitoring Partnership). 2012. Guidance for Implementing Successful Data Management \& Sharing. PNAMP Publication No. 2012004. http://www.pnamp.org/sites/default/files/2012-03-06pnamp-roadmap-pubno2012-004.pdf
Rayton, M.D., and P. Wagner 2006. KWA Ecological Sciences. Field Manual: Okanogan Basin Monitoring and Evaluation Program Rotary Screw Trap Protocols. Internal report prepared for the Colville Tribes adapted from trapping protocols by Murdoch et al. (2001) and Seiler and Volkhardt (2005). Retrieved from http://nrd.colvilletribes.com/obmep/Reports.htm.

Reisenbichler, R.R., and J.D. McIntyre. (1977). Genetic differences in growth and survival of juvenile hatchery and wild steelhead trout, Salmo gairdneri. Journal of the Fisheries Research Board of Canada, 34: 123-128.

Ryding, K.E., 2002. Appendix A - Review: 2000 Skagit River wild 0+ Chinook production evaluation in D. Seiler, S. Neuhauser, and L. Kishimoto 2002. 2001 Skagit River wild 0+ Chinook production evaluation, FPA 02-11. Washington Department of Fish and Wildlife, Olympia, WA.

Schmit, R.M., Charles O. Hamstreet, and Rebecca Christopherson. 2014. Chinook salmon spawning ground surveys on the Entiat River, 2013. U. S. Fish and Wildlife Service, Leavenworth Washington.
http://www.fws.gov/midcolumbiariverfro/pdf/2011\ Entiat\ River\ SCS \%20and\%20SUS\%20Chinook\%20SGS.pdf

Scofield, C., and J. Griffith. 2014. 2013 annual report. Stillaguamish River smolt trapping project. Stillaguamish Tribe of Indians Natural Resource Department. http://www.stillaguamish.com/images/Natural_Resources/Mainstem_Stillaguamis h_Smolt_Trap_Annual_Report_2013.pdf

Seber, G.A.F. 1982. The estimation of animal abundance and related parameters ( $2^{\text {nd }}$ ed.). London: Griffin.

USGS (U.S. Geological Survey). 2005. Water resources data - Washington water year 2005. Retrieved from http://pubs.usgs.gov/wdr/2005/wdr-wa-05-1/.

UCSRB (Upper Columbia Salmon Recovery Board) 2007. Upper Columbia spring Chinook salmon and steelhead recovery plan. 307 pages. Seattle, Washington. http://www.nwr.noaa.gov/Salmon-Recovery-Planning/Recovery-Domains/Interior-Columbia/Upper-Columbia/upload/UC_Plan.pdf.

Williamson, K.S., A.R. Murdoch, T.N. Pearsons, E.J. Ward, and M.J. Ford. Factors influencing the relative fitness of hatchery and wild spring Chinook salmon (Oncorhynchus tshawytscha) in the Wenatchee River, Washington, USA. (2010). Canadian Journal of Fisheries and Aquatic Sciences. 67:1840-1851.

Zhou, S. 2002. Size-dependent recovery of Chinook salmon in carcass surveys. Transactions of the American Fisheries Society 131:1194-1202.

## APPENDIX A

## Hatchery operations and production

The CJH's central facility is a 15 acre facility located immediately below Chief Joseph Dam along the right bank of the Columbia River at rkm. 872 near Bridgeport, WA. There are two CJH acclimation facilities on the Okanogan River, Omak (rkm 51) and Riverside (rkm 64) acclimation ponds. There is an additional acclimation facility on the Similkameen River (rkm 6.4) that is part of the CJH program but is operated by WDFW and funded by the CPUD.

Construction of the hatchery was completed in 2013 and broodstock were brought on station for the first time. The goal of the CJHP is to contribute to the increased abundance, productivity, temporal-spatial diversity, re-colonization of Chinook in the Okanogan Basin, and provide increased harvest for all fishers.

## Production Objectives

Full program production totals 2.9 million Chinook Salmon, including 2 million summer/fall Chinook and 900,000 spring Chinook. The summer/fall Chinook program incorporates both an integrated program ( 1.1 million smolts) supported by Okanogan River natural-origin broodstock and a segregated program ( 900,000 smolts) supported by hatchery-origin adults returning from the integrated program. The spring Chinook program includes a segregated program ( 700,000 smolts) supported by Leavenworth National Fish Hatchery (LNFH) broodstock and a re-introduction program (200,000 smolts) supported by WNFH broodstock (Met Comp stock) to reintroduce spring Chinook to the Okanogan under section $10(\mathrm{j})$ of the ESA.

In 2016, the summer/fall Chinook program production level did not meet full production as planned, due to higher than expected pre-spawn mortality on both the integrated and segregated summer/fall brood. The segregated spring Chinook program also did not meet full production goals due to an infrastructural automatic power transfer switch failure resulting in egg loss, which has since been corrected. The 10(j) spring Chinook reintroduction program was full program..

## Spring Chinook Salmon

## BY 2015 LEAVENWORTH Spring Chinook Rearing and ReLEase

Pre-spawn mortality was average and BKD prevalence was low, resulting in the program meeting its goal for fry ponded. A total of 784,165 fish were ad-clipped, with a total of 200,000 also receiving a CWT. This group also received 5,000 PIT tags, with a total
of 4,815 released ( 4,283 detected at release). During the month of April, reservoir water temperatures increased steadily, triggering a good smolt response. Feeding rates were increased for final grow out. A forced, nocturnal release strategy was implemented beginning on April 17 th with the last of the fish being pushed out April $18^{\text {th }}$.

## Cumulative egg to smolt survival

The cumulative egg to smolt survival for the 2015 brood Leavenworth-stock Spring Chinook was $96.7 \%$ (Table A 1). This includes ponding loss, rearing loss, and subtracting the shortage realized at marking. This overall survival metric will be a critical assessment of the hatchery's performance each brood year. The target egg to smolt survival identified in the original spring Chinook HGMP was 77\% (CCT 2008a).

Table A 1. Chief Joseph Hatchery BY 2015 Spring Chinook rearing summary, April 2017.

| Month | Total on hand | Mortality | Feed Fed | Fish per pound | Cumulative Survival (\%) |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | Total | Total | Total | Total | Total |
| Feb. 29 | 767,279 | 13,762 | 497 | 1,300 | 98.2\% |
| Mar. 31 | 763,160 | 4,119 | 1,874 | 750 | 97.7\% |
| Apr 30 | 761,868 | 1,292 | 1,561 | 150 | 97.5\% |
| May 31 | 761,239 | 629 | 1,920 | 100 | 97.4\% |
| June 30 | 784,563 | 372 | 1,311 | 70 | 97.4\% |
| July 31 | 747,086** | 1,597 | 2,693 | 50 | 97.1\% |
| Aug. 31 | 746,780 | 306 | 2,024 | 40 | 97.1\% |
| Sept. 30 | 746,419 | 361 | 3,666 | 35 | 97.0\% |
| Oct. 31 | 746,066 | 353 | 5,500 | 20 | 97.0\% |
| Nov. 30 | 745,362 | 704 | 2,464 | 20 | 96.9\% |
| Dec. 31 | 744,871 | 491 | 2,904 | 18 | 96.9\% |
| Jan. 31 | 744,560 | 311 | 2,508 | 16 | 96.8\% |
| Feb. 28 | 744,324 | 236 | 4,840 | 16 | 96.8\% |
| Mar. 31 | 744,199 | 125 | 2,816 | 15 | 96.8\% |
| Apr. 18 | 743,996* | 203 | 3,872 | 15 | 96.7\% |
| Cumulative | 743,996 | 24,861 | 40,450 | 15 | 96.7\% |

*Planted **Shortage of 35,880 in remaining raceways - total shortage overall of 12,184

## BY 2015 10j Met Comp Spring Chinook rearing and release

On October 26 ${ }^{\text {th }}, 2015$, CCT staff transported 209,956 MetComp Spring Chinook eyed eggs from the USFWS Winthrop National Fish Hatchery for rearing at CJH. On October

25th, 2015 fish were transferred to the Riverside Acclimation Pond. Under Permit No. 18928, issued by the National Marine Fisheries Service, this group is designated as an (10j) experimental population, for the reintroduction of Spring Chinook into the Okanogan Basin.

Temperatures at both Omak and Riverside dropped dramatically during December, and both ponds iced over. Over the course of the spring, temperatures rose steadily, and the fish growth stayed on target for release. These fish were forced released on April 15 ${ }^{\text {th }}$, 2017. Table A 2 illustrates feed fed, feeding rate, and mortality to date. After subtracting mortality and shed tags, a total of 5,036 PIT tags were released ( 661 were detected at release.)

Table A 2. Riverside Acclimation Pond BY 2015 integrated Spring Chinook rearing summary, April 2017.
${ }^{\text {a }}$ Released

| Month | Total on hand | Rearing Mortality <br> Total | Feed Fed <br> Total | Fish per pound <br> Total | Cumulative Survival (\%) <br> Total |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | Total |  |  |  |  |
| Nov 30 | 202,859 | 141 | 1,980 | 30 | 96.6\% |
| Dec. 31 | 202,326 | 533 | 132 | 24 | 96.5\% |
| Jan. 31 | 202,225 | 101 | 0 | 24 | 96.5\% |
| Feb. 28 | 201,952 | 273 | 88 | 25 | 96.4\% |
| Mar. 31 | 201,836 | 116 | 1,496 | 20 | 96.4\% |
| Apr. 15 | 201,821 ${ }^{\text {a }}$ | 15 | 572 | 15.9 | 96.4\% |
| Cumulative | 201,821 | 5,662 ${ }^{\text {b }}$ | 10,234 | 15.9 | 96.4\% |

## BY 2016 LEAVENWORTH SPRING CHINOOK

## 2016 Brood Collection

The segregated spring Chinook production goal for the 2016 brood is a release of 700,000 yearlings in April of 2018. The calculated number of brood needed to meet this production was 640 adults, based on a 50/50 ratio of males and females. This includes $10 \%$ pre-spawn mortality, up to $20 \%$ culling for Bacterial Kidney Disease (BKD) management, $10 \%$ egg loss, and rearing mortality of $15 \%$. The mortality per life stage benchmarks were based on historical performance at LNFH. As with any new facility, baseline data collected
during initial production years will be the basis for adjusting broodstock requirements in future years.

A total of 639 spring Chinook broodstock were transferred from LNFH to CJH June $1^{\text {st }}$ through $2^{\text {nd }}, 2016$; including 319 females, 320 males (Table A 3). The 639 spring Chinook transferred represents $99.8 \%$ of the collection objective.

Transport water was obtained from LNFH to fill the transport trucks, at a temperature of $51^{\circ} \mathrm{F}$ and the adult holding pond temperature, at LNFH, was $48^{\circ} \mathrm{F}$. Transport densities on both days were 0.60 lbs . /gal. (Table A 3). All transport included Vita Life, a calming agent superior to salt, at a rate of 500 ml per 2,000 gal., and supplemental oxygen at $8 \mathrm{~L} / \mathrm{min}$. There were no mortalities associated with the transport.

Broodstock were off loaded, via water-to-water transfer, into adult ponds \#5 and \#6, at CJH. The receiving water was $55^{\circ} \mathrm{F}$. The adult pond had a flow rate of 380 gpm . and an exchange rate of 60 minutes, representing a Flow Index (FI) of 0.42 and 0.20 for pond \#5 and \#6, respectively (Table A 4). The Density Index (DI) was 0.04 and 0.02 for ponds \#5 and \#6, respectively. Both adult ponds were $100 \%$ well water. Both ponds were treated two days per week with hydrogen peroxide at 1:10,000 for a copepod infection on the gills, three days per week with formalin at 1:6,000 to control fungus. All treatments were performed for one exchange.

On July 6 th, 2016 and again on August 3rd, 2016 USFWS DVM Trista Welsh-Becker assisted hatchery staff with inoculations for all spring Chinook brood. Each female was inoculated with Draxxin at a rate of $10 \mathrm{mg} / \mathrm{kg}$ for reduction of BKD, and Vetrimycin - 200 (Oxytetracycline) IP at $20 \mathrm{mg} / \mathrm{kg}$; for reduction of pre-spawn mortality due to furunculosis. Pre-spawn mortality was very low at $99.1 \%$ survival. (Table A 5)

Table A 3. Chief Joseph Hatchery spring Chinook broodstock transfer summary for 2016.

| Date | Trap site | Receiving Facility | Males |  |  | Females | Total Brood | Holding <br> Temp $\left({ }^{\circ} \mathbf{F}\right)$ | Transport Temp.$\left({ }^{\circ} \mathbf{F}\right)$ | Transport Density (lbs./gal) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Adult | Jack | Total |  |  |  |  |  |
| 6/1/2016 | LNFH | CJH | 156 | 0 | 156 | 165 | 321 | 48 | 51 | 0.48 |
| 6/2/2016 | LNFH | CJH | 164 | 0 | 164 | 154 | 318 | 48 | 51 | 0.48 |
| Total |  |  | 320 | 0 | 320 | 319 | 639 |  |  |  |

Table A 4. Spring Chinook broodstock adult holding conditions for 2016.

| Transfer Date | Adult <br> Pond | Males |  |  | Females | Total Brood | Transport <br> Temp ( ${ }^{\circ}$ F) | Holding Temp. ( ${ }^{\circ}$ F) | Flow <br> Index | Density Index |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Adult | Jack | Total |  |  |  |  |  |  |
| 6/1/2016 | \#5 | 156 | 0 | 156 | 165 | 321 | 51 | 56 | 0.26 | 0.05 |
| 6/2/2016 | \#6 | 164 | 0 | 164 | 154 | 318 | 51 | 56 | 0.26 | 0.05 |

Table A 5. Chief Joseph Hatchery spring Chinook broodstock holding and survival summary for 2016. ( $\mathrm{M}=$ adult males, $\mathrm{J}=$ jacks, and $\mathrm{F}=$ adult females). The survival standard for this life stage was $90 \%$.

| Beginning |  |  | Ending |  |  | Mortality |  |  | Cumulative Survival (\%) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| M | J | F | M | J | F | M | J | F | M | J | F |
| 320 | 0 | 319 | 316 | 0 | 317 | 4 | 0 | 2 | 98.8\% | NA | 99.4\% |

## Spawning

Spring Chinook spawning started August 17 ${ }^{\text {th }}$ and concluded on August 31 ${ }^{\text {st }}$ (Error! eference source not found.). The spawn consisted of 279 females and 279 males, with nine non-viable (green) females killed resulting in a green egg take of approximately 1,060,200 (Error! Reference source not found.).

Spawning occurred inside the spawning shed adjacent to the adult holding raceways, and gametes were then transported to the main facilities egg entry room for processing. Each individually numbered female was fertilized with a primary male initially, and then a backup male was added to ensure fertilization. Each female's eggs were then placed in the corresponding numbered trey. The eggs from one female was culled due to high or moderate ELISA results (culled eggs from ELISA results are not included in Error! eference source not found.). This was approximately $0.4 \%$ of the females spawned and was far less than planned for (up to 20\%).

## Incubation

Each female's eggs were initially incubated separately to facilitate culling based on ELISA results. Once eyed, egg mortality was removed and eggs were combined for hatching. All spring Chinook eggs were placed on varying degrees of chilled water. The water temperature was gradually dropped, on the first egg take, to $37^{\circ} \mathrm{F}$ degrees. This process was done over a several hour period four days after spawning. The second egg take was left on well water ( $60^{\circ} \mathrm{F}$ ) until such time as the total numbers of temperature units
(TUs) were earned to equal the first egg take, then the same procedure was used to lower water temperature to $37^{\circ} \mathrm{F}$. This process provided the ability to control when, and how many, fish are brought out of the incubators and placed into early rearing.

Green egg to eyed egg survival was 51.1\% (Error! Reference source not found.). his survival was far lower than the key assumption (90\%) due to a flaw in power monitoring switches that provide electricity to the well pumps, resulting in mortality associated with oxygen deprivation.

## Rearing

Due to the manipulation of TUs, the first group of spring Chinook was brought out of incubation and transferred into early rearing troughs on April 28 ${ }^{\text {th }}, 2017$, and the second group on May $12^{\text {th }}, 2017$. During the month, the first group was introduced to feed in the early rearing troughs, and reared for a period of two weeks. After the initial rearing period inside, this group was transferred outside to the standard raceways via the fry transfer line. The second group, however, were transferred immediately after ponding to outdoor raceways via transfer line, due to a major water pipeline crisis two hours after ponding from incubator into early-rearing trough. Although a higher than normal level of early rearing mortality did occur due the stress from this premature transfer, $<1.6 \%$ of population was lost (as opposed to $0.5 \%$ for the first ponding group). No inventory was taken at ponding to prevent excess handling stress. Survival from incubation to ponding was $94.5 \%$ which was on par with the standard (95\%) for this life stage (Error! Reference ource not found.).

Table A 6. Chief Joseph Hatchery spring Chinook spawning and survival summary for 2016 ( $\mathrm{M}=$ adult males, $\mathrm{J}=$ jacks and $\mathrm{F}=$ adult females). The target survival standard for this life stage was $90 \%$.

| Spawn Date | Total Adults Spawned |  |  | Green Egg Take <br> Total | Eyed Egg <br> Total | Mortality (Pick off) <br> Total | Cumulative <br> Survival (\%) <br> Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | M | J | F |  |  |  |  |
| August 17 | 94 | 0 | 89 | 338,200 | 310,332 | 47,210 | 86.80\% |
| August 24 | 136 | 0 | 134 | 509,200 | 287,879 | 210,162 | 57.80\% |
| August 31 | 57 | 0 | 56 | 212,800 | 17,320 | 185,067 | 8.56\% |
| Total | 287 | 0 | 279 | 1,060,200 | 615,532 | 442,439 | 51.05\% |

Table A 7. Chief Joseph Hatchery spring Chinook ponding summary for BY 2016 as of May 2017.

|  | $\underline{\text { Total Fry }}$ | $\underline{\text { Ponding }}$ | $\underline{\text { Monthly }}$ | $\underline{\text { Monthly }}$ | Ponding | Cumulative <br> Survival (\%) |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Production <br> Group | Total | $\underline{\text { Mortality }}$ | $\underline{\text { Teed }}$ | $\underline{\text { Mortality }}$ | $\underline{\text { Loss (\%) }}$ | Total |

The key assumption survival for this life stage is $95 \%$.

Spring Chinook were fed Bio-Oregon's BioPro 2 diet, and converted at an average of 0.88:1 for their first six weeks of rearing. Post ponding rearing is on schedule, with no fish health issues and minimal mortality to date (Table A 8). Survival for this life stage will be reported in subsequent annual reports once all release information is available.

Table A 8. Chief Joseph Hatchery BY 2016 segregated spring Chinook rearing summary as of May 2017.

| Month | Total on hand | Mortality | Feed Fed | Fish per pound | $\frac{\text { Cumulative }}{\frac{\text { Survival }}{\text { (\%) }}}$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | Total | Total | Tot <br> al | Total | Total |
| LVN |  |  |  |  |  |
| May 31 | 581,528 | 8,317 | 486 | 620 | 98.6\% |

## Summer/Fall Chinook Salmon

BY 2015 Summer/FALL Chinook SaLmon Rearing and ReLease
A total of 423,658 sub-yearling summer/fall Chinook were brought out of incubation from December 23 rd, 2015 through January 21st, 2016. An addition 603,421 yearling summer/fall Chinook were also brought out of incubation from March 31st, 2016 through April 26th, 2016.

Rearing proceeded on schedule, with the marking and releasing of both the integrated and segregated sub-yearlings in April. On May $3^{\text {rd }}$, a total of 175,872 integrated sub-yearlings were transferred to the Omak Acclimation Pond, at 58 fpp . This group was released on May 23rd, 2016, with a post transfer survival of $88.0 \%$, and a cumulative survival from ponding of $87.9 \%$ (Table A 9). Approximately 5,000 PIT tags were added to each group and after subtracting shed tags and mortality, a total of 4,979 PIT tags were released ( 1,237 were detected at release).

Table A 9. Chief Joseph Hatchery brood year 2015 integrated summer/fall Chinook subyearling rearing summary.

|  | Total Planted | Mortality | Feed Fed | Fish per pound | $\frac{\text { Cumulative }}{\text { Survival }}$ <br> $(\%)$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Month | Total | Total | Total | Total | Total |
| May 23 | 175,771 | 22,040 | 2,650 | 44 | 87.9 |
| Cumulative | 175,771 | 22,040 | 2,650 | 44 | 87.9 |

A total of 218,459 segregated summer/fall sub-yearlings were marked and transferred into rearing Pond B, for final rearing and release. This group was released on May $16^{\text {th }}, 2016$, at 56.7 fpp . Cumulative rearing survival was $90.3 \%$ (Table A 10). Approximately 5,000 PIT tags were added to each group and after subtracting shed tags and mortality, a total of 3,936 PIT tags were detected at release

Table A 10. Chief Joseph Hatchery brood year 2015 segregated summer/fall Chinook subyearling rearing summary.

|  | Total Planted | Mortality | Feed Fed | Fish per pound | $\begin{gathered} \frac{\text { Cumulative }}{\text { Survival }} \\ \text { (\%) } \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Month | Total | Total | Total | Total | Total |
| May 16 | 218,383 | 24,386 | 2,951 | 56.7 | 90.3 |
| $\begin{gathered} \text { Cumulativ } \\ \mathrm{e} \\ \hline \end{gathered}$ | 218,383 | 24,386 | 2,951 | 56.7 | 90.3 |

The yearling summer/fall Chinook rearing proceeded on schedule, with both the integrated and segregated groups being marked in July and August. Marking was completed, for both the integrated and the segregated programs, on August 4th, 2016. The segregated Summer Chinook were 100\% ad-clipped, with a 100k CWT group tagged. The integrated Summer Chinook were 100\% AD/CWT. As shown in Error! Reference source ot found., ponding and rearing mortality for the segregated program has been lower than anticipated, although both stocks were short of book numbers, at marking. The segregated production was marked into rearing Pond B, while the integrated program was marked into the lower raceways, and reared until transfer to the acclimation ponds in late October. Both groups were released the evening of April 18 ${ }^{\text {th }}, 2017$. Approximately 5,000 PIT tags were added to each group in October 2016. After subtracting shed tags and mortality, a total of 5,024 PIT tags were released from the segregated group (4,194 were detected at release). Final feed conversion from rearing stage to release was 1.09:1.

The integrated Summer Chinook were shipped to the Omak Acclimation Pond, and the Similkameen Acclimation Pond, on October 24th, 2016. Reporting for the Similkameen Pond will reside with WDFW through release. A nocturnal forced release took place on April 20 th, 2017 at the Omak Pond.

Table A 11. Chief Joseph Hatchery BY 2015 segregated Summer/Fall Chinook rearing summary.
*Overage at marking - 20,473

| HORs <br> Month | Total on hand | Mortality | Feed Fed | Fish per pound | Cumulative <br> Survival (\%) |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | Total | Total | Total | Total | Total |
| Apr. 30 | 216,291 | 1,333 | 120 | 650 | 99.4\% |
| May 31 | 214,660 | 1,631 | 495 | 270 | 98.6\% |
| June 30 | 213,815 | 845 | 673 | 150 | 98.2\% |
| July 31 | 233,985* | 303 | 1,239 | 100 | 98.2\% |
| Aug. 31 | 233,718 | 267 | 1,056 | 60 | 98.1\% |
| Sept. 30 | 233,642 | 76 | 1,247 | 40 | 98.1\% |
| Oct. 31 | 233,472 | 170 | 1,936 | 30 | 98.0\% |
| Nov. 30 | 232,960 | 512 | 968 | 20 | 97.9\% |
| Dec. 31 | 232,638 | 322 | 1,320 | 20 | 97.9\% |
| Jan. 31 | 232,459 | 179 | 1,320 | 16 | 97.8\% |
| Feb. 28 | 232,379 | 80 | 1,936 | 15 | 97.8\% |
| Mar. 31 | 232,280 | 99 | 4,180 | 14 | 97.7\% |
| Apr. 18 | 232,103^ | 177 | 4,620 | 12 | 97.7\% |
| Sub Total | 232,103 | 5,994 | 21,110 | 12 | 97.7\% |

## Omak Acclimation Pond

On October 24th, 2016 Chief Joseph Hatchery staff transferred 213,830 Integrated BY 15 Summer Chinook from Chief Joseph Hatchery to the Omak Acclimation Pond. At the time of transfer, the fish were approximately 30 fpp , and were programmed to be reared over winter, with a target size at release of 10 fpp . An additional 167,291 BY 15 Summer Chinook were transferred to WDFW's Similkameen Pond, as part of the cost share agreement. These fish were forced released April 20 th, 2017. Table A 12 illustrates feed fed, feeding rate, and mortality to date for the integrated summer/fall Chinook transferred to the Omak Acclimation pond. After subtracting mortality and shed tags, a total of 4,830 PIT tags were released ( 1,578 were detected at release).

Table A 12. Omak Acclimation Pond BY 15 integrated summer/fall Chinook rearing summary.

| Month | Total on hand | Mortality | Feed Fed | Fish per pound | $\begin{aligned} & \text { Cumulative } \\ & \text { Survival (\%) } \\ & \hline \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | Total | Total | Total | Total | Total |
| Nov. 30 | 213,621 | 209 | 1,980 | 20 | 98.0\% |
| Dec. 31 | 213,543 | 78 | 176 | 20 | 98.0\% |
| Jan. 31 | 213,259 | 284 | 0 | 20 | 97.9\% |
| Feb. 28 | 212,898 | 361 | 88 | 21 | 97.8\% |
| Mar. 31 | 212,790 | 108 | 1,884 | 17 | 97.8\% |
| Apr. 20 | 212,726 | 64 | 1,279 | 13.8 | 97.7\% |
| Cumulative | 212,726 | 8,398 | $\begin{gathered} 15,61 \\ 2 \\ \hline \end{gathered}$ | 13.8 | 97.7\% |

## Riverside Acclimation Pond

Riverside Acclimation Pond was not used to rear BY 2015 summer/fall Chinook, but was utilized to rear BY 15 10j Spring Chinook, as noted above.

## Similkameen Acclimation Pond

Similkameen Pond was used to rear yearling summer Chinook per the WDFW program funded by CPUD. Adult broodstock used to generate the juveniles for BY 2015 were collected via the CCT purse seine as part of the transition to the collaborative CJH program. On October 24 ${ }^{\text {th }}, 2016$, Chief Joseph Hatchery staff transferred 167,291 summer/fall Chinook to the Similkameen Pond, with the assistance of WDFW's Eastbank Hatchery staff. At the time of transfer, the fish were approximately 30 fpp , and were programmed for over winter acclimation, with a target size at release of 10 fpp . These fish began volitional release on April 12 ${ }^{\text {th }}$, with an end release date of April 15 nd, 2017. Cumulative survival, at the date of transfer, was $98.1 \%$. Survival from transfer to release was 78.4\%.

## Cumulative egg to smolt survival

The target egg to smolt survival identified in the original summer/fall Chinook HGMP was $77.5 \%$ for sub-yearlings and $73.5 \%$ for yearlings (CCT 2008b). The cumulative egg to smolt survival, for the BY 2015 sub-yearlings, was $89.1 \%$. The cumulative egg to smolt survival, for the BY 2015 yearlings, was $97.7 \%$.

## 2016 Broodstock collection

Collection of summer/fall Chinook for BY 2016 occurred between July 5th, 2016 and August 28 ${ }^{\text {th }}, 2016$ via the CCT purse seine operation at the mouth of the Okanogan River. Both hatchery- and natural-origin brood were collected to supply the integrated and segregated production programs at CJH. As the seine was being pursed, 9-meter transport barges approached the seine vessel and tied off on the opposite side. The broodstock transport barges have two transport tanks, a 300 gallon for HORs and a 600 gallon for NORs. Brood fish were removed from the seine and placed headfirst in a rubber tube, or boot, containing some water and handed to the staff on the barges for placement in the holding tanks. A maximum of 14 HOR and 28 NOR brood could be loaded per barge. Once full, or at the commencement of the purse seine haul, the barges returned to the offload area at Mosquito Park approximately 2 km away. The brood was then removed from the tanks by hand, placed into a boot, then delivered to one of two 2,500 gallon tanker trucks and transported 16 km to the hatchery.

Water temperatures were of major concern during these operations and monitored to minimize trauma to the adult brood. Okanogan River temperatures during July ranged from $66^{\circ} \mathrm{F}\left(19^{\circ} \mathrm{C}\right)$ to $78^{\circ} \mathrm{F}\left(25.5^{\circ} \mathrm{C}\right)$. In order to limit the effects of the temperature changes we monitored the temperature of all transport vessels and strived to not expose brood to changes greater than $8^{\circ} \mathrm{F}$. We accomplish this by utilizing both well water and surface water when filling the barges and transport tankers, and monitoring our raceway temperatures.

A weekly quota was developed to ensure that brood collections occurred across as much of the summer run timing as possible (Table A 13). If brood collection failed to meet the weekly quota it was adjusted the following week. The purse seine is only effective when there is a thermal barrier at the mouth of the Okanogan, therefore broodstock can only be collected there until late August or early September. Broodstock were offloaded, via water-to-water transfer, into adult ponds at CJH. The receiving water was approximately $57^{\circ} \mathrm{F}$. The adult ponds had a flow rate of 380 gpm ., and an exchange rate of 60 minutes, representing a Flow Index (FI) of 0.15 and a Density Index (DI) of 0.02. Upon arrival, adult ponds were put on well water. All adult ponds were treated a minimum of four days per week with formalin to control fungus at a rate of 1:6000, for one exchange. Additionally, brood fish were treated two days per week with Chloramine-T at 12 ppm for one exchange to control Columnaris bacteria.

On August 10 ${ }^{\text {th }}, 2016$ USFWS DVM Trista Welsh-Becker assisted hatchery staff with inoculations for all summer/fall Chinook brood. Each female was inoculated with Draxxin
at a rate of $10 \mathrm{mg} / \mathrm{kg}$ IP, for reduction of BKD, and Vetrimycin - 200 (Oxytetracycline) IP, at the $20 \mathrm{mg} / \mathrm{kg}$ for reduction of pre-spawn mortality due to furunculosis.

A total of 552 HOB were collected including 281 females, 271 adult males and 0 jacks (Error! Reference source not found.). A total of 572 NOB was collected including 87 females, 285 adult males, and 0 jacks (Error! Reference source not found.). No steelhead or Bull trout were encountered during broodstock collection efforts.

The brood collection objective for the weir was 84 NOB. Fourteen NOB Chinook were collected from the weir trap between September 5 and September 14, 2016. The adults were transported to shore via a fish boot (rubber tire inner tube) and placed into an 800-gallon hatchery truck. The fish were then transported approximately 32 km to Chief Joseph Hatchery where they were held in the broodstock raceways until the first spawn date the first week in October. The broodstock shortfall ( $\mathrm{n}=72$ ) for late-arriving fish was filled by beach seining the Kline site on the Similkameen River. We recognize that fish collected late may have arrived at any point in their run timing; however, the efforts to collect fish in September at least offer the opportunity to include fish that arrive later in the run timing. These adults were $100 \%$ otolith sampled at spawning. The goal was to ensure that, prior to being included in the integrated production; there would be no unmarked Priest Rapids Hatchery fish in this group. Based on results from the WDFW Otolith Lab, there were no unmarked Priest Rapids Hatchery fish collected in September for the integrated program; therefore, it was not necessary to place any of these fish into the segregated program.

Table A 13. Chief Joseph Hatchery summer/fall Chinook weekly broodstock collection objectives and results for brood year 2016.

| Week | Weekly Quota |  | Cumulative Proportion | Cumulative Collection |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\frac{\text { Natural }}{\text { Origin }^{2}}$ | $\begin{aligned} & \text { Hatchery } \\ & \text { Origin }^{3} \end{aligned}$ |  | $\begin{aligned} & \text { Natural } \\ & \underline{\text { Origin }} \\ & \hline \end{aligned}$ | $\begin{aligned} & \text { Hatchery } \\ & \text { Origin } \end{aligned}$ |
| July 4 - July 10 | 22 | 22 | 0.04 | 22 | 22 |
| July 11 - July |  |  |  |  |  |
| 17 | 22 | 22 | 0.08 | 44 | 44 |
| July 18 - July 24 | 108 | 104 | 0.27 | 152 | 148 |
| July 25 - July |  |  |  |  |  |
| 31 | 108 | 104 | 0.46 | 260 | 252 |
| Aug. 1 - Aug. 7 | 132 | 126 | 0.69 | 392 | 378 |
| $\text { Aug. } 8 \text { - Aug. }$ $14$ | 132 | 126 | 0.92 | 524 | 504 |
| $\begin{aligned} & \text { Aug. } 15 \text { - Aug. } \\ & 21 \end{aligned}$ | 36 | 36 | 0.98 | 560 | 540 |
| $\begin{aligned} & \text { Aug. } 22 \text { - Aug. } \\ & 28 \end{aligned}$ | 12 | 12 | 1.00 | 572 | 552 |
| *Sept. 15 Oct. 15 *NOR weir collection | 84 |  |  | 656 |  |
| ${ }^{1}$ Weekly collection short-fall to be added to following week's collection |  |  |  |  |  |
| ${ }^{2}$ Combined collection strategies in priority order (purse seine, tangle-net, Okanogan weir beach seine and CJH ladder |  |  |  |  |  |
| ${ }^{3}$ Combined collection strategies in priority order: purse seine, tangle-net, CJH ladder, Okanogan weir and beach seine |  |  |  |  |  |

Table A 14. Chief Joseph Hatchery summer/fall Chinook Hatchery-Origin Broodstock (HOB) transfer summary for brood year 2016.

| Date | $\begin{gathered} \text { Trapping } \\ \text { site } \\ \hline \end{gathered}$ | Receiving Facility | Males | Females | Jacks | Total <br> Broodstock | River Temp (f0) | Barge Temp (F0) |  | Adult <br> Pond <br> Temp <br> (f0) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 7/5/2016 | SEINE | CJH | 1 | 0 | 0 | 1 | 72 | 62 | 58 | 57 |
| 7/6/2016 | SEINE | CJH | 10 | 11 | 0 | 22 | 72 | 62 | 58 | 57 |
| 7/11/2016 | SEINE | CJH | 1 | 3 | 0 | 26 | 72 | 62 | 58 | 57 |
| 7/12/2016 | SEINE | CJH | 7 | 10 | 0 | 43 | 74 | 66 | 60 | 57 |
| 7/18/2016 | SEINE | CJH | 17 | 11 | 0 | 71 | 76 | 64 | 59 | 57 |
| 7/19/2016 | SEINE | CJH | 39 | 35 | 0 | 145 | 76 | 64 | 59 | 57 |
| 7/25/2016 | SEINE | CJH | 12 | 14 | 0 | 171 | 76 | 64 | 59 | 57 |
| 7/26/2016 | SEINE | CJH | 2 | 4 | 0 | 177 | 76 | 64 | 59 | 57 |
| 7/27/2016 | SEINE | CJH | 6 | 4 | 0 | 187 | 78 | 64 | 59 | 57 |
| 7/28/2016 | SEINE | CJH | 31 | 34 | 0 | 242 | 78 | 64 | 59 | 57 |
| 7/29/2016 | SEINE | CJH | 9 | 8 | 0 | 269 | 80 | 64 | 59 | 57 |
| Total |  |  | 135 | 134 | 0 | 269 |  |  |  |  |

Table A 15. Chief Joseph Hatchery summer/fall Chinook Natural-Origin Broodstock (NOB) transfer summary for brood year 2016.

| Date | Trapping site | Receiving Facility | Males | Females | Jacks | Total Broodstock | River Temp (f0) | Barge Temp (F0) | Transport Temp. (F0) | Adult <br> Pond <br> Temp <br> (f0) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 7/5/2016 | SEINE | СJH | 0 | 3 | 0 | 3 | 74 | 64 | 59 | 57 |
| 7/6/2016 | SEINE | CJH | 9 | 13 | 0 | 25 | 74 | 64 | 59 | 57 |
| 7/11/2016 | SEINE | CJH | 8 | 6 | 0 | 39 | 76 | 64 | 59 | 57 |
| 7/12/2016 | SEINE | CJH | 1 | 7 | 0 | 47 | 76 | 64 | 59 | 57 |
| 7/18/2016 | SEINE | СЈН | 61 | 49 | 0 | 157 | 76 | 64 | 59 | 57 |
| 7/19/2016 | SEINE | CJH | 0 | 0 | 0 | 157 | 76 | 64 | 59 | 57 |
| 7/25/2016 | SEINE | CJH | 32 | 32 | 0 | 221 | 76 | 64 | 59 | 57 |
| 7/26/2016 | SEINE | CJH | 7 | 19 | 0 | 247 | 76 | 64 | 59 | 57 |
| 7/27/2016 | SEINE | CJH | 12 | 6 | 0 | 265 | 78 | 64 | 59 | 57 |
| 7/28/2016 | SEINE | СЈH | 0 | 0 | 0 | 265 | 78 | 64 | 59 | 57 |
| 7/29/2016 | SEINE | CJH | 15 | 25 | 0 | 305 | 80 | 64 | 56 | 56 |
| Total |  |  | 145 | 160 | 0 | 305 |  |  |  |  |

The cumulative pre spawn holding survival, for all Summer/Fall brood collected, was $95.9 \%$ for HOB and $93.4 \%$ for NOB (Table A 16). Both programs met the survival standard (90\%).
Table A 16. Chief Joseph Hatchery summer/fall Chinook Hatchery (HOB) and Natural (NOB) origin broodstock holding survival summary for brood year 2016. ( $\mathrm{M}=$ adult males, $\mathrm{J}=\mathrm{jacks}$ and $\mathrm{F}=$ adult females). The survival standard for this life stage was $90 \%$.

| Month | Beginning Month |  |  | Ending Month |  |  | Mortality |  |  | Monthly Survival$\qquad$ (\%) |  |  | Cumulative Survival$(\%)$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | M | J | F | M | J | F | M | J | F | M | J | F | M | J | F |
| HOR |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Aug | 271 | 0 | 281 | 269 | 0 | 275 | 2 | 0 | 6 | 99.2\% | 100\% | 97.8\% | 99.2\% | 100\% | 97.8\% |
| Sept | 269 | 0 | 275 | 269 | 0 | 276* | 0 | 0 | 5 | 100\% | 100\% | 98.3\% | 99.2\% | 100\% | 96.1\% |
| Oct | 269 | 0 | 276 | 265 | 0 | 270 | 4 | 0 | 6 | 98.5\% | 100\% | 97.8\% | 97.7\% | 100\% | 94.0\% |
| NOR |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Aug | 285 | 0 | 287 | 283 | 0 | 282 | 2 | 0 | 5 | 99.2\% | 100\% | 98.2\% | 99.2\% | 100\% | 98.2\% |
| Sept | 283 | 0 | 282 | $323 \wedge$ | 0 | $324 \wedge$ | 2 | 0 | 2 | 99.4\% | 100\% | 99.4\% | 98.8\% | 100\% | 97.9\% |
| Oct | 323 | 0 | 324 | 314 | 0 | 307 | 9 | 0 | 17 | 97.2\% | 100\% | 94.7\% | 96.0\% | 100\% | 92.7\% |
| Nov. | 314 | 0 | 307 | 314 | 0 | 298 | 0 | 0 | 9 | 100\% | 100\% | 97.0\% | 96.0\% | 100\% | 90.5\% |

* Weir Brood added ^Added 44 NOR Females and 42 NOR males


## Spawning

Spawning of Summer Chinook began on October 5th, 2016, and continued through November $2^{\text {nd }}, 2016$. As with the Spring Chinook, the Summer Chinook program is also $100 \%$ ELISA sampled. For the 2016 brood, we experienced a much lower than normal disease profile, and as a result no females were culled.

Total NOB spawned included 287 males, zero jacks, and 297 females. (Table A 17) None of the 16 NOB taken from the weir and spawned had an otolith mark; indicating that there were no ad-present Priest Rapids Hatchery adults included in the integrated brood. Total HOR spawn included 244 males, zero jacks, and 238 females. In addition, 13 nonviable NOR females and eight non-viable HOR females were spawned. Total eyed egg take for the season was $2,004,861$. Egg survival from green egg to eyed egg for NOB averaged 75.13\% (Table A 17). Egg survival for HOB averaged 81.25\%. Survival was lower than the key assumption of ( $90 \%$ ) for this life stage.

Table A 17. Chief Joseph Hatchery brood year 2016 summer/fall Chinook spawning results.

| Month | $\frac{\text { Total Adults }}{\text { Spawned }}$ |  |  | Green Egg Take <br> Total | Eyed Egg <br> Total | Mortality (Pick Off) <br> Tota <br> 1 | Cumulative Survival (\%) <br> Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | M | J | F |  |  |  |  |
| NOR |  |  |  |  |  |  |  |
| Oct. 5 | 20 | 0 | 19 | 95,000 | 67,670 | 9,323 | 87.9\% |
| Oct. 12 | 67 | 0 | 66 | 330,000 | 264,719 | 25,264 | 91.2\% |
| Oct. 19 | 101 | 0 | 99 | 495,000 | 388,042 | 45,276 | 89.6\% |
| Oct. 26 | 96 | 0 | 96 | 480,000 | 342,017 | 61,363 | 84.7\% |
| Nov 2 | 3 | 0 | 17 | 85,000 | 36,251 | 39,943 | 47.5\% |
|  |  |  | 29 | 1,485,00 |  | 181,16 |  |
| Sub-total | 287 | 0 | 7 | 0 | 1,098,699 | 9 | 85.8\% |
| HOR |  |  |  |  |  |  |  |
| Oct. 5 | 24 | 0 | 23 | 115,000 | 95,517 | 8,717 | 91.6\% |
| Oct. 12 | 97 |  | $\begin{aligned} & 94 \\ & 10 \end{aligned}$ | 470,000 | 386,930 | 26,858 | 92.6\% |
| Oct. 19 | 101 | 0 | 0 | 500,000 | 264,990 | 17,786 | 93.7\% |
| Oct. 26 | 22 | 0 | 21 | 105,000 | 158,725 | 96,753 | 62.1\% |
|  |  |  | 23 | 1,190,00 |  | 150,11 |  |
| Sub-total | 244 | 0 | 8 | 0 | 906,162 | 4 | 85.7\% |
|  |  |  | 51 | 2,675,00 |  |  |  |
| Total | 528 | 0 | 8 | 0 | 2,004,861 | 331,283 | 85.8\% |

## Broodstock Origin

Broodstock were interrogated for coded-wire tags on four different spawning events during October: $10 / 5,10 / 12,10 / 19$ and $10 / 26$. When a wire was detected, the snout was collected for extraction and analysis that occurred in the laboratory at a later date. All of the brood stock collected for the summer Chinook segregated program came from an Upper Columbia River Hatchery program. The CJH integrated program was the largest contributor to segregated brood with ( $\mathrm{n}=252$ ) 53.6\% of adults coming from either the Similkameen or Omak Pond. Other Upper Columbia River Hatcheries contributed ( $\mathrm{n}=162$ ) $34.4 \%$, most of which were from Wells (20.0\%) and Chelan Falls (9.2\%) hatcheries. A large portion of snouts ( $\mathrm{n}=47$ ) indicated detection during spawning events but a coded-wire tag was not found during extraction. Reasons for this include but are not limited to rapidly shaking a Northwest Marine Technologies (NMT) T-Wand when scanning for a cwt (false positive in the field), failure to detect a tag in the lab (false negative), metals in the soil that transfer to a fish during handling or hooks or other metal debris in the fish's head. The unknown component represents $11.9 \%$ of the 2016 segregated brood (Table A 18). A relatively large percentage of the segregated CJH does not receive a CWT, however, in 2016 only jacks returned from the CJH segregated program and jacks were not collected for broodstock. A few 3 year old males may still have been included in the brood collection effort if they exceeded 66 cm fork length. The high rate of no CWT detection in the lab is concerning because all summer Chinook programs upstream of Priest Rapids Dam are supposed to have a $100 \%$ tag rate (except for CJH segregated). Most tag rates reported in RMIS are greater than $98 \%$, indicating that there should be a relatively small percentage of adult summer Chinook that have an adipose clip and no CWT, particularly in 2016.

Table A 18. Composition of hatchery-origin brood, by program, collected for the CJH segregated summer/fall Chinook program in 2016.

| Category | Hatchery Program | \# tags | \% of brood |  |
| :---: | :---: | :---: | :---: | :---: |
| Okanogan Integrated | Similkameen | 233 | $49.6 \%$ | $53.6 \%$ |
|  | Omak Pond | 19 | $4.0 \%$ |  |
| Other UCR summer/fall | Carlton | 6 | $1.3 \%$ |  |
|  | Wells | 94 | $20.0 \%$ |  |
|  | Chelan Falls | 43 | $9.2 \%$ | $34.4 \%$ |
|  | Entiat | 10 | $2.1 \%$ |  |
|  | Dryden | 9 | $1.9 \%$ |  |
| Unknown | Lost/scratched Tags | 9 | $1.9 \%$ | $11.9 \%$ |
|  | No tag in snout | 47 | $10.0 \%$ |  |
| Total |  | 470 | $100.0 \%$ |  |

## Integrated Program Broodstock Age Structure

Scales are taken from summer Chinook integrated Program broodstock in order to capture the age of successfully spawned fish. In 2016, the integrated program was comprised of mostly four-year old male and female fish, whereas the segregated program skewed heavily to five-year old fish. (Figure A 1).





Figure A 1. The total and salt ages of the 2016 broodstock, males and females, collected for the Okanogan summer/fall Chinook integrated program.

## Segregated Program Broodstock Age Structure

Coded wire tags are extracted from summer Chinook segregated program broodstock and later read in order to capture the age of successfully spawned fish.





Figure A 2. The total and salt ages of the 2016 broodstock, males and females, collected for the Chief Joseph Hatchery segregated program.

## Incubation

Eggs from each female summer/fall Chinook were placed in individual incubators (Heath Trays) and remained individually incubated until ELISA results were obtained. Once eye-up occurred, eggs from any moderate and high ELISA would be removed; no females were discarded from the 2016 brood. The cull rate for this production plan allows for a rate of $5 \%$ for segregated and $3 \%$ for integrated. After eye-up, egg mortality was removed and the eggs were inventoried and put into incubators at 5,000 eggs per tray for hatching. Incubation water temperatures were manipulated to the level necessary to synchronize the hatching and ponding of the spawn takes throughout October and November 2016 and to achieve the size-at-release target for both yearling and sub-yearling summer Chinook programs. Four days after spawning, the incubation water temperatures were gradually reduced on yearling egg-takes to a temperature of $37^{\circ} \mathrm{F}$. Once each take achieved 100 TUs, incubation temperatures were, manipulated to either advance or delay maturation. Variable incubation water temperatures were required to synchronize hatching dates associated with variable spawn dates throughout the spawn period within yearling and sub-yearling production groups and to achieve target hatching date associated with size-at-release targets, based on projected growth rates and release dates for the respective production groups.

## Rearing

The first groups of sub-yearlings were brought out of incubation and transferred into early rearing troughs on February $1^{\text {st }}, 2017$. During the month of February, this group was introduced to feed in the early rearing troughs, and reared for a period of two weeks. Ponding was completed on February 11 th, 2017. After the initial rearing period inside, all groups were transferred outside to the standard raceways via the fry transfer line. No inventories were taken during transfers, to prevent excess handling stress. All subyearlings are released in the first spring of life, and after marking, both the integrated and segregated sub-yearlings were released in May of 2017. The integrated sub-yearlings were transferred to the Omak Acclimation pond after marking. In addition, both groups included a 5,000 PIT tag component. PIT tagging was contracted to USFWS, and was completed on April 20 ${ }^{\text {th }}, 2017$.

Table A 19. Chief Joseph Hatchery brood year 2016 summer/fall Chinook sub-yearling ponding summary. The survival standard for this life stage was $95 \%$.

| Production Group | $\begin{aligned} & \frac{\text { Total }}{\underline{\text { Fry }}} \\ & \text { Ponded } \end{aligned}$ | Ponding Mortality | Monthly Feed | $\begin{aligned} & \text { Monthly } \\ & \text { Mortality } \end{aligned}$ | Ponding Loss (\%) | Cumulative Survival (\%) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Total | Total | Total | Total | Total | Total |
| NOR |  |  |  |  |  |  |
| Subs | 239,646 | 6,012 | 190 | 1,848 | 2.5\% | 96.7\% |
| Sub-total | 239,646 | 6,012 | 190 | 1,848 | 2.5\% | 96.7\% |
| HOR |  |  |  |  |  |  |
| Subs | 187,237 | 1,624 | 138 | 1,355 | 0.9\% | 98.4\% |
| Sub-total | 187,237 | 1,624 | 138 | 1,355 | 0.9\% | 98.4\% |
| Total | 426,883 | 7,636 | 328 | 3,203 | 3.4\% | $\begin{aligned} & 97.5 \\ & \% \\ & \hline \end{aligned}$ |

Table A 20. Chief Joseph Hatchery brood year 2016 summer/fall Chinook sub-yearling rearing summary., May 2017.

|  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |

The first group of integrated yearlings was brought out of incubation and transferred into early rearing troughs in late May, and continued into June. Once ponded, all groups were introduced to feed in the early rearing troughs, and remained in early rearing for a period of two weeks. After the initial rearing period inside, groups will be transferred outside to the standard raceways via the fry transfer line. Ponding survival ranged from $96 \%$ to $99 \%$ and averaged $98 \%$ across all groups which exceeded the survival standard (95\%) for this life stage (Table A 21).

Table A 21. Chief Joseph Hatchery brood year 2016 summer/fall Chinook yearling ponding summary. The survival standard for this life stage was $95 \%$.

|  | $\frac{\text { Total Fry }}{\text { Ponded }}$ | $\begin{aligned} & \text { Ponding } \\ & \text { Mortality } \end{aligned}$ | $\frac{\text { Monthly }}{\text { Feed }}$ | $\begin{aligned} & \text { Monthly } \\ & \hline \text { Mortality } \end{aligned}$ | Ponding Loss (\%) | $\begin{aligned} & \text { Cumulative } \\ & \underline{\text { Survival (\%) }} \text { ( } \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Production Group | Total | Total | $\begin{aligned} & \text { Tota } \\ & 1 \\ & \hline \end{aligned}$ | Total | Total | Total |
| NORs |  |  |  |  |  |  |
| May | $\begin{gathered} 384,48 \\ 0 \\ \hline \end{gathered}$ | $\begin{gathered} 11,20 \\ 0 \\ \hline \end{gathered}$ | 17 | 2,283 | 2.91\% | 96.5\% |
| Sub-total | $\begin{gathered} 384,48 \\ 0 \end{gathered}$ | $\begin{gathered} 11,20 \\ 0 \\ \hline \end{gathered}$ | 17 | 2,283 | 2.91\% | 96.5\% |
| HORs |  |  |  |  |  |  |
| May | $\begin{gathered} 262,01 \\ 8 \\ \hline \end{gathered}$ | 3,458 | 15 | 1,042 | 1.31\% | 98.3\% |
| Sub-total | $\begin{gathered} 262,01 \\ 8 \end{gathered}$ | 3,458 | 15 | 1,042 | 1.31\% | 98.3\% |
| Total | $\begin{gathered} 646,49 \\ 8 \end{gathered}$ | $\begin{gathered} 14,65 \\ 8 \end{gathered}$ | 32 | 3,325 | 2.11\% | 97.4\% |

## Chief Joseph Hatchery Ladder

The CJH ladder is operated with the primary purpose of adult management (reducing pHOS ) but can also be utilized to collect brood for the segregated program in years of low abundance or if the purse seine is not effective due to environmental conditions. In 2016, a small portion (3\%) of hatchery fish collected from the CJH ladder were used for brood in the segregated program (Table A 22). The CJH fish ladder began operation on June $27^{\text {th }}, 2016$, with the first adult management activities occurring on June 29th. In 2016, only 3 year old fish were returning from the CJH program so adult returns were stray fish (or wandering9) from the Similkameen Pond program (pre-CJH) and other downstream facilities. All hatchery Chinook and Sockeye were removed from the ladder and utilized for Tribal subsistence and ceremonial food purposes. All steelhead and NOR Chinook were returned to the river via a water to water transfer.

From June $27^{\text {th }}$ thru October $6^{\text {th }}, 2016,6,354$ hatchery-origin summer/fall Chinook and five Sockeye were removed at the CJH ladder. 6,206 of the Chinook were utilized for Tribal subsistence purposes and 148 were shipped to the landfill as they were not deemed

[^10]suitable for human consumption. 556 natural-origin Summer/Fall Chinook, and 56 steelhead were trapped, handled and released back to the Columbia River. (Table, A 22,Table A 23). Of the AD-present steelhead handled and released, $67 \%$ were determined to be of hatchery origin, based on the stock composition developed by WDFW at Wells Dam.

The encounter/handling and release of four NOR steelhead represents $35 \%$ of the allowable incidental take provided in the Biological Opinion (BiOp) for Chief Joseph Hatchery collection facilities (NMFS 2008). There were no observed immediate steelhead mortalities during the ladder operations in 2016.

Table A 22. Chief Joseph Hatchery adult summer/fall Chinook ladder operations from June to October 2016.

| Date | \# of Ladder Trap Checks | HOR Males surplussed | HOR <br> Females surplussed | HOR Jacks <br> (1) <br> surplussed | HOR <br> Brood <br> Males | HOR <br> Brood <br> emales | NOR <br> Males RTS | NOR <br> Females RTS | NOR <br> Jacks <br> RTS |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| June | 1 | 79 | 89 | 88 | 0 | 0 | 7 | 8 | 2 |
| July | 8 | 1,566 | 1,530 | 400 | 70 | 70 | 144 | 106 | 42 |
| Aug | 5 | 816 | 677 | 329 | 27 | 28 | 72 | 54 | 27 |
| Sept | 3 | 261 | 195 | 176 | 0 | 0 | 39 | 32 | 20 |
| Oct | 2 | 74 | 72 | 2 | 0 | 0 | 2 | 1 | 0 |
| Total | 19 | 2,796 | 2,563 | 995 | 97 | 98 | 264 | 201 | 91 |

RTS= Return to stream

Table A 23. Chief Joseph Hatchery adult spring Chinook, sockeye and steelhead ladder operations from June to October 2016.

| Date | \# of <br> Ladder <br> Trap <br> Checks | HOR <br> Spring <br> Chinook | HOR <br> Spring <br> Chinook <br> Jacks | NOR <br> Spring Chinook RTS | NOR <br> Spring Chinook Jacks RTS | Sockeye | AD Present Steelhead RTS | AD Absent Steelhead RTS | Coho RTS |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| June | 1 | 21 | 82 | 0 | 0 | 0 | 0 | 0 | 0 |
| July | 8 | 22 | 11 | 0 | 0 | 3 | 1 | 2 | 0 |
| Aug | 5 | 0 | 0 | 0 | 0 | 2 | 1 | 13 | 0 |
| Sept | 3 | 0 | 0 | 0 | 0 | 0 | 8 | 28 | 0 |
| Oct | 2 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 0 |
| Total | 19 | 43 | 93 | 0 | 0 | 5 | 11 | 45 | 0 |

RTS= Return to stream

Table A 24. Chief Joseph Hatchery annual summer/fall Chinook, sockeye, and steelhead collected during ladder operations.

| Date | HOR <br> Chinook surplussed | HOR jacks ${ }^{(1)}$ surplussed | NOR <br> Chinook <br> RTS | NOR <br> jack <br> RTS | HOR <br> Chinook Brood | Sockeye | AD Present Steelhead RTS | AD Absent Steelhead RTS | Coho RTS |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Aug.- |  |  |  |  |  |  |  |  |  |
| Nov. 2013 | 1,263 | 523 | 247 | 69 | 9 | 10 | 38 | 0 | 0 |
| July- |  |  |  |  |  |  |  |  |  |
| Nov. |  |  |  |  |  |  |  |  |  |
| $\begin{aligned} & 2014 \\ & \text { July- } \end{aligned}$ | 2,835 | 1,778 | 861 | 245 | 87 | 31 | 69 | 122 | $181{ }^{6}$ |
| $\begin{aligned} & \text { Oct. } \\ & 2015 \end{aligned}$ | 6,773 | 1,651 | 1,671 | 369 | 2174 | 180 | 1192 | 401 | 2 |
| June- |  |  |  |  |  |  |  |  |  |
| Oct. |  |  |  |  |  |  |  |  |  |
| 2016 | 5,359 | 995 | 465 | 91 | 1965 | 5 | $11^{3}$ | 45 | 0 |
| Total | 14,967 | 4,424 | 2,997 | 705 | 509 | 216 | 199 | 568 | 183 |

${ }^{(1)}$ Includes mini-jacks
${ }^{(2)} 24 \%$ AD Present Steelhead were HORs
(3) $67 \%$ AD Present Steelhead were HORs
(4) 147 adults ( 80 males, 67 females) taken for transfer to Eastbank Hatchery
${ }^{(5)} 98$ males and 98 females taken in July and August,
${ }^{(6)}$ Surplussed fish
RTS= Return to stream

Table A 25. Chief Joseph Hatchery spring Chinook collected during ladder operations in 2016.

| Date | HOR Chinook surplussed | HOR Chinook Jacks surplussed | NOR <br> Chinook <br> RTS | NOR <br> Chinook Jacks RTS | HOR <br> Brood |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 6/29 | 21 | 82 | 0 | 0 | 0 |
| 7/7 | 5 | 0 | 0 | 0 | 0 |
| 7/13 | 1 | 3 | 0 | 0 | 0 |
| 7/19 | 8 | 1 | 0 | 0 | 0 |
| 7/21 | 1 | 1 | 0 | 0 | 0 |
| 7/22 | 5 | 3 | 0 | 0 | 0 |
| 7/26 | 1 | 1 | 0 | 0 | 0 |
| 7/28 | 1 | 2 | 0 | 0 | 0 |
| Total | 43 | 93 | 0 | 0 | 0 |

RTS= Return to stream
The ladder was closed and dewatered on October $6^{\text {th }}, 2016$, for the season. The protocol was to sample 20\% (one of five) of the adipose-clipped summer/fall Chinook for code-wire tags (CWT). Snouts with positive CWT detection were held frozen until December 2016 when CWT extraction and reading took place in the Chief Joseph Hatchery lab. Coded wire tags were extracted and read from snouts collected from 2013-2016 (we had to catch up on 3 years of backlogged snouts after WDFW decided they could not process them without additional funding). Recovery data were expanded by the tag rate at the hatchery of origin and the sample rate at the ladder. Please refer to the Methods section for details on the expansion process for recovered tags. Beginning with jacks in 2016, snouts without a tag were assumed to be from the CJH segregated program.

Eight summer/fall Chinook hatchery programs were encountered during ladder operations at CJH in 2013, most of which came from upstream of Priest Rapids (Table A 26). The Chelan (33\%) and Wells (22\%) hatcheries as well as the Dryden Pond (26\%) were the most common. Similkameen Pond was a small proportion (10\%) of the composition (Table A 26). The Omak Pond and CJH segregated programs were not encountered in the ladder surplus, primarily because the first CJH releases didn't occur until 2014.

Table A 26. Summary of summer/fall Chinook coded-wire tags encountered and expansions for the CJH ladder in 2013.

| Category | Hatchery <br> Program | \# Tags | Expanded Abundance | \% of Ladder Surplus |
| :---: | :---: | :---: | :---: | :---: |
| Okanogan <br> Integrated | Omak Yearlings | 0 | 0 | 0\% |
|  | Omak Subyearlings | 0 | 0 | 0\% |
|  | Similkameen | 38 | 86 | 10\% |
| CJH Segregated | Segregated yearlings | 0 | 0 | 0\% |
|  | Segregated subyearlings | 0 | 0 | 0\% |
|  | No CWT, assumed Seg. | 0 | 0 | 0\% |
| Other UCR summer/fall Chinook hatchery | Wells | 87 | 201 | 22\% |
|  | Chelan | 128 | 300 | 33\% |
|  | Carlton | 34 | 77 | 8\% |
|  | Entiat | 0 | 0 | 0\% |
|  | Dryden | 99 | 237 | 26\% |
|  | Priest Rapids | 2 | 4 | 0.5\% |
| Out of ESU hatchery | Lyons Ferry | 2 | 4 | 0.5\% |
| Total |  | 390 | 909 | 100\% |

In 2014 eleven summer/fall Chinook hatchery programs were encountered at the CJH ladder, most of which still came from upstream of Priest Rapids (Table A 27). Similar to 2013, Wells (28\%), and Chelan (26\%) hatcheries were the most common. Dryden (11\%), Carlton (8\%) and Similkameen (10\%) were a smaller proportion of the composition (Table A 27). The Omak Pond and CJH segregated programs were not represented in the ladder surplus because there were no adult returns from those programs in 2014 (the first releases occurred in 2014).

Table A 27. Summary of summer/fall Chinook coded-wire tags encountered and expansions for the CJH ladder in 2014.

| Category | Hatchery <br> Program | \# Tags | Expanded Abundance | \% of Ladder <br> Surplus |
| :---: | :---: | :---: | :---: | :---: |
| Okanogan <br> Integrated | Omak Yearlings | 0 | 0 | 0\% |
|  | Omak Subyearlings | 0 | 0 | 0\% |
|  | Similkameen | 111 | 197 | 10\% |
| CJH Segregated | Segregated yearlings | 0 | 0 | 0\% |
|  | Segregated subyearlings | 0 | 0 | 0\% |
|  | No CWT, assumed Seg. | 190 | 300 | 15\% |
| Other UCR summer/fall Chinook hatchery | Wells | 312 | 569 | 28\% |
|  | Chelan | 290 | 526 | 26\% |
|  | Carlton | 84 | 152 | 8\% |
|  | Entiat | 18 | 32 | 2\% |
|  | Dryden | 123 | 220 | 11\% |
|  | Priest Rapids | 1 | 2 | 0.1\% |
| Out of ESU hatchery | Irrigon | 1 | 2 | 0.1\% |
|  | Lyons Ferry | 2 | 4 | 0.2\% |
|  | Nez Perce Tribal | 1 | 2 | 0.1\% |
|  | Oxbow | 1 | 2 | 0.1\% |
| Total |  | 1,134 | 2,008 | 100\% |

In 2015 twelve summer/fall Chinook hatchery programs were encountered at the CJH ladder, most of which still came from upstream of Priest Rapids (Table A 28). Again, similar to previous years, Wells (34\%) and Chelan (29\%) hatcheries were the most common and Dryden (12\%) and Similkameen (13\%) remained a small proportion of the composition (Table A 28). The Omak Pond and CJH segregated programs were represented at less than $3 \%$ of the overall ladder surplus.

Table A 28. Summary of summer/fall Chinook coded-wire tags encountered and expansions for the CJH ladder in 2015.

| Category | Hatchery <br> Program | \# Tags | Expanded Abundance | \% of <br> Ladder <br> Surplus |
| :---: | :---: | :---: | :---: | :---: |
| Okanogan <br> Integrated | Omak Yearlings | 2 | 7 | 0.1\% |
|  | Omak Subyearlings | 2 | 8 | 0.1\% |
|  | Similkameen | 224 | 815 | 13\% |
| CJH Segregated | Segregated yearlings | 0 | 0 | 0\% |
|  | Segregated subyearlings | 0 | 0 | 0\% |
|  | No CWT, presumed Seg | 30 | 86 | 1\% |
| Other UCR summer/fall Chinook hatchery | Wells | 610 | 2212 | 34\% |
|  | Chelan | 518 | 1898 | 29\% |
|  | Carlton | 102 | 378 | 6\% |
|  | Dryden | 218 | 795 | 12\% |
|  | Entiat | 72 | 261 | 4\% |
|  | Priest Rapids | 4 | 15 | 0.2\% |
| Out of ESU hatchery | Irrigon | 1 | 4 | 0.1\% |
|  | Nez Perce Tribal | 3 | 11 | 0.2\% |
|  | Oxbow | 1 | 4 | 0.1\% |
|  | Lyons Ferry | 4 | 15 | 0.2\% |
|  | McCall | 1 | 4 | 0.1\% |
| Total |  | 1,792 | 6,513 | 100\% |

Eleven summer/fall Chinook hatchery programs were encountered at the CJH ladder in 2016, most of which still came from upstream of Priest Rapids (Table A 29). Wells (50\%) and Chelan (26\%) hatcheries were the most common and Dryden (8\%) and Similkameen (3\%) still remained a small proportion of the composition (Table A 29). The Omak Pond (2\%) program made up a small percentage of the overall ladder surplus while the CJH segregated program made up 5\% of the composition. Most of these recoveries were from ad-clipped, non-coded wire tagged (CWT) fish and are presumed to be from the CJH segregated program since this is the only hatchery above Priest Rapids that releases ad-clipped, non-CWT fish.

Table A 29. Summary of summer/fall Chinook coded-wire tags encountered and expansions for the CJH ladder in 2016.

| Category | Hatchery <br> Program | \# Tags | Expanded Abundance | \% of <br> Ladder <br> Surplus |
| :---: | :---: | :---: | :---: | :---: |
| Okanogan Integrated | Omak Yearlings | 14 | 64 | 1\% |
|  | Omak Subyearlings | 12 | 57 | 1\% |
|  | Similkameen | 37 | 170 | 3\% |
| CJH Segregated | Segregated yearlings | 0 | 0 | 0\% |
|  | Segregated subyearlings | 24 | 110 | 2\% |
|  | No CWT, presumed Segr. | 76 | 321 | 5\% |
| Other UCR summer/fall Chinook hatchery | Wells | 626 | 2864 | 50\% |
|  | Chelan | 323 | 1486 | 26\% |
|  | Carlton | 21 | 110 | 2\% |
|  | Entiat | 28 | 128 | 2\% |
|  | Dryden | 95 | 440 | 8\% |
| Out of ESU hatchery | Irrigon | 2 | 9 | 0.2\% |
|  | Nez Perce Tribal | 1 | 5 | 0.1\% |
|  | Lyons Ferry | 1 | 5 | 0.1\% |
|  | Umatilla | 1 | 5 | 0.1\% |
| Total |  | 1,261 | 5,774 | 100\% |

Table A 30. Percent of CJH ladder surplus summer/fall Chinook each year estimated to be from various facilities based on CWT assessment. Similkameen includes some returns from Bonaparte Pond releases (2010 and 2011). Chelan includes returns from the Turtle Rock program (2010 and 2011). 2016 was the first year of jack returns ( 3 year old males) to CJH. 2018 will be the first return year with a full complement of brood years in the return (through age 5).

|  | Surplus Fish | Facility/Program |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | $\begin{aligned} & \hline \text { CJH } \\ & \text { Seg. } \end{aligned}$ | Omak | Similk ${ }^{\text {b }}$ | Wells | Chelan ${ }^{\text {c }}$ | Carlton | Entiat | Dryden ${ }^{\text {d }}$ | Priest | Other ${ }^{\text {e }}$ |
| 2013 | 1,061 | 0\% | 0\% | 10\% | 22\% | 33\% | 8\% | 0\% | 26\% | 1\% | 1\% |
| 2014 | 2,008 | 0\% | 0\% | 10\% | 28\% | 26\% | 8\% | 2\% | 11\% | 0\% | 0\% |
| 2015 | 6,802 | 1\% | 0\% | 13\% | 34\% | 29\% | 6\% | 4\% | 12\% | 0\% | 0\% |
| 2016 | 5,788 | 5\% | 2\% | 3\% | 50\% | 26\% | 2\% | 2\% | 8\% | 0\% | 0\% |
| Average | 3,915 | 2\% | 1\% | 9\% | 26\% | 29\% | 6\% | 2\% | 14\% | 0\% | 0\% |

'Includes recoveries with 'no coded wire tags' under 62 cm . in fork length, 2015 and 2016
${ }^{\text {b }}$ Includes Bonaparte pond releases, all years

${ }^{d}$ Includes releases by the Eastbank Hatchery into the Wenatchee R. (2013)
${ }^{\text {e }}$ Includes releases by Lyons Ferry and Snake River basin hatcheries

The spring Chinook CWT recovery data from the CJH ladder represents spring Chinook encountered during summer Chinook ladder operations (June 29-October 6). In 2016, the majority (43\%) of spring Chinook encountered at the CJH ladder were from the CJH segregated program, although only jacks were returning from that program (Table A 31). Thirteen percent of the Spring Chinook encountered at the ladder were from the Okanogan 10(j) program at Riverside Pond. and the contribution from LNFH dropped dramatically to 13\% compared to previous years. Interestingly, the percent of 'Other' out-ofESU contributors increased in 2016 and came from 2 different hatcheries (Table A 31). However, conclusions should not be made without sampling during the earlier run timing of spring Chinook and considering that a full set of brood years was not returning from the releases at CJH .

Table A 31. Percent of CJH ladder surplus spring Chinook each year estimated to be from various facilities based on CWT assessment of spring Chinook. Estimated number of annual spring Chinook coded wire tag recoveries, by release hatchery, from Chief Joseph Hatchery ladder operations in June to October.

|  | $\#$ <br> Surplus <br> Fish | Facility/Program |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Riverside Pond | CJH | Winthrop | Leavenworth | Chiwawa <br> Pond | Methow <br> Hatchery | Othera |  |  |
| 2013 | 3 | $0 \%$ | $0 \%$ | $0 \%$ | $0 \%$ | $100 \%$ | $0 \%$ | $0 \%$ |  |
| 2014 | 46 | $0 \%$ | $0 \%$ | $0 \%$ | $91 \%$ | $7 \%$ | $2 \%$ | $0 \%$ |  |
| 2015 | 24 | $0 \%$ | $0 \%$ | $4 \%$ | $75 \%$ | $17 \%$ | $0 \%$ | $4 \%$ |  |
| 2016 | 17 | $13 \%$ | $43 \%$ | $6 \%$ | $13 \%$ | $13 \%$ | $6 \%$ | $6 \%$ |  |
| Average | $\mathbf{2 3}$ | $\mathbf{3 \%}$ | $\mathbf{1 1 \%}$ | $\mathbf{3 \%}$ | $\mathbf{4 4 \%}$ | $\mathbf{3 4 \%}$ | $\mathbf{2 \%} \%$ | $\mathbf{3 \%} \%$ |  |

[^11]
## Appendix B

## 2017 Production Plan

Table B 1. Summer Chinook Early - Integrated Program (Similkameen Release)

| Chief Joseph Hatchery Production Plan |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Brood Year: | 2017 |  |  |  |  |  | Planting Goal: | 267,000 |  |  |
| Species: | Summer Chinook - Early |  |  |  |  |  | Pounds: | 26,700 |  |  |
| Stock: | Okanogan |  |  |  |  |  |  |  |  |  |
| Origin: | Wild |  |  |  |  |  |  |  |  |  |
| Program: | Integrated |  |  |  |  |  |  |  |  |  |
|  | 360,000 |  |  |  |  |  |  |  |  |  |
| Egg Take Goal: |  |  |  |  |  |  | Adult Goal: | 158 |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Estimated Release Data: |  |  |  |  |  |  |  |  |  |  |
| Start Date:$04 / 15 / 19$ | End Date: |  |  |  |  |  |  |  |  |  |
|  |  | Num Released | fish per lb. | Wt. grams | Total weight (lb.) | Total weight (kg) | Life Stage | Release Site | Mark Type | Tagged |
|  | 04/30/19 | 267,000 | 10 | 45 | 26,700 | 12,015 | yearlings | Similkameen | Ad Clipped | CWT |
|  |  |  |  |  |  |  |  |  |  |  |
| Notes: | Egg take goal includes 3\% for culling. |  |  |  |  |  |  |  |  |  |
|  | Adult Goal includes 10\% pre-spawn mortality |  |  |  |  |  |  |  |  |  |
|  | 10\% Green to Eyed egg mortality |  |  |  |  |  |  |  |  |  |
|  | Rearing mortality is $16.7 \%$ |  |  |  |  |  |  |  |  |  |
| Rearing Summary: |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Species | Source | Date | Number Green Eggs | Number Eyed Eggs | Number Ponded | Fed Fry | Released | Location |  |  |
| EA SU Chinook YR | Okanogan | April | 349,200 | 314,280 | 298,566 | 283,638 | 267,000 | Similkameen |  |  |

Table B 2. Summer Chinook Late - Integrated Program (Omak Acclimation Pond Release)

| Chief Joseph Hatchery Production Plan |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Brood Year: | 2017 |  |  |  |  |  | Planting Goal: | 566,000 |  |  |
| Species: | Summer Chinook - Late |  |  |  |  |  | Pounds: | 32,600 |  |  |
| Stock: | Okanogan |  |  |  |  |  |  |  |  |  |
| Origin: | Wild |  |  |  |  |  |  |  |  |  |
| Program: | Integrated |  |  |  |  |  |  |  |  |  |
| Egg Take Goal: | 765,000 |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  | Adult Goal: | $340^{\circ}$ |  |  |
| Estimated Release Data: |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Start Date: <br> 04/15/19 | End Date:$04 / 30 / 19$ | Num Released | fish per lb. | Wt. grams | Total weight (lb.) | Total weight (kg) | Life Stage | Release Site | Mark Type | Tagged |
|  |  | 266,000 | 10 | 45 | 26,600 | 11,970 | yearlings | Omak | Ad Clipped | CWT |
| 05/15/18 | 06/01/18 | 300,000 | 50 | 11 | 6,000 | 3,300 | sub-yearling | Omak | Ad Clipped | CWT |
| Notes: | Egg take goal includes 3\% for culling. |  |  |  |  |  |  |  |  |  |
|  | Adult Goal includes 10\% pre-spawn mortality |  |  |  |  |  |  |  |  |  |
|  | 10\% Green to Eyed egg mortality |  |  |  |  |  |  |  |  |  |
|  | Rearing mortality is $15.7 \%$ for yearlings, $16.2 \%$ for subs. |  |  |  |  |  |  |  |  |  |
| Rearing Summary: |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Species | Source | Date | Number Green Eggs | Number Eyed Eggs | Number Ponded | Fed Fry | Released | Location |  |  |
| EA SU Chinook YR | Okanogan | April | 348,764 | 313,887 | 298,193 | 283,283 | 266,000 | Omak |  |  |
| EA SU Chinook Sub | Okanogan | June | 393,287 | 353,958 | 336,260 | 319,447 | 300,000 | Omak |  |  |

Table B 3. Summer Chinook Late - Segregated Program (CJH Site Release)


Table B 4. Summer Chinook Early - Integrated Program (Riverside Acclimation Pond Release)


Table B 5. Summer Chinook Early - Segregated Program (CJH Release Site)

| Chief Joseph Hatchery Production Plan |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Brood Year: | 2017 |  |  |  |  |  | Planting Goal: | 450,000 |  |  |
| Species: | Summer Chinook - Early |  |  |  |  |  | Pounds: | 29,000 |  |  |
| Stock: | Okanogan |  |  |  |  |  |  |  |  |  |
| Origin: | Hatchery |  |  |  |  |  |  |  |  |  |
| Program: | Segregated |  |  |  |  |  |  |  |  |  |
| Egg Take Goal: | - |  |  |  |  |  |  |  |  |  |
|  | 620,000 |  |  |  |  |  | Adult Goal: | 276 |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Estimated Release Data: |  |  |  |  |  |  |  |  |  |  |
|  | End Date: Num Released |  |  |  |  |  |  |  |  |  |
| Start Date: |  |  | fish per lb. | Wt. grams | Total weight (lb.) | Total weight (kg) | Life Stage | Release Site | Mark Type | Tagged |
| 04/15/19 | 04/30/1906/01/18 | 250,000 | 10 | 45 | 25,000 | 11,250 | yearlings | CJ hatchery | Ad Clipped |  |
| 05/15/18 |  | 200,000 | 50 | 9 | 4,000 | 1,800 | sub-yearling | CJ hatchery | Ad Clipped |  |
| Notes: | Egg take goal includes 5\% for culling. |  |  |  |  |  |  |  |  |  |
|  | Adult Goal includes 10\% pre-spawn mortality |  |  |  |  |  |  |  |  |  |
|  | 10\% Green to Eyed egg mortality |  |  |  |  |  |  |  |  |  |
|  | Rearing mortality is $14.4 \%$ for yearlings, $16.5 \%$ for subs. |  |  |  |  |  |  |  |  |  |
| Rearing Summary: |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Species | Source | Date | Number Green Eggs | $\begin{gathered} \hline \text { Number } \\ \text { Eyed Eggs } \\ \hline \end{gathered}$ | Number Ponded | Fed fry | Released | Location |  |  |
| EA SU Chinook YR | Okanogan | April | 323,950 | 291,555 | 276,977 | 263,128 | 250,000 | CJ Hatchery |  |  |
| EA SU Chinook Sub | Okanogan | June | 265,050 | 238,545 | 226,618 | 215,287 | 200,000 | CJ Hatchery |  |  |

Table B 6. Spring Chinook - Leavenworth (CJH Release)


Table B 7. Spring Chinook - Met Comp (Riverside Acclimation Pond Release)


## Appendix C

## Reach Weighted Effective pHOS

Table C 1. pHOS information for adjustments based on hatchery fish effectiveness (relative reproductive success assumption) and the reach weighting based on the proportion of redds in each reach in the Okanogan River from 2006 to 2016.

| Brood Year |  | Number of Summer Chinook Carcasses |  |  |  |  |  |  |  | Total | Effective Reach Weighted $\mathbf{p H O S}{ }^{1}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Okanogan |  |  |  |  |  | Similkameen |  |  |  |
|  |  | O-1 | O-2 | O-3 | O-4 | O-5 | O-6 | S-1 | S-2 |  |  |
| 2006 | NOS | 2 | 2 | 22 | 10 | 105 | 247 | 370 | 73 | 831 | 18.0\% |
|  | HOS | 2 | 1 | 9 | 6 | 15 | 44 | 138 | 33 | 248 |  |
|  | Effective pHOS ${ }^{2}$ | 44.4\% | 28.6\% | 24.7\% | 32.4\% | 10.3\% | 12.5\% | 23.0\% | 26.6\% |  |  |
|  | \% Redds | 0.2\% | 1.3\% | 4.1\% | 3.4\% | 19.7\% | 32.0\% | 29.9\% | 9.5\% | 100\% |  |
| 2007 | NOS | 1 | 0 | 30 | 1 | 284 | 322 | 405 | 20 | 1063 | 31.7\% |
|  | HOS | 1 | 0 | 25 | 0 | 169 | 197 | 253 | 9 | 654 |  |
|  | Effective pHOS ${ }^{2}$ | 44.4\% | 0.0\% | 40.0\% | 0.0\% | 32.3\% | 32.9\% | 33.3\% | 26.5\% |  |  |
|  | \% Redds | 0.2\% | 0.8\% | 5.8\% | 3.1\% | 27.3\% | 27.6\% | 31.0\% | 4.3\% | 100\% |  |
| 2008 | NOS | 2 | 1 | 14 | 11 | 107 | 324 | 347 | 41 | 847 | 54.3\% |
|  | HOS | 2 | 9 | 26 | 25 | 141 | 341 | 512 | 116 | 1172 |  |
|  | Effective pHOS ${ }^{2}$ | 44.4\% | 87.8\% | 59.8\% | 64.5\% | 51.3\% | 45.7\% | 54.1\% | 69.4\% |  |  |
|  | \% Redds | 0.2\% | 2.4\% | 2.8\% | 4.5\% | 17.4\% | 26.2\% | 37.3\% | 9.3\% | 100\% |  |
| 2009 | NOS | 2 | 3 | 13 | 14 | 189 | 347 | 330 | 75 | 973 | 40.4\% |



|  | Effective pHOS ${ }^{2}$ | 100.0\% | 0.0\% | 20.2\% | 10.6\% | 12.6\% | 7.0\% | 13.9\% | 10.8\% |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | \% Redds | 0.3\% | 1.3\% | 4.5\% | 2.6\% | 20.0\% | 23.7\% | 40.8\% | 6.7\% | 100\% |  |
| 2015 | NOS | 0 | 5 | 39 | 9 | 209 | 931 | 1186 | 176 | 2555 | 21.0\% |
|  | HOS | 0 | 5 | 22 | 2 | 74 | 63 | 516 | 56 | 738 |  |
|  | Effective pHOS ${ }^{2}$ | 20.8\% | 44.4\% | 31.1\% | 15.1\% | 22.1\% | 5.1\% | 25.8\% | 20.3\% |  |  |
|  | \% Redds | 0.8\% | 2.6\% | 6.6\% | 1.8\% | 23.6\% | 20.1\% | 37.7\% | 6.7\% | 100\% |  |
| 2016 | NOS | 0 | 6 | 13 | 7 | 186 | 1019 | 819 | 121 | 2171 | 15.5\% |
|  | HOS | 0 | 6 | 1 | 4 | 44 | 56 | 395 | 78 | 584 |  |
|  | Effective pHOS ${ }^{2}$ | 15.5\% | 44.4\% | 5.8\% | 31.4\% | 15.9\% | 4.2\% | 27.8\% | 34.0\% |  |  |
|  | \% Redds | 0.0\% | 1.1\% | 1.0\% | 2.5\% | 17.2\% | 44.3\% | 31.2\% | 2.7\% | 100\% |  |
| Average \% Redds |  | 0.3\% | 1.5\% | 4.1\% | 2.9\% | 19.4\% | 29.7\% | 35.9\% | 6.2\% |  |  |
| Average | ctive pHOS | 44.4\% | 52.2\% | 42.0\% | 46.5\% | 33.0\% | 19.1\% | 38.9\% | 42.6\% |  |  |
| Average Reach Weighted Effective pHOS $=$ |  |  |  |  |  |  |  |  |  |  | 29.4\% |

Table C 2 Number of hatchery- and natural-origin (wild) summer Chinook carcasses collected in each reach of the Okanogan (01-06) and Similkameen rivers from 1993 to 2016.

| Survey year | Origin | Survey reach |  |  |  |  |  |  |  | Tota I |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | O-1 | O-2 | O-3 | O-4 | O-5 | O-6 | S-1 | S-2 |  |
| $1993{ }^{\text {a }}$ | Wild | 0 | 0 | 3 | 0 | 13 | 4 | 48 | 1 | 69 |
|  | Hatchery | 0 | 2 | 0 | 0 | 10 | 9 | 25 | 0 | 46 |
| $1994{ }^{\text {b }}$ | Wild | 0 | 0 | 1 | 0 | 7 | 1 | 113 | 22 | 144 |
|  | Hatchery | 0 | 4 | 3 | 0 | 20 | 4 | 205 | 38 | 274 |
| 1995 | Wild | 0 | 0 | 1 | 0 | 10 | 0 | 66 | 4 | 81 |
|  | Hatchery | 0 | 0 | 1 | 0 | 20 | 0 | 173 | 11 | 205 |
| 1996 | Wild | 0 | 0 | 0 | 1 | 3 | 1 | 53 | 0 | 58 |
|  | Hatchery | 0 | 0 | 0 | 1 | 2 | 1 | 173 | 0 | 177 |
| 1997 | Wild | 0 | 0 | 1 | 0 | 0 | 3 | 83 | 0 | 87 |
|  | Hatchery | 0 | 0 | 1 | 0 | 9 | 0 | 142 | 1 | 153 |
| 1998 | Wild | 0 | 1 | 3 | 1 | 6 | 5 | 162 | 4 | 182 |
|  | Hatchery | 0 | 0 | 5 | 0 | 1 | 2 | 178 | 0 | 186 |
| 1999 | Wild | 0 | 0 | 0 | 0 | 9 | 23 | 293 | 9 | 334 |
|  | Hatchery | 0 | 0 | 3 | 2 | 14 | 30 | 473 | 39 | 561 |
| 2000 | Wild | 0 | 0 | 8 | 8 | 24 | 11 | 189 | 4 | 244 |
|  | Hatchery | 0 | 2 | 12 | 7 | 23 | 5 | 538 | 37 | 624 |
| 2001 | Wild | 0 | 10 | 23 | 5 | 67 | 42 | 390 | 54 | 591 |
|  | Hatchery | 0 | 16 | 52 | 5 | 60 | 70 | 751 | 51 | $\begin{gathered} 1,00 \\ 5 \end{gathered}$ |
| 2002 | Wild | 6 | 14 | 20 | 10 | 81 | 212 | 340 | 72 | 755 |
|  | Hatchery | 4 | 18 | 63 | 25 | 123 | 360 | 925 | 187 | $\begin{gathered} 1,70 \\ 5 \end{gathered}$ |
| $2003{ }^{\text {c }}$ | Wild | 0 | 0 | 13 | 0 | 12 | 152 | 231 | 124 | 532 |
|  | Hatchery | 0 | 0 | 15 | 0 | 5 | 91 | 365 | 257 | 733 |
| 2004 | Wild | 0 | 2 | 19 | 19 | 108 | 225 | 1,125 | 260 | $\begin{gathered} 1,75 \\ 8 \end{gathered}$ |
|  | Hatchery | 0 | 2 | 12 | 5 | 38 | 58 | 267 | 38 | 420 |
| 2005 | Wild | 0 | 5 | 51 | 21 | 256 | 364 | 531 | 176 | $\begin{gathered} 1,40 \\ 4 \end{gathered}$ |
|  | Hatchery | 0 | 3 | 42 | 16 | 115 | 70 | 200 | 100 | 546 |
| 2006 | Wild | 2 | 2 | 22 | 10 | 105 | 247 | 370 | 73 | 831 |
|  | Hatchery | 2 | 1 | 9 | 6 | 15 | 44 | 138 | 33 | 248 |


| 2007 | Wild | 1 | 0 | 30 | 1 | 284 | 322 | 405 | 20 | 1,06 3 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Hatchery | 1 | 0 | 25 | 0 | 169 | 197 | 253 | 9 | 654 |
| 2008 | Wild | 2 | 1 | 14 | 11 | 107 | 324 | 347 | 41 | 847 |
|  | Hatchery | 2 | 9 | 26 | 25 | 141 | 341 | 512 | 116 | $\begin{gathered} 1,17 \\ 2 \end{gathered}$ |
| 2009 | Wild | 2 | 3 | 13 | 14 | 189 | 347 | 330 | 75 | 973 |
|  | Hatchery | 0 | 4 | 18 | 18 | 159 | 153 | 373 | 75 | 800 |
| 2010 | Wild | 1 | 5 | 19 | 18 | 154 | 180 | 329 | 69 | 775 |
|  | Hatchery | 2 | 5 | 11 | 24 | 87 | 172 | 296 | 79 | 676 |
| 2011 | Wild | 0 | 0 | 21 | 4 | 201 | 362 | 216 | 19 | 823 |
|  | Hatchery | 0 | 0 | 34 | 10 | 160 | 116 | 537 | 95 | 952 |
| 2012 | Wild | 0 | 0 | 18 | 9 | 133 | 427 | 206 | 23 | 816 |
|  | Hatchery | 1 | 0 | 38 | 6 | 123 | 110 | 288 | 31 | 597 |
| $2013{ }^{\text {d,e }}$ | Wild | 0 | 0 | 22 | 7 | 37 | 352 | 191 | 4 | 613 |
|  | Hatchery | 0 | 0 | 8 | 2 | 15 | 80 | 188 | 4 | 297 |
| 2014 | Wild | 0 | 1 | 60 | 47 | 233 | 716 | 641 | 425 | 2123 |
|  | Hatchery | 1 | 0 | 19 | 7 | 42 | 67 | 129 | 64 | 329 |
| 2015 | Wild | 0 | 5 | 39 | 9 | 209 | 931 | 1186 | 176 | 2555 |
|  | Hatchery | 0 | 5 | 22 | 2 | 74 | 63 | 516 | 56 | 738 |
| 2016 | Wild | 0 | 6 | 13 | 7 | 186 | 1019 | 819 | 121 | 2171 |
|  | Hatchery | 0 | 6 | 1 | 4 | 44 | 56 | 395 | 78 | 584 |
| Average | Wild | 1 | 2 | 18 | 9 | 105 | 273 | 374 | 77 | 859 |
|  | Hatchery | 1 | 3 | 18 | 7 | 63 | 91 | 342 | 60 | 586 |

${ }^{\text {a }} 25$ additional carcasses were sampled on the Similkameen and 46 on the Okanogan without any reach designation.
${ }^{\mathrm{b}}$ One additional carcass was sampled on the Similkameen without any reach designation.
${ }^{\text {c }} 793$ carcasses were sampled on the Similkameen before initiation of spawning (pre-spawn mortality) and an additional 40 carcasses were sampled on the Okanogan. The cause of the high mortality (Ichthyophthirius multifilis and Flavobacterium columnarae) was exacerbated by high river temperatures.
${ }^{\mathrm{d}}$ In 2013, carcass recoveries were combined in reaches O-3 and O-4, and S-1 and S-2. Then re-apportioned based on redd counts within each reach.
e 2013 data have been updated to reflect age and origin data acquired from scale reading since the publication of the 2013 annual report

## Age at Maturity

Table C 3 Salt age of recovered carcasses in the Okanogan and Similkameen Rivers.

| Hatchery-Origin Male <br> Salt Age Carcasses Recovered |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Survey Year | 0 | 1 | 2 | 3 | 4 | 5 | Total |
| 1993 | 0 | 0 | 33 | 0 | 0 | 0 | 33 |
| 1994 | 0 | 5 | 23 | 92 | 0 | 0 | 120 |
| 1995 | 0 | 2 | 23 | 27 | 17 | 0 | 69 |
| 1996 | 0 | 3 | 17 | 24 | 5 | 0 | 49 |
| 1997 | 0 | 0 | 1 | 25 | 2 | 0 | 28 |
| 1998 | 0 | 9 | 64 | 12 | 9 | 0 | 94 |
| 1999 | 2 | 0 | 35 | 74 | 2 | 0 | 113 |
| 2000 | 7 | 65 | 6 | 104 | 8 | 0 | 190 |
| 2001 | 0 | 47 | 625 | 3 | 11 | 0 | 686 |
| 2002 | 0 | 10 | 267 | 419 | 0 | 1 | 697 |
| 2003 | 0 | 18 | 30 | 146 | 27 | 0 | 221 |
| 2004 | 0 | 2 | 100 | 67 | 18 | 0 | 187 |
| 2005 | 0 | 12 | 19 | 104 | 15 | 0 | 150 |
| 2006 | 0 | 7 | 15 | 11 | 27 | 0 | 60 |
| 2007 | 0 | 122 | 116 | 56 | 5 | 3 | 302 |
| 2008 | 0 | 18 | 460 | 137 | 3 | 0 | 618 |
| 2009 | 0 | 43 | 33 | 158 | 2 | 0 | 236 |
| 2010 | 4 | 20 | 293 | 29 | 7 | 0 | 353 |
| 2011 | 0 | 144 | 47 | 118 | 0 | 0 | 309 |
| 2012 | 1 | 31 | 168 | 63 | 7 | 0 | 270 |
| 2013 | 0 | 7 | 27 | 22 | 2 | 1 | 59 |
| 2014 | 0 | 55 | 58 | 39 | 0 | 0 | 152 |
| 2015 | 0 | 17 | 234 | 49 | 0 | 0 | 300 |
| 2016 | 0 | 6 | 15 | 74 | 4 | 0 | 99 |
| Average | 1 | 27 | 113 | 77 | 7 | 0 | 225 |

Hatchery-Origin Female
Salt Age Carcasses Recovered

| Survey Year | $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\mathbf{1 9 9 3}$ | 0 | 0 | 10 | 1 | 0 | 0 | 11 |
| $\mathbf{1 9 9 4}$ | 0 | 0 | 3 | 141 | 1 | 0 | 145 |
| $\mathbf{1 9 9 5}$ | 0 | 0 | 9 | 44 | 82 | 0 | 135 |
| $\mathbf{1 9 9 6}$ | 0 | 0 | 21 | 74 | 31 | 1 | 127 |
| $\mathbf{1 9 9 7}$ | 0 | 0 | 2 | 107 | 16 | 0 | 125 |
| $\mathbf{1 9 9 8}$ | 0 | 1 | 28 | 30 | 32 | 0 | 91 |
| $\mathbf{1 9 9 9}$ | 1 | 0 | 31 | 393 | 13 | 2 | 440 |
| $\mathbf{2 0 0 0}$ | 0 | 1 | 4 | 307 | 49 | 0 | 361 |
| $\mathbf{2 0 0 1}$ | 0 | 1 | 256 | 19 | 42 | 0 | 318 |
| $\mathbf{2 0 0 2}$ | 0 | 0 | 54 | 921 | 9 | 0 | 984 |
| $\mathbf{2 0 0 3}$ | 0 | 1 | 9 | 368 | 54 | 0 | 432 |
| $\mathbf{2 0 0 4}$ | 0 | 0 | 22 | 103 | 69 | 0 | 194 |
| $\mathbf{2 0 0 5}$ | 0 | 0 | 11 | 303 | 64 | 2 | 380 |
| $\mathbf{2 0 0 6}$ | 0 | 0 | 10 | 21 | 48 | 0 | 79 |
| $\mathbf{2 0 0 7}$ | 0 | 0 | 53 | 178 | 22 | 4 | 257 |
| $\mathbf{2 0 0 8}$ | 0 | 0 | 197 | 267 | 25 | 1 | 490 |
| $\mathbf{2 0 0 9}$ | 0 | 0 | 9 | 516 | 22 | 0 | 547 |
| $\mathbf{2 0 1 0}$ | 0 | 0 | 155 | 120 | 42 | 1 | 318 |
| $\mathbf{2 0 1 1}$ | 0 | 1 | 22 | 602 | 6 | 0 | 631 |
| $\mathbf{2 0 1 2}$ | 0 | 1 | 153 | 140 | 25 | 0 | 319 |
| $\mathbf{2 0 1 3}$ | 1 | 0 | 34 | 188 | 7 | 0 | 230 |
| $\mathbf{2 0 1 4}$ | 0 | 0 | 23 | 127 | 5 | 0 | 155 |
| $\mathbf{2 0 1 5}$ | 0 | 1 | 138 | 102 | 5 | 0 | 246 |
| $\mathbf{2 0 1 6}$ | 0 | 0 | 6 | 283 | 13 | 0 | 302 |
| Average | $\boldsymbol{0}$ | $\boldsymbol{0}$ | $\mathbf{5 3}$ | $\mathbf{2 2 3}$ | $\mathbf{2 8}$ | $\boldsymbol{0}$ | $\mathbf{3} 305$ |


| Natural-Origin Male <br> Salt Age Carcasses Recovered |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Survey <br> Year | 0 | 1 | 2 | 3 | 4 | 5 | Total |
| 1993 | 0 | 0 | 8 | 19 | 3 | 0 | 30 |
| 1994 | 0 | 3 | 13 | 22 | 10 | 0 | 48 |
| 1995 | 0 | 0 | 6 | 11 | 4 | 0 | 21 |
| 1996 | 0 | 1 | 7 | 4 | 1 | 0 | 13 |
| 1997 | 0 | 3 | 8 | 8 | 1 | 0 | 20 |
| 1998 | 0 | 3 | 32 | 27 | 5 | 0 | 67 |
| 1999 | 0 | 0 | 22 | 39 | 8 | 1 | 70 |
| 2000 | 0 | 6 | 24 | 27 | 12 | 0 | 69 |
| 2001 | 0 | 13 | 82 | 168 | 8 | 0 | 271 |
| 2002 | 0 | 15 | 85 | 232 | 52 | 1 | 385 |
| 2003 | 0 | 12 | 55 | 171 | 34 | 0 | 272 |
| 2004 | 0 | 19 | 226 | 166 | 303 | 3 | 717 |
| 2005 | 0 | 1 | 129 | 447 | 28 | 4 | 609 |
| 2006 | 0 | 1 | 14 | 189 | 116 | 0 | 320 |
| 2007 | 0 | 17 | 67 | 53 | 226 | 5 | 368 |
| 2008 | 0 | 8 | 258 | 263 | 13 | 2 | 544 |
| 2009 | 0 | 10 | 21 | 276 | 31 | 0 | 338 |
| 2010 | 0 | 3 | 90 | 123 | 50 | 0 | 266 |
| 2011 | 0 | 10 | 46 | 228 | 17 | 0 | 301 |
| 2012 | 1 | 14 | 160 | 112 | 58 | 0 | 345 |
| 2013 | 0 | 6 | 83 | 140 | 12 | 0 | 241 |
| 2014 | 0 | 43 | 135 | 633 | 76 | 0 | 887 |
| 2015 | 0 | 8 | 809 | 402 | 113 | 0 | 1332 |
| 2016 | 0 | 1 | 53 | 548 | 109 | 1 | 712 |
| Average | 0 | 8 | 101 | 180 | 54 | 1 | 344 |


| Salt Age Carcasses Recovered |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Survey <br> Year | $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | Total |  |
| $\mathbf{1 9 9 3}$ | 0 | 0 | 5 | 25 | 3 | 0 | 33 |  |
| $\mathbf{1 9 9 4}$ | 0 | 0 | 2 | 36 | 29 | 0 | 67 |  |
| $\mathbf{1 9 9 5}$ | 0 | 0 | 7 | 27 | 11 | 0 | 45 |  |
| $\mathbf{1 9 9 6}$ | 0 | 0 | 3 | 18 | 2 | 0 | 23 |  |
| $\mathbf{1 9 9 7}$ | 0 | 0 | 12 | 31 | 10 | 0 | 53 |  |
| $\mathbf{1 9 9 8}$ | 0 | 0 | 21 | 51 | 12 | 0 | 84 |  |
| $\mathbf{1 9 9 9}$ | 0 | 0 | 32 | 132 | 34 | 0 | 198 |  |
| $\mathbf{2 0 0 0}$ | 0 | 0 | 9 | 106 | 32 | 0 | 147 |  |
| $\mathbf{2 0 0 1}$ | 0 | 0 | 11 | 237 | 12 | 0 | 260 |  |
| $\mathbf{2 0 0 2}$ | 0 | 0 | 18 | 199 | 90 | 0 | 307 |  |
| $\mathbf{2 0 0 3}$ | 2 | 2 | 29 | 130 | 45 | 0 | 208 |  |
| $\mathbf{2 0 0 4}$ | 0 | 0 | 37 | 233 | 539 | 2 | 811 |  |
| $\mathbf{2 0 0 5}$ | 0 | 0 | 28 | 566 | 71 | 7 | 672 |  |
| $\mathbf{2 0 0 6}$ | 0 | 0 | 2 | 250 | 256 | 2 | 510 |  |
| $\mathbf{2 0 0 7}$ | 0 | 0 | 8 | 72 | 601 | 12 | 693 |  |
| $\mathbf{2 0 0 8}$ | 0 | 0 | 12 | 269 | 19 | 3 | 303 |  |
| $\mathbf{2 0 0 9}$ | 0 | 0 | 3 | 473 | 112 | 0 | 588 |  |
| $\mathbf{2 0 1 0}$ | 0 | 0 | 20 | 195 | 226 | 1 | 442 |  |
| $\mathbf{2 0 1 1}$ | 0 | 0 | 12 | 416 | 58 | 0 | 486 |  |
| $\mathbf{2 0 1 2}$ | 0 | 0 | 15 | 195 | 196 | 0 | 406 |  |
| $\mathbf{2 0 1 3}$ | 0 | 0 | 5 | 254 | 27 | 0 | 286 |  |
| $\mathbf{2 0 1 4}$ | 0 | 3 | 24 | 809 | 189 | 0 | 1025 |  |
| $\mathbf{2 0 1 5}$ | 0 | 0 | 66 | 342 | 426 | 1 | 835 |  |
| $\mathbf{2 0 1 6}$ | 0 | 0 | 4 | 927 | 288 | 4 | 1223 |  |
| Average | $\boldsymbol{0}$ | $\boldsymbol{0}$ | $\mathbf{1 6}$ | $\mathbf{2 5 0}$ | $\mathbf{1 3 7}$ | $\mathbf{1}$ | 404 |  |
|  |  |  |  |  |  |  |  |  |

Table C 4 Salt age structure (percent of recovered carcasses) for sex-origin classes.

## Hatchery-Origin Male

Salt Age - Percent of carcasses recovered within origin/sex class

| Survey <br> Year | $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\mathbf{1 9 9 3}$ | $0 \%$ | $0 \%$ | $100 \%$ | $0 \%$ | $0 \%$ | $0 \%$ | 1 |
| $\mathbf{1 9 9 4}$ | $0 \%$ | $4 \%$ | $19 \%$ | $77 \%$ | $0 \%$ | $0 \%$ | 1 |
| $\mathbf{1 9 9 5}$ | $0 \%$ | $3 \%$ | $33 \%$ | $39 \%$ | $25 \%$ | $0 \%$ | 1 |
| $\mathbf{1 9 9 6}$ | $0 \%$ | $6 \%$ | $35 \%$ | $49 \%$ | $10 \%$ | $0 \%$ | 1 |
| $\mathbf{1 9 9 7}$ | $0 \%$ | $0 \%$ | $4 \%$ | $89 \%$ | $7 \%$ | $0 \%$ | 1 |
| $\mathbf{1 9 9 8}$ | $0 \%$ | $10 \%$ | $68 \%$ | $13 \%$ | $10 \%$ | $0 \%$ | 1 |
| $\mathbf{1 9 9 9}$ | $2 \%$ | $0 \%$ | $31 \%$ | $65 \%$ | $2 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 0 0}$ | $4 \%$ | $34 \%$ | $3 \%$ | $55 \%$ | $4 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 0 1}$ | $0 \%$ | $7 \%$ | $91 \%$ | $0 \%$ | $2 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 0 2}$ | $0 \%$ | $1 \%$ | $38 \%$ | $60 \%$ | $0 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 0 3}$ | $0 \%$ | $8 \%$ | $14 \%$ | $66 \%$ | $12 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 0 4}$ | $0 \%$ | $1 \%$ | $53 \%$ | $36 \%$ | $10 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 0 5}$ | $0 \%$ | $8 \%$ | $13 \%$ | $69 \%$ | $10 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 0 6}$ | $0 \%$ | $12 \%$ | $25 \%$ | $18 \%$ | $45 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 0 7}$ | $0 \%$ | $40 \%$ | $38 \%$ | $19 \%$ | $2 \%$ | $1 \%$ | 1 |
| $\mathbf{2 0 0 8}$ | $0 \%$ | $3 \%$ | $74 \%$ | $22 \%$ | $0 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 0 9}$ | $0 \%$ | $18 \%$ | $14 \%$ | $67 \%$ | $1 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 1 0}$ | $1 \%$ | $6 \%$ | $83 \%$ | $8 \%$ | $2 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 1 1}$ | $0 \%$ | $47 \%$ | $15 \%$ | $38 \%$ | $0 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 1 2}$ | $0 \%$ | $11 \%$ | $62 \%$ | $23 \%$ | $3 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 1 3}$ | $0 \%$ | $12 \%$ | $46 \%$ | $37 \%$ | $3 \%$ | $2 \%$ | 1 |
| $\mathbf{2 0 1 4}$ | $0 \%$ | $36 \%$ | $38 \%$ | $26 \%$ | $0 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 1 5}$ | $0 \%$ | $6 \%$ | $78 \%$ | $16 \%$ | $0 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 1 6}$ | $0 \%$ | $6 \%$ | $15 \%$ | $75 \%$ | $4 \%$ | $0 \%$ | 1 |
| Average | $\mathbf{0 \%}$ | $\mathbf{1 2 \%}$ | $\mathbf{4 1 \%}$ | $\mathbf{4 0 \%}$ | $\mathbf{6 \%}$ | $\boldsymbol{0} \%$ | $\mathbf{1}$ |

Hatchery-Origin Female
Salt Age - Percent of carcasses recovered within origin/sex class

| Survey Year | $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\mathbf{1 9 9 3}$ | $0 \%$ | $0 \%$ | $91 \%$ | $9 \%$ | $0 \%$ | $0 \%$ | 1 |
| $\mathbf{1 9 9 4}$ | $0 \%$ | $0 \%$ | $2 \%$ | $97 \%$ | $1 \%$ | $0 \%$ | 1 |
| $\mathbf{1 9 9 5}$ | $0 \%$ | $0 \%$ | $7 \%$ | $33 \%$ | $61 \%$ | $0 \%$ | 1 |
| $\mathbf{1 9 9 6}$ | $0 \%$ | $0 \%$ | $17 \%$ | $58 \%$ | $24 \%$ | $1 \%$ | 1 |
| $\mathbf{1 9 9 7}$ | $0 \%$ | $0 \%$ | $2 \%$ | $86 \%$ | $13 \%$ | $0 \%$ | 1 |
| $\mathbf{1 9 9 8}$ | $0 \%$ | $1 \%$ | $31 \%$ | $33 \%$ | $35 \%$ | $0 \%$ | 1 |
| $\mathbf{1 9 9 9}$ | $0 \%$ | $0 \%$ | $7 \%$ | $89 \%$ | $3 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 0 0}$ | $0 \%$ | $0 \%$ | $1 \%$ | $85 \%$ | $14 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 0 1}$ | $0 \%$ | $0 \%$ | $81 \%$ | $6 \%$ | $13 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 0 2}$ | $0 \%$ | $0 \%$ | $5 \%$ | $94 \%$ | $1 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 0 3}$ | $0 \%$ | $0 \%$ | $2 \%$ | $85 \%$ | $13 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 0 4}$ | $0 \%$ | $0 \%$ | $11 \%$ | $53 \%$ | $36 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 0 5}$ | $0 \%$ | $0 \%$ | $3 \%$ | $80 \%$ | $17 \%$ | $1 \%$ | 1 |
| $\mathbf{2 0 0 6}$ | $0 \%$ | $0 \%$ | $13 \%$ | $27 \%$ | $61 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 0 7}$ | $0 \%$ | $0 \%$ | $21 \%$ | $69 \%$ | $9 \%$ | $2 \%$ | 1 |
| $\mathbf{2 0 0 8}$ | $0 \%$ | $0 \%$ | $40 \%$ | $54 \%$ | $5 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 0 9}$ | $0 \%$ | $0 \%$ | $2 \%$ | $94 \%$ | $4 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 1 0}$ | $0 \%$ | $0 \%$ | $49 \%$ | $38 \%$ | $13 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 1 1}$ | $0 \%$ | $0 \%$ | $3 \%$ | $95 \%$ | $1 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 1 2}$ | $0 \%$ | $0 \%$ | $48 \%$ | $44 \%$ | $8 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 1 3}$ | $0 \%$ | $0 \%$ | $15 \%$ | $82 \%$ | $3 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 1 4}$ | $0 \%$ | $0 \%$ | $15 \%$ | $82 \%$ | $3 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 1 5}$ | $0 \%$ | $0 \%$ | $56 \%$ | $41 \%$ | $2 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 1 6}$ | $0 \%$ | $0 \%$ | $2 \%$ | $94 \%$ | $4 \%$ | $0 \%$ | 1 |
| Average | $0 \%$ | $0 \%$ | $22 \%$ | $64 \%$ | $14 \%$ | $0 \%$ | 1 |

Natural-Origin Male
Salt Age - Percent of carcasses recovered within origin/sex class

| Survey Year | $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\mathbf{1 9 9 3}$ | $0 \%$ | $0 \%$ | $27 \%$ | $63 \%$ | $10 \%$ | $0 \%$ | 1 |
| $\mathbf{1 9 9 4}$ | $0 \%$ | $6 \%$ | $27 \%$ | $46 \%$ | $21 \%$ | $0 \%$ | 1 |
| $\mathbf{1 9 9 5}$ | $0 \%$ | $0 \%$ | $29 \%$ | $52 \%$ | $19 \%$ | $0 \%$ | 1 |
| $\mathbf{1 9 9 6}$ | $0 \%$ | $8 \%$ | $54 \%$ | $31 \%$ | $8 \%$ | $0 \%$ | 1 |
| $\mathbf{1 9 9 7}$ | $0 \%$ | $15 \%$ | $40 \%$ | $40 \%$ | $5 \%$ | $0 \%$ | 1 |
| $\mathbf{1 9 9 8}$ | $0 \%$ | $4 \%$ | $48 \%$ | $40 \%$ | $7 \%$ | $0 \%$ | 1 |
| $\mathbf{1 9 9 9}$ | $0 \%$ | $0 \%$ | $31 \%$ | $56 \%$ | $11 \%$ | $1 \%$ | 1 |
| $\mathbf{2 0 0 0}$ | $0 \%$ | $9 \%$ | $35 \%$ | $39 \%$ | $17 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 0 1}$ | $0 \%$ | $5 \%$ | $30 \%$ | $62 \%$ | $3 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 0 2}$ | $0 \%$ | $4 \%$ | $22 \%$ | $60 \%$ | $14 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 0 3}$ | $0 \%$ | $4 \%$ | $20 \%$ | $63 \%$ | $13 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 0 4}$ | $0 \%$ | $3 \%$ | $32 \%$ | $23 \%$ | $42 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 0 5}$ | $0 \%$ | $0 \%$ | $21 \%$ | $73 \%$ | $5 \%$ | $1 \%$ | 1 |
| $\mathbf{2 0 0 6}$ | $0 \%$ | $0 \%$ | $4 \%$ | $59 \%$ | $36 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 0 7}$ | $0 \%$ | $5 \%$ | $18 \%$ | $14 \%$ | $61 \%$ | $1 \%$ | 1 |
| $\mathbf{2 0 0 8}$ | $0 \%$ | $1 \%$ | $47 \%$ | $48 \%$ | $2 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 0 9}$ | $0 \%$ | $3 \%$ | $6 \%$ | $82 \%$ | $9 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 1 0}$ | $0 \%$ | $1 \%$ | $34 \%$ | $46 \%$ | $19 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 1 1}$ | $0 \%$ | $3 \%$ | $15 \%$ | $76 \%$ | $6 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 1 2}$ | $0 \%$ | $4 \%$ | $46 \%$ | $32 \%$ | $17 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 1 3}$ | $0 \%$ | $2 \%$ | $34 \%$ | $58 \%$ | $5 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 1 4}$ | $0 \%$ | $5 \%$ | $15 \%$ | $71 \%$ | $9 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 1 5}$ | $0 \%$ | $1 \%$ | $61 \%$ | $30 \%$ | $8 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 1 6}$ | $0 \%$ | $7 \%$ | $77 \%$ | $15 \%$ | $0 \%$ | $0 \%$ | 1 |
| Average | $0 \%$ | $4 \%$ | $32 \%$ | $49 \%$ | $14 \%$ | $0 \%$ | 1 |

Natural-Origin Female
Salt Age - Percent of carcasses recovered within origin/sex class

| Sample Year | $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\mathbf{1 9 9 3}$ | $0 \%$ | $0 \%$ | $15 \%$ | $76 \%$ | $9 \%$ | $0 \%$ | 1 |
| $\mathbf{1 9 9 4}$ | $0 \%$ | $0 \%$ | $3 \%$ | $54 \%$ | $43 \%$ | $0 \%$ | 1 |
| $\mathbf{1 9 9 5}$ | $0 \%$ | $0 \%$ | $16 \%$ | $60 \%$ | $24 \%$ | $0 \%$ | 1 |
| $\mathbf{1 9 9 6}$ | $0 \%$ | $0 \%$ | $13 \%$ | $78 \%$ | $9 \%$ | $0 \%$ | 1 |
| $\mathbf{1 9 9 7}$ | $0 \%$ | $0 \%$ | $23 \%$ | $58 \%$ | $19 \%$ | $0 \%$ | 1 |
| $\mathbf{1 9 9 8}$ | $0 \%$ | $0 \%$ | $25 \%$ | $61 \%$ | $14 \%$ | $0 \%$ | 1 |
| $\mathbf{1 9 9 9}$ | $0 \%$ | $0 \%$ | $16 \%$ | $67 \%$ | $17 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 0 0}$ | $0 \%$ | $0 \%$ | $6 \%$ | $72 \%$ | $22 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 0 1}$ | $0 \%$ | $0 \%$ | $4 \%$ | $91 \%$ | $5 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 0 2}$ | $0 \%$ | $0 \%$ | $6 \%$ | $65 \%$ | $29 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 0 3}$ | $1 \%$ | $1 \%$ | $14 \%$ | $63 \%$ | $22 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 0 4}$ | $0 \%$ | $0 \%$ | $5 \%$ | $29 \%$ | $66 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 0 5}$ | $0 \%$ | $0 \%$ | $4 \%$ | $84 \%$ | $11 \%$ | $1 \%$ | 1 |
| $\mathbf{2 0 0 6}$ | $0 \%$ | $0 \%$ | $0 \%$ | $49 \%$ | $50 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 0 7}$ | $0 \%$ | $0 \%$ | $1 \%$ | $10 \%$ | $87 \%$ | $2 \%$ | 1 |
| $\mathbf{2 0 0 8}$ | $0 \%$ | $0 \%$ | $4 \%$ | $89 \%$ | $6 \%$ | $1 \%$ | 1 |
| $\mathbf{2 0 0 9}$ | $0 \%$ | $0 \%$ | $1 \%$ | $80 \%$ | $19 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 1 0}$ | $0 \%$ | $0 \%$ | $5 \%$ | $44 \%$ | $51 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 1 1}$ | $0 \%$ | $0 \%$ | $2 \%$ | $86 \%$ | $12 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 1 2}$ | $0 \%$ | $0 \%$ | $4 \%$ | $48 \%$ | $48 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 1 3}$ | $0 \%$ | $0 \%$ | $2 \%$ | $89 \%$ | $9 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 1 4}$ | $0 \%$ | $0 \%$ | $2 \%$ | $79 \%$ | $18 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 1 5}$ | $0 \%$ | $0 \%$ | $8 \%$ | $41 \%$ | $51 \%$ | $0 \%$ | 1 |
| $\mathbf{2 0 1 6}$ | $0 \%$ | $0 \%$ | $0 \%$ | $76 \%$ | $24 \%$ | $0 \%$ | 1 |
| Average | $\mathbf{0 \%}$ | $\mathbf{0 \%}$ | $\mathbf{7 \%}$ | $\mathbf{6 5 \%}$ | $\mathbf{2 8 \%}$ | $\mathbf{0} \%$ | $\mathbf{1}$ |

## Contribution to Fisheries

Table C 5. Estimated number and percent of hatchery-origin Okanogan/Similkameen summer Chinook captured in different fisheries, brood years 1989-2010.

| Brood year | Ocean fisheries | Columbia River Fisheries |  |  | Total | Percent of brood year escapement harvested ${ }^{\text {a }}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Tribal | Commercial (Zones 1-5) | Recreational (sport) |  |  |
| 1989 | 2,360 (80) | 553 (19) | 0 (0) | 53 (2) | 2,966 | 39.8 |
| 1990 | 355 (89) | 34 (8) | 0 (0) | 12 (3) | 401 | 28.2 |
| 1991 | 220 (86) | 37 (14) | 0 (0) | 0 (0) | 257 | 14.0 |
| 1992 | 422 (91) | 28 (6) | 2 (0) | 10 (2) | 462 | 20.0 |
| 1993 | 24 (80) | 6 (20) | 0 (0) | 0 (0) | 30 | 25.6 |
| 1994 | 372 (92) | 23 (6) | 2 (0) | 7 (2) | 404 | 26.1 |
| 1995 | 643 (93) | 9 (1) | 12 (2) | 25 (4) | 689 | 23.8 |
| 1996 | 6 (100) | 0 (0) | 0 (0) | 0 (0) | 6 | 18.2 |
| 1997 | 6,483 (92) | 136 (2) | 36 (1) | 424 (6) | 7,079 | 37.1 |
| 1998 | 4,414 (89) | 251 (5) | 45 (1) | 223 (5) | 4,933 | 62.8 |
| 1999 | 1,359 (68) | 224 (11) | 31 (2) | 384 (19) | 1,998 | 70.0 |
| 2000 | 3,139 (69) | 533 (12) | 222 (5) | 675 (15) | 4,559 | 67.1 |
| 2001 | 184 (58) | 81 (25) | 31 (10) | 23 (7) | 319 | 74.9 |
| 2002 | 706 (56) | 200 (16) | 90 (7) | 258 (21) | 1,254 | 63.2 |
| 2003 | 711 (38) | 568 (30) | 130 (7) | 466 (25) | 1,875 | 53.3 |
| 2004 | 3,153 (39) | 2,162 (26) | 694 (8) | 2,168 (27) | 8,177 | 60.9 |
| 2005 | 470 (46) | 306 (30) | 79 (8) | 167 (16) | 1,022 | 61.1 |
| 2006 | 3,136 (37) | 3,352 (40) | 469 (6) | 1,419 (17) | 8,376 | 61.0 |
| 2007 | 1,549 (44) | 992 (28) | 67 (2) | 905 (26) | 3,513 | 70.8 |
| 2008 | 4,226 (38) | 2,576 (23) | 218 (2) | 3,969 (36) | 10,989 | 73.5 |
| 2009 | 2,005 (36) | 2,155 (39) | 207 (4) | 1,138 (21) | 5,505 | 77.2 |
| 2010 | 3,193 (38) | 3,933 (46) | 247 (3) | 1,110 (13) | 8,483 | 79.0 |
| 2011 | 5,801 (40) | 5,812 (40) | 456 (3) | 2,598 (18) | 14,667 | 78.0 |
| Average | 1,953 (51) | 1,042 (27) | 132 (3) | 697 (18) | 3,825 | 51.5 |
| Median | 1,359 (68) | 251 (19) | 45 (2) | 258 (13) | 1,998 | 61.0 |

a Percent of brood year escapement harvested = Total brood year harvest/ (Total brood year harvest $+\Sigma$ Hatchery collection $+\Sigma$ escapement) * 100. In other words, this indicates the percentage of all detected CWTs that ended up in harvest.

## APPENDIX D

## Glossary of Terms, Acronyms, and Abbreviations

The following is a list of key terms and variables used in the Chief Joseph Hatchery Program and in this Annual Report. This is not a complete list, but provides many of the main terms used in this report or that will likely be used in future CJHP Annual Report.

Accord/MOA = A ten-year agreement (2008-2018) between BPA and the CCT whereas BPA agreed to fund pre-determined fish and wildlife projects and CCT agreed not to sue the Action Agencies regarding the BiOp for the FCRPS.

CJHP Master Plan = A three-step development and review process required for all new hatcheries funded by BPA in the Columbia Basin.
eDNA = environmental DNA; dissolved or cell-bound DNA that persists in the environment.
Escapement Target $=$ Number of fish of all origins targeted to pass upstream of the Okanogan Adult Fish weir

HOB = the number of hatchery-origin fish used as hatchery broodstock.
HOR = hatchery-origin recruit. The number of HORs equals the sum of HOS + HOB + hatchery-origin fish intercepted in fisheries.

HOR Terminal Run Size = Number of Chief Joseph Hatchery HORs returning to Wells Dam HOS = the number of hatchery-origin fish spawning naturally.

Juvenile Abundance = annual abundance of out-migrant juveniles estimated by expanding data from juveniles captured at the rotary screw trap.

Met Comp = Methow composite Spring Chinook. These fish are part of the Winthrop NFH program and are intended to be used for the Okanogan reintroduction pending approval under section 10 (j) of the ESA.

NOB = the number of natural-origin fish used as hatchery broodstock.
NOR = natural-origin recruit. The number of NOR's equals the sum of NOB, + NOS + natural-origin fish intercepted in fisheries.

NOR Terminal Run Size = Number of Okanogan (and Similkameen, combined) NOR's returning to Wells Dam.

NOS = the number of natural-origin fish spawning naturally.
$\mathbf{p H O S}=$ proportion of natural spawners composed of HORs. Equals HOS/ (NOS + HOS).

PNI = proportion of natural influence on a composite hatchery-/natural-origin population. Can also be thought of as the percentage of time the genes of a composite population spend in the natural environment. Equals 1 - pNOB/ (pNOB + pHOS).
$\mathbf{p N O B}=$ proportion of hatchery broodstock composed of NORs. Equals NOB/ (HOB + NOB).
$\mathbf{S A R}=$ smolt to adult return.
Recovery Plans = Federally-required plans under the Endangered Species Act that describe species status, recovery criteria and expected restoration actions.

Relative Reproductive Success = The probability that an HOR produce adult offspring and summer/fall expressed as a fraction of the same probability for a NOR

Spatial Distribution = Geographic spawning distribution of adult salmon.
Spawner Abundance $=$ Total number of adult spawners each year.
Subbasin Plans = Plans developed in the early 2000s for the NPCC project funding process describing "limiting factors" used for development of regional recovery and protection strategies.

Total NOR Recruitment = Annual number of adult recruits (catch plus escapement)

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AHA = All H Analyzer
APPT = Annual Program Planning Tool
APR = Annual Program Review
BiOp = Biological Opinion
BKD = Bacterial Kidney Disease
BPA = Bonneville Power Administration
CA = Coordinated Assessments
CBFWA = Columbia Basin Fish and Wildlife Authority
CCT = Confederated Tribes of the Colville Indian Reservation
cfs = Cubic feet per second
CJH = Chief Joseph Hatchery
CJHP = Chief Joseph Hatchery Program
Colville Tribes = Confederated Tribes of the Colville Reservation
CTFWP = Colville Tribes Fish &Wildlife Program
CRITFC = Columbia River Inter-Tribal Fish Commission
CWT = Coded Wire Tag
DI = Density Index
```

DPS = Distinct Population Segment
EDT = Ecosystem Diagnostic \& Treatment
ELISA = Enzyme-Linked Immunosorbent Assay
ESA = Endangered Species Act
ESU = Evolutionarily Significant Unit
FCRPS = Federal Columbia River Power System
FI = Flow Index
FPP = Fish per pound
FWS = U.S. Fish and Wildlife Service
GIS = Geographic Information System
Gpm = gallons per minute
GPS = Global Positioning System
HCP = Habitat Conservation Plan(s)
HGMP = Hatchery Genetic Management Plan(s)
HSRG = Hatchery Science Review Group
ISIT = In-season Implementation Tool
ISRP = Independent Scientific Review Panel
KMQ = Key Management Questions
LNFH = Leavenworth National Fish Hatchery
NEPA = National Environmental Policy Act
NMFS = National Marine Fisheries Service
NOAA = National Oceanic and Atmospheric Administration
NPCC = Northwest Power and Conservation Council
OBMEP = Okanogan Basin Monitoring and Evaluation Program
ODFW = Oregon Department of Fish and Wildlife
ONA = Okanagan Nation Alliance
PBT = Parental Based Tagging
PIT = Passive Integrated Transponder
PNAMP = Pacific Northwest Aquatic Monitoring Partnership
PSMFC = Pacific States Marine Fisheries Commission
PTAGIS = PIT Tag Information System
PUD = Public Utility District
PL

RKM= River Kilometer
RM = River Mile
RMIS = Regional Mark Information System
RM\&E = Research, Monitoring, and Evaluation
RST = Rotary Screw Trap
SNP = Single Nucleotide Polymorphism
TAC $=$ Technical Advisory Committee
TRMP = Tribal Resources Management Plan
$\mathbf{T U}=$ Temperature Unit
UCSRB = Upper Columbia Salmon Recovery Board
USGS = U.S. Geological Survey
WDFW = Washington Department of Fish and Wildlife
WNFH = Winthrop National Fish Hatchery


[^0]:    ${ }^{1}$ Adapted from the Hatchery Reform Project, the Hatchery Science Review Group reports and independent science review.

[^1]:    ${ }^{2}$ website: http://www.psmfc.org/Regional Mark Processing Center RMPC

[^2]:    ${ }^{3}$ There could have been some hatchery-origin fish with an intact adipose fin. Although all summer/fall Chinook hatchery programs in the Upper Columbia strive for a $100 \%$ adipose fin clip rate, a small percentage ( $\sim 1 \%$ ) may not receive the fin clip due to mechanical failure in the marking trailer. Additionally, not all fall Chinook programs, such as Priest Rapids Hatchery, clip the adipose fin of their releases.

[^3]:    ${ }^{4}$ This $80 \%$ correction factor has also been suggested by the HSRG as a default value when no direct estimates are available (HSRG 2009). Also see HSRG 2014 for a discussion about the definition and calculation effective pHOS.

[^4]:    ${ }^{5}$ A radio tracking study showed that fewer than $50 \%$ of the natural-origin fish tagged at Wells Dam ended up in the Okanogan in 2011 and 2012 (Mann and Snow 2013).

[^5]:    ${ }^{6}$ Not all hatchery steelhead released in the Okanogan receive an adipose fin clip. In 2016, 20,349 steelhead were released into the Similkameen River with an adipose clip, and 451 unclipped steelhead were released. In Omak Creek, 32,763 unclipped steelhead were released and in Salmon Creek, adipose clipped steelhead and 6,477 unclipped steelhead were released.

[^6]:    * Spawning escapement was calculated using the "Modified Meekin Method" (i.e., $3.1 \times$ jack multiplier).

[^7]:    ${ }^{7}$ Origin assignments take into account all scale, ad-mark, coded wire tag and PIT tag information available at time of publication. Values may be updated in future annual reports depending on availability of data.

[^8]:    ${ }^{\text {a }}$ Includes all tag codes and CWT released fish (CWT + Ad Clip fish and CWT-only fish).
    ${ }^{\mathrm{b}}$ Includes estimated recoveries (spawning grounds, hatcheries, all harvest - including the ocean and Columbia river basin, etc.) and observed recoveries if estimated recoveries were unavailable.

[^9]:    ${ }^{8}$ From McElhany, 2000 (NOAA), a viable salmonid population is an independent population of any Pacific salmonid (genus Oncorhynchus) that has a negligible risk of extinction due to threats from demographic variation, local environmental variation, and genetic diversity changes over a 100-year time frame. The four VSP parameters are abundance, productivity, spatial structure and diversity.

[^10]:    ${ }^{9}$ A wandering fish is one that has bypassed its facility (or river) of origin and been intercepted by another facility. These fish are often categorized as a stray, but had the intercepting facility not retained the fish it may have fallen back and returned to its place of origin.

[^11]:    ${ }^{\text {a }}$ Releases from Out of ESU hatcheries include:, Parkdale and Nez Perce hatcheries

