# The Chief Joseph Hatchery Program Summer/Fall Chinook 

2020 Annual Report

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This report includes both hatchery production/operations and the corresponding monitoring activities completed through April of 2021. It is structured to meet the RM\&E technical report formatting requirements for BPA, and therefore the hatchery production portion is included in Appendix A.

Reports, program descriptions, annual review materials and background information, news and contact information can be found on our website at: https://www.cct-fnw.com/reports/.

All photos are credited to Confederated Tribes of the Colville Reservation Fish and Wildlife Department - Chief Joseph Hatchery Program unless otherwise noted.

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## Executive Summary

The Colville Confederated Tribes (CCT) Chief Joseph Hatchery (CJH) is the fourth hatchery obligated under the Grand Coulee Dam/Dry Falls project, originating in the 1940s. Leavenworth, Entiat, and Winthrop National Fish Hatcheries were built and operated as mitigation for salmon blockage at Grand Coulee Dam, but the fourth hatchery was not built, and the obligation was nearly forgotten. After the Colville Tribes successfully collaborated with the United States to resurrect the project, planning of the hatchery began in 2001 and construction was completed in 2013. The monitoring program began in 2012 and adult Chinook Salmon were brought on station for the first time in June 2013. Bonneville Power Administration (BPA) is the primary funding source for CJH , and the Mid-Columbia PUDs (Douglas, Grant, and Chelan County) have entered into cost-share agreements with the tribes and BPA in order to meet some of their mitigation obligations.

The CJH production level was set at $100 \%$ in 2020 during the eighth year of operation for the Summer/Fall Chinook program. In July and August the CCT used a purse seine vessel to collect 1,287 summer/fall Chinook for broodstock for both the integrated and segregated programs (including Similkameen). Additionally, 84 summer/fall Chinook were collected at the Okanogan adult weir in August and September. The summer/fall Chinook program collected enough brood to meet full production level. The cumulative pre spawn holding survival, for all Summer/Fall brood collected, was $82.7 \%$ for hatchery-origin broodstock (HOB) and $80.0 \%$ for natural-origin broodstock (NOB). The survival standard (90\%) was not met by both the hatchery-origin and natural-origin brood. Total green egg take for the season was 1,639,336 (60\% of full program). Egg survival from green egg to eyed egg averaged 80.4\% for NOB and 81.4\% for HOB, both under the survival standard (90\%) for this life stage. After in-hatchery mortalities from pre-spawn holding through ponding there were 1,156,450 fish on hand at the end of April for the yearling releases in 2022. ( $89 \%$ of the yearling program) and 177,115 fish on hand for the sub-yearling releases in May 2021 (25\% of full program).

2020 was the sixth year for Summer/Fall Chinook hatchery yearlings released from the CJH, Similkameen and Omak acclimation ponds. In April, 298,988 integrated yearling summer/fall Chinook were released from the Omak acclimation pond and 409,348 were released by Washington Department of Fish \& Wildlife (WDFW) from the Similkameen Pond; combined these programs were at 100\% of the full program goal of 800,000 integrated yearlings. There were no integrated or segregated sub-yearlings from brood year (BY) 2019 released in May 2020. However, there were 568,625 yearling Chinook released directly from Chief Joseph Hatchery (100\% of full program).

After release, the yearling programs from CJH and Omak Pond had lower survival when compared to previous years and other programs. Although survival was lower, the travel time was shorter, which was a confusing result that could not be explained. In contrast, subyearling
survival was similar to or better than previous years. The majority ( $>95 \%$ ) of PIT tagged hatchery smolts released from Omak Pond migrated to the lower Okanogan River within one month of release. Although overall outmigration was slower than in previous years, this assessment suggests that the program was successful at releasing actively migrating smolts.

The CJH monitoring project collected field data to determine Chinook population status, trend, and hatchery effectiveness centered on five major activities; 1) beach seine (naturalorigin smolt PIT tagging, smolt to adult return) 2) lower Okanogan adult fish pilot weir (adult escapement, proportion of hatchery-origin spawners [pHOS], broodstock) 3) spawning ground surveys (redd and carcass surveys)(viable salmonid population [VSP] parameters) 4) eDNA collection (VSP parameter-distribution/spatial structure) and 5) coded wire tag lab (extraction and reading). The rotary screw trap project was suspended in 2020 due to the program's inability to operate under CCT's COVID-19 safety guidelines.

Beach seining captured 20,340 juvenile Chinook and 18,700 (92\%) were PIT tagged and released. Pre- and post-tag mortality was $1.0 \%$ and $7.0 \%$ respectively. In 2020 , wild summer Chinook tagged at the mouth of the Okanogan had a minimum apparent survival of $43 \%$ ( $3 \%$ SE) to Rocky Reach Juvenile Bypass (RRJ) and 85\% (34\% SE) from RRJ to McNary (MCN).

The lower Okanogan Adult Fish Weir was deployed on August 17 when discharge was 1,900 cfs. The thermal barrier was present in the lower Okanogan after installation until August 25th when the mean Okanogan River temperature began dropping below $22.5^{\circ} \mathrm{C}$, allowing Chinook to migrate up the Okanogan. Trapping began on August 27 and continued until September 24 with the majority ( $66 \%$ ) being caught from August 30-September 7. Eight hundred and seventy adult Chinook were trapped in 2020. Eighty-four natural-origin Chinook were transported to the hatchery and held as broodstock for the integrated program. Adult brood were transported from the weir trap to the hatchery brood truck by foot using a rubber boot. There were no immediate mortalities of these fish within the first week after transport to the hatchery. All other natural-origin fish were released upstream of the weir unharmed. One hundred and fifty-eight hatchery-origin were removed from the weir trap for adult management purposes. $6.6 \%$ of the Chinook spawning escapement was detected in the trap. All Chinook and Sockeye mortality encountered at the weir were categorized as impinged on the upstream side, indicating that they most likely died upstream and floated down onto the weir. There were only 145 Chinook carcasses collected from August 28 to September 24. The head differential, river velocity, and trap capacity were within the NOAA standard operating criteria. Water quality information, including dissolved oxygen, turbidity, and total dissolved solids were collected to assess potential impacts to increased fish mortality. Weir trapping operations ceased on September 24.

Spawning ground surveys estimated 4,127 summer/fall Chinook redds and 2,604 carcasses were recovered (1,908 natural-origin and 696 hatchery-origin). Adult summer/fall Chinook spawning escapement in 2020 was estimated to be 11,019, with 7,957 natural-origin spawners and 3,062 hatchery origin spawners. In 2020, the effective pHOS (0.24) and
proportion of natural influence (PNI) (0.82) met the program objectives ( $<0.30 \mathrm{pHOS}$; $>0.67$ PNI). The five-year average for $\mathrm{pHOS}(0.25)$ and PNI (0.75) met the long-term goal ( $<0.30$ pHOS; >0.67 PNI). Selective harvest activities by CCT and WDFW contributed to the reduced pHOS and increased PNI in 2020. CCT removed 1,141 hatchery fish, including 88 jacks, during surplus events at the CJH ladder and trap, and tribal members removed another 1,658, including 49 jacks, at the Chief Joseph Dam tailrace fishery. The Harvest program's purse seine removed 170 hatchery fish, including 145 jacks. 62 natural-origin fish, including 3 jacks, were released during surplus at the Chief Joseph Hatchery ladder. The purse seine released 531 natural-origin fish, including 121 jacks during their efforts. The Okanogan temporary weir encountered 870 fish in 2020, in which 160 hatchery fish, including 6 jacks, were removed and 625 natural-origin fish, including 19 jacks, and 3 hatchery-origin fish were released back to the river. Within the WDFW state fishery above Wells Dam in the Columbia River, 1,088 hatchery Chinook (segregated and integrated fish), including 39 jacks, were harvested and 115 natural-origin Chinook, including 14 jacks were released back to the river.

The management strategy for the CJH integrated hatchery program in the Okanogan River appears to be having some of the intended effects on the spawning grounds. The intent of adding the Omak Acclimation Pond was to reduce spawning density and pHOS in the high density reaches of the upper Okanogan (06) and lower Similkameen (S1) and to increase spawning in the under-utilized lower and middle reaches of the Okanogan (02-05).

Indeed, spawner distributions have changed during the CJH-era (2016-2020) compared to years prior. We find an increased proportion of redds in reaches 02 thru 06, and reduced proportions in reaches S1 and S2 for years 2016-2020 compared to years 2006-2015. Additionally, carcass recovery data show shifts in the composition of spawners, with increased pHOS in the lower basin (Reach O2) and reduced pHOS upstream. These changes in composition and distribution of spawners across the basin are likely the results of hatchery acclimation strategies, specifically with hatchery fish relating to their Omak Pond acclimation site in the lower basin and should help with the effectiveness of natural-origin spawners in the prime spawning habitat in the upper basin (Reach 06 and S1).

The CJH coded wire tag lab was in its fifth year of operation in 2020. Coded wire tags were extracted and read from Chinook snout recoveries from broodstock, ladder surplus, purse seine harvest, and creel and spawning ground surveys. The development of in-house CWT reading continues to be a huge success, providing age- and origin data within 2-3 months of the spawning ground surveys utilizing Colville tribal staff, rather than outsourcing to another lab. The majority of the summer Chinook adult returns to the CJH ladder were CJH Segregated (59\%) followed by Wells Hatchery (18\%), Chelan Falls (9\%), Okanogan integrated (6\%), and three other programs made up the remaining $8 \%$.

The majority (66\%) of hatchery-origin spawners recovered on the spawning grounds in 2020 were from Similkameen (48\%) and Okanogan (27\%). Chief Joseph Hatchery segregated Chinook comprised $19 \%$ of the HOS on the Okanogan spawning grounds. The level of segregated
hatchery fish on the spawning grounds (5.4\%) did not meet the program objective of < $5 \%$ segregated pHOS and future management efforts should focus on reducing the stray rate of segregated hatchery fish to the Okanogan spawning grounds. Overall, the majority of fish acclimated at Similkameen Pond ended up spawning throughout the upper reaches of the Okanogan (reaches 05 \& 06) (31\%) and Similkameen Rivers (66\%). Reach S1, the location of the Similkameen acclimation site in the Similkameen River accounted for just over half of the estimated spawning by Similkameen Pond fish (55\%).

Fish released within the Okanogan Basin have consistently homed to their natal stream, and 2020 was not an exception. One of the goals of the CJHP is to redistribute Chinook spawners to the middle and lower portion of the Okanogan River instead of inundating the already saturated Similkameen River with additional spawners. Juvenile Chinook releases from the Omak Pond acclimation site are primarily spawning in the Okanogan River (92\% in 2018, $90 \%$ in 2019 , and $83 \%$ in 2020) instead of the Similkameen River. Specifically, the Omak Pondreared Chinook have spawned almost exclusively in the lower ( 03 reach) and middle ( 05 reach) sections of the Okanogan River.

The most recent brood year that could be fully assessed (through age 5) for stray rate of Okanogan/Similkameen fish to spawning areas outside the Okanogan was 2015. The 2015 brood year had a stray of $0.4 \%$ to non-target basins and $1.7 \%$ to non-target hatcheries, which was similar to the long term and recent five-year average ( $0.9 \%$ for non-target basins and $0.4 \%$ to non-target hatcheries).

An Annual Program Review (APR) was held in April 2021 to share hatchery production and monitoring data, review the salmon forecast for the upcoming year, and develop action plans for the hatchery, selective harvest, and monitoring projects. Based on an average preseason forecast of 59,600 Upper Columbia summer/fall Chinook, the plan for 2021 is to operate the hatchery at full program levels of 2 million summer/fall Chinook with $100 \%$ pNOB. pNOB was set at $50 \%$ natural-origin broodstock for the integrated program and CCT will plan to harvest their allocation of 5,618 with the selective harvest program, including removals at the purse seine, the weir, and at the hatchery ladder.

## Introduction

Salmon (Oncorhynchus spp.) and steelhead (O. mykiss) faced many anthropogenic challenges ever since European settlement of the Pacific Northwest. Harvest, hydropower development, and habitat alteration/disconnection have all had a role in reducing productivity or eliminating entire stocks of salmon and steelhead (MacDonald 1894; UCSRB 2007). These losses and reductions in salmon had a profound impact on Native American tribes, including the Confederated Tribes of the Colville Reservation. Hatcheries have been used as a replacement or to supplement the wild production of salmon and steelhead throughout the Pacific Northwest.

However, hatcheries and hatchery practices can pose a risk to wild populations (Busack and Currens 1995; Ford 2002; McClure et al. 2008). As more studies lead to a better understanding of hatchery effects and effectiveness, hatchery reform principles were developed (Mobrand et al. 2005; Paquet et al. 2011). The CJHP is one of the first of its kind to be structured using many of the recommendations emanating from Congress's Hatchery Reform Project, the Hatchery Science Review Group (HSRG) and multiple independent science reviews. Principally, the success of the program is not based on the ability to meet the same fixed smolt output or the same escapement goal each year. Instead, the program is managed for variable smolt production and natural escapement. Success is based on meeting targets for abundance and composition of natural escapement and hatchery broodstock (HSRG 2009). Chief Joseph Hatchery Program (CJHP) managers and scientists are accountable for accomplishments and/or failures, and therefore, have well-defined response alternatives that guide annual program decisions. For these reasons, the program is operated in a manner where hundreds of variables are monitored, and activities are routinely and transparently evaluated. Functionally, this means that directed research, monitoring, and evaluation (RM\&E) are used to determine status and trends and population dynamics and are conducted to assess the program's progress in meeting specified biological targets, measure hatchery performance, and in reviewing the key assumptions used to define future actions for the entire CJHP.

The actions being implemented by the Colville Tribes, in coordination with regional management partners, represent an extraordinary effort to recover Okanogan and Columbia River natural-origin Chinook Salmon populations. In particular, the Tribes have embraced hatchery program elements that seek to find a balance between artificial and natural production and address the goals of increased harvest and conservation.

Two hatchery genetic management plans (HGMPs) were initially developed for the CJH during the Northwest Power and Conservation Council (NPCC) three-step planning process one for summer/fall Chinook (CCT 2008a) and one for spring Chinook (CCT 2008b). Each of the two plans included an integrated and a segregated component. Integrated hatchery fish have a high proportion of natural origin parents, are released into the Okanogan River system and a proportion of these fish are expected to spawn in the natural environment. Segregated fish have primarily hatchery parents, are to be released from CJH directly into the Columbia River and adult returns are targeted exclusively for harvest.

In 2010 the CCT requested that the National Marine Fisheries Service (NMFS) designate a non-essential experimental population of spring Chinook in the Okanogan utilizing section 10(j) of the Endangered Species Act (ESA). In order to obtain a permit to transfer ESA listed fish from the Methow River to the Okanogan River, a new HGMP was developed (CCT 2013). Biological Opinions (BiOps) and permits have been issued by NMFS for the 2008 HGMPs, and CCT acquired a BiOp and permit for the 2013 spring Chinook in 2014. The program will be guided by all three HGMPs.

At full program the facility will rear up to 2 million summer/fall Chinook and 900,000 spring Chinook. Up to 1.1 million summer/fall Chinook will be released in the Okanogan and Similkameen Rivers as an integrated program and 900,000 will be released from CJH as a segregated program. Up to 700,000 segregated spring Chinook will be released from CJH and up to 200,000 Met Comp spring Chinook from the Winthrop National Fish Hatchery (WNFH) will be used to reintroduce spring Chinook to the Okanogan under section 10 (j) of the ESA. In 2017, the summer/fall and spring Chinook program's production level was set at full production capacity.

The CJHP will increase harvest opportunity for all anglers throughout the Columbia River and Pacific Ocean. Additionally, the Colville Tribes and other salmon co-managers have worked with the mid-Columbia Public Utility Districts to meet some of their hydro-system mitigation through hatchery production (CPUD 2002a; CPUD 2002b; DPUD 2002).

In order to make full use of the best science available the program operates on the following general principles ${ }^{1}$ :

1. Monitor, evaluate and adaptively manage hatchery and science programs
2. Manage hatchery broodstock to achieve proper genetic integration with, or segregation from natural populations
3. Promote local adaptation of natural and hatchery populations
4. Minimize adverse ecological interactions between hatchery- and natural-origin fish
5. Minimize effects of hatchery facilities on the ecosystem
6. Maximize survival of hatchery fish in integrated and segregated programs
7. Develop clear, specific, quantifiable harvest and conservation goals for natural and hatchery populations within an "All-H" (Hatcheries, Habitat, Harvest and Hydro) context
8. Institutionalize and apply a common analysis, planning, and implementation framework
9. Use the framework to sequence and or prioritize actions
10. Hire, train, and support staff in a manner consistent with successful implementation of the program
11. Conduct annual reviews to include peers, stakeholders, and regional managers, and
12. Develop and maintain database and information systems and a highly functional informational web-presence.

The CJHP annual RM\&E activities were focused on six primary field activities to provide data for answering key management questions. These activities included:

1. Rotary screw traps (juvenile outmigration, natural-origin smolt PIT tagging)
2. Beach seine (natural-origin smolt PIT tagging)
3. Lower Okanogan adult fish pilot weir (adult escapement, pHOS , broodstock)

[^0]4. Spawning ground surveys (redd and carcass surveys)(VSP parameters)
5. eDNA collection (VSP parameter—distribution/spatial structure)
6. Coded wire tag lab (extraction, reading, reporting)

Additional data compilation activities occurred and were necessary in conjunction with our field efforts to answer the key management questions. These included:

1. Harvest (ocean, lower Columbia, terminal sport, and CCT)
2. Query RMIS for coded wire tag (CWT) recoveries to evaluate strays and stock composition
3. Query PTAGIS for PIT tag returns at mainstem dams and tributaries
4. EDT model estimates for abundance and productivity (from OBMEP)

In-hatchery monitoring/data collection was focused in five areas (see Appendix A):

1. Broodstock collection and bio-sampling
2. Life stage survival
3. Disease monitoring
4. Tagging, marking, and release
5. Ladder surplus / pHOS reduction

## Study Area

The primary study area of the CJHP lies within the Okanogan River Subbasin and Columbia River near Chief Joseph Dam in north central Washington State (Figure 1). The Okanogan River measures approximately 185 km long and drains 2,316,019 ha, making it the third largest subbasin to the Columbia River. Its headwaters are in Okanagan Lake in British Columbia, from which it flows south through a series of four lakes before crossing into Washington State at Lake Osoyoos. Seventy-six percent of the area lies in Canada. Approximately 14 km south of the border, the Okanogan is joined by its largest tributary, the Similkameen River. The Similkameen River watershed is 510 km long and drains roughly 756,096 ha. The Similkameen contributes approximately 75\% of the flow to the Okanogan River. The majority of the Similkameen is located in Canada. However, part of its length within Washington State composes an important study area for CJHP. From Enloe Dam (Similkameen rkm 14) to its confluence with the Okanogan, the Similkameen River contains important Chinook pre-spawn holding and spawning grounds. Downstream of the Similkameen confluence, the Okanogan River continues to flow south for 119 km until its confluence with the Columbia River at Columbia River km 853, between Chief Joseph and Wells dams, near the town of Brewster, Washington.


Figure 1. Map of the U.S. portion of the Okanogan River Basin, the Chief Joseph Hatchery (CJH), Winthrop National Fish Hatchery (WNFH), Okanogan adult weir (Weir), Rotary screw trap (RST), and Chinook Salmon acclimation sites. Horizontal coordinate information is referenced to the North American Datum of 1983 (NAD

Similar to many western rivers, the hydrology of the Okanogan River watershed is characterized by high spring runoff and low flows occurring from late summer through winter. Peak flows coincide with spring rains and melting snowpack (Figure 2). Low flows coincide with minimal summer precipitation, compounded by the reduction of mountain snowpack. Irrigation diversions in the lower valley also contribute to low summer flows. As an example, at the town of Malott, Washington (rkm 27), Okanogan River discharge can fluctuate annually from less than 1,000 cfs to over 30,000 cfs (USGS 2005).

The Okanogan Subbasin experiences a semi-arid climate, with hot, dry summers and cold winters. Water temperature can exceed $25^{\circ} \mathrm{C}$ in the summer, and the Okanogan River surface usually freezes during the winter months. Precipitation in the watershed ranges from more than 102 cm in the western mountain region to approximately 20 cm at the confluence of the Okanogan and Columbia Rivers (NOAA 1994). About 50\% to $75 \%$ of annual precipitation falls as snow during the winter months.

For most of its length, the Okanogan River is a broad, shallow, low gradient channel with relatively homogenous habitat. There are few pools and limited large woody debris. Fine sediment levels and substrate embeddedness are high and large woody debris is rare (Miller et al. 2013). Towns, roads, agricultural fields, and residential areas are adjacent to the river through most of the U.S. reaches.

Near its mouth, the Okanogan River is affected by the Wells Dam on the Columbia River, which creates a lentic influence on the lowermost 27 km of the Okanogan River. Water level fluctuates frequently because of operational changes (power generation, storage) at Wells Dam.


Figure 2. Okanogan River mean daily discharge (blue lines) and water temperature (red lines) at Malott, WA (USGS Stream Gage 12447200).

## METHODS

## Tag and Mark Plan

HATCHERY SUMMER/FALL CHINOOK. -All summer/fall hatchery-origin Chinook were marked with an adipose fin clip to ensure differentiation from natural-origin fish in the field and in fisheries. Additionally, all summer/fall Chinook raised for the integrated program have been/will be tagged with a CWT (with distinct codes differentiated by release location), which is inserted into the snout of fish while in residence at the hatchery. A batch of 200,000 summer/fall Chinook in the segregated program will receive a CWT, so the presence or absence of a CWT in adipose-clipped fish is a partial diagnostic as to which program an ad-clipped, hatchery-origin fish belongs (Table 1). This will allow for selective efforts in broodstock collection, purse seining, and hatchery trapping activities to be program specific by determining the presence or absence of a CWT in the field. It was decided that losing some resolution on
field differentiation of the segregated and integrated populations was a good tradeoff in order to get the harvest information back from the batch of 200,000 CWT in the segregated program.

Under this strategy, a returning adult from the CJH with an adipose fin clip and CWT would be considered part of the integrated program and either collected for broodstock in the segregated program, allowed to escape to the spawning grounds (if pHOS is within acceptable levels), or removed from the population (for harvest or pHOS management). If a fish has an adipose fin clip but no CWT, then it is assumed from the segregated program (or a stray from another hatchery program) and removed for harvest or pHOS management. In this way, CWTs assist with in-season management of hatchery-origin stocks in the field. The 200,000 segregated fish with a CWT represent about $15 \%$ of the combined segregated $(900,000)$ and integrated ( 1.1 million) hatchery fish with a CWT. If smolt to adult survival and adult holding/migration behaviors are identical, this would mean that $15 \%$ of the subsequent generation of segregated fish would have a segregated parent and would not be consistent with the 'stepping stone' approach. However, segregated fish should spend less time holding at the mouth of the Okanogan and therefore have a lower probability of being collected as broodstock in the purse seine. CWT monitoring from broodstock collections during the first several years of returns will provide insight to this tradeoff.

Coded wire tags are recovered from salmon carcasses during Chief Joseph Hatchery ladder surplus, CCT creel surveys, CCT purse seine, Okanogan weir trapping, and spawning ground surveys in the Okanogan Basin. All recovered CWTs are sent to the Chief Joseph Hatchery coded wire tag lab for extraction, reading, and data upload to the Regional Mark Processing Center operated by the Pacific States Marine Fisheries Commission (PSMFC) ${ }^{2}$. These data are used to develop estimates of total recruitment, rate of return to point of release (homing), contribution to fisheries, survival rates, mark rate, and other parameters, helping inform future management and production decisions within the CJHP.

[^1]Table 1. General mark and tag plan for Chief Joseph Hatchery summer/fall Chinook.

| Mark Group | Target max <br> smolt released | Life-stage <br> released | \% CWT | Adipose <br> Fin-Clip | PIT tag |
| :--- | :--- | :--- | :--- | :--- | :--- |
| Okanogan <br> Integrated | $1,100,000$ |  |  |  |  |
| Similkameen | 400,000 | Yearling | $100 \%$ | $100 \%$ |  |
| Omak Pond | 400,000 | Yearling | $100 \%$ | $100 \%$ | 5,000 |
|  | 300,000 | Sub- <br> yearling | $100 \%$ | $100 \%$ | 5,000 |
| Chief Joseph <br> Segregated | 500,000 | Yearling | $20 \%$ | $100 \%$ | 5,000 |
|  | 400,000 | Sub- <br> yearling | $25 \%$ | $100 \%$ | 5,000 |
| Natural-Origin | RST and <br> Confluence Seine | N/A | $0 \%$ | $0 \%$ | $\leq 25,000$ |

${ }^{1}$ The original plan was to use Riverside Pond for approximately $1 / 3$ of the summer Chinook yearling production, however, to date it has been only been used to acclimate the $10(\mathrm{j})$ spring Chinook because Tonasket Pond has not been rehabilitated for acclimation of spring Chinook.

In addition to the adipose fin-clip and CWT, a subset of hatchery-origin fish will be PITtagged to further assist with fish monitoring efforts in subsequent years. Table 1 represents the general plan at full production.

NATURAL-ORIGIN FISH TAGGING. -The RM\&E plan called for up to 25,000 PIT tags in juvenile natural-origin summer/fall Chinook parr/smolts. PIT tagging of natural-origin summer/fall Chinook occurs at the rotary screw trap and the juvenile beach seine annually. Please see those sections for details.

## Genetic Sampling/Archiving

The CJHP collects and archives genetic samples for future analysis of allele frequency and genotyping of naturally spawned and hatchery Chinook populations. Genetic samples (fin clips) from outmigrant juvenile Chinook were collected during rotary screw trap operations. Samples were preserved in 200-proof molecular grade ethanol and are currently archived at the Chief Joseph Hatchery Science Program office in Omak, WA. Annual tissue collection targets are approximately $n=200$ samples for: (1) natural-origin sub-yearling Chinook handled at the rotary screw trap/beach seine; (2) natural-origin yearling ( $>130 \mathrm{~mm}$ ) Chinook handled at the rotary screw trap/beach seine and (3) natural- and hatchery-origin (100 each) Chinook encountered during carcass surveys on the spawning grounds and (4) natural-origin Chinook encountered during juvenile electrofishing surveys.

The CJHP has also supported requests from Columbia River Inter-tribal Fish Commission (CRITFC) to provide genetic samples (caudal punches) from CJH summer-Chinook broodstock to
aid in the development of a Columbia River Parentage Based Tagging (PBT) program. Samples were preserved on pre-labeled Whatman (GE Healthcare, Pittsburg, PA, USA) cellulose chromatography paper and shipped to CRITFC Lab in Hagerman, ID, USA. Genetic samples will continue to be collected from all hatchery broodstock at CJH.

## Rotary Screw Traps

CJH M\&E staff were regrettably unable to operate the RST due to the program's inability to operate under the Colville Tribe's COVID-19 safety guidelines.

## Juvenile Beach Seine and PIT Tagging

Portions of the following text describing the methods were taken directly from a draft DPUD report (DPUD 2014).

Beach seining took place from June 2- July 15 in the area near the confluence of the Okanogan and Columbia Rivers. Efforts at the confluence were focused on beaches along the North bank of the Columbia River, downstream of the mouth of the Okanogan ( $48^{\circ} 612.46 " \mathrm{~N}$, $119^{\circ} 44^{\prime} 35.48$ "W) (Figure 3). This area is known as Gebber's Landing. This location provided reasonable catch rates, limited bycatch, and provided suitable substrates (limited debris loads/underwater snags) for efficient sampling. Juvenile Chinook from this location were likely primarily fish originating from the Okanogan River based on results from stable isotope analysis conducted in 2018.


Figure 3. Seining location downstream (Gebber's Landing) of the confluence.
A single beach seine ( $30.49 \mathrm{~m} \times 3.05 \mathrm{~m}$ with a $28.32 \mathrm{~m}^{3}$ 'bag'; Christensen Net Works, Everson, WA) was used to capture fish. Netting was Delta woven 6.4 mm mesh with "fish-green" treatment. Weights ( $3-5 \mathrm{~kg}$ ) were attached to each end of the seine to help keep it open during retrieval.

To capture fish, one end of the seine was tied off to an anchor point onshore, while the other was towed out by boat until the seine was stretched perpendicular to shore. The boat would then pull the seine upstream and return to shore, causing the seine to form a semi-circle intersected by the shoreline (Figure 4). The seine bridle was handed from the boat to a shore crew that would retrieve the seine. Juvenile Chinook were transferred to a 10 -gallon tub filled with river water and transferred to a nearby floating net pen. Handling/holding time in the tub
was generally <15 minutes. Floating net pens were approximately $5 \mathrm{~m}^{3}$ and consisted of a PVC pipe frame covered with black $19.1-\mathrm{mm}$ and $3.2-\mathrm{mm}$ mesh. The mesh allowed for adequate water exchange, retained juvenile Chinook, and prevented the entrance of predators. Noticeable bycatch, most commonly three-spine stickleback (Gasterosteus aculeatus) were released from the seine without enumeration. Any bycatch inadvertently transferred to the floating net pen were later sorted and released during tagging (untagged).


Figure 4. Juvenile beach seine being retrieved by CCT staff near the confluence of the Okanogan and Columbia Rivers.

In most circumstances, juvenile Chinook were held 24 hours prior to tagging to assess capture/handling effects. Occasionally, due to staff availability or other complicating circumstances, fish were held for two days or released shortly after recovery from anesthesia. Chinook $\geq 65 \mathrm{~mm}$ were tagged with a full duplex 12 mm PIT tag, and Chinook between 65 and 50 mm were tagged with a full duplex 9 mm PIT tag. After tagging, fish were returned to a floating net pen for 24 hours post-tagging to assess tag loss and tag application/handling mortality rates. Fish were then released to the Columbia River (Wells Pool) several hundred meters downstream of their capture location.

TAGGING PROCEDURES. -Tagging was conducted by CCT staff with support from USGS using a mobile tagging station (Biomark, Co., Boise, ID, USA). The tagging station consisted of an approximately $1 \mathrm{~m}^{2}$ aluminum work surface with a trough for holding fish during the tagging process as well as all the necessary electronics (computer, tag reader, and antenna) needed for tagging. Water was pumped directly from the river using a $1 / 4$ horsepower pump. When tagging water temperatures were $>17{ }^{\circ} \mathrm{C}$, water was replaced in the trough with cooler water from the river. A solution of 4.0 g Tricaine methanesulfonate (MS-222) per 1 L of water was used to anesthetize fish prior to tagging. The applied concentration of MS-222 would sedate fish to the desired level of stage-2 anesthesia in approximately 3 to 4 minutes. All fish were tagged within 10 minutes of the initial exposure. Recovery time was approximately 1 to 2 minutes.

The tagging location had two net pens: one containing the fish to be tagged, and an empty pen for holding fish post-tagging. Fish to be tagged were collected from the respective net pens using a dip net and placed into an 18.9 L bucket of water. Up to 40 fish at a time were then transferred from the bucket using a smaller dip net and placed into the trough containing the anesthetic solution.

Fish were tagged with 12.5 mm 134.2 kHz ISO PIT tags using pre-loaded, 12-gauge hypodermic needles (BIO12.BPLT) fitted onto injection devices (MK-25). 12.5 mm PIT tags were used to maximize detection at downstream locations, particularly the Rocky Reach Juvenile Bypass and the Bonneville Dam Corner Collector, although 9 mm PIT tags were used in fish. Detection efficiencies at both of the former sites would dramatically suffer when using the smaller PIT tags available. The tagging crew consisted of one or two taggers and one data collector. The data collector interrogated the tag in each tagged fish, recorded its fork length and noted any anomalies. Tagged fish were transferred to the recovery/holding pen via a PVC pipe with flowing water.

Data collected during tagging were stored using PITTAG4 (P4) software (Pacific States Marine Fisheries Commission). After completion of the tagging events, tag files were consolidated, uploaded to PTAGIS (www.ptagis.org).

FISH RELEASES. -Tagged fish were released the morning after they had been tagged. Prior to release, the net pen was opened, and all observed mortalities and moribund fish were removed. Once the mortalities were removed, expelled (shed) tags from live fish were recovered from the mesh floor via a powerful magnet. After that was completed, the net pen was tilted to allow the fish to volitionally exit. PIT tags were recovered from dead/moribund fish, the associated tag codes were marked as "Mortalities" in the tag files and the tag codes were deleted. Carcasses of summer Chinook were returned back to the river.

## Lower Okanogan Adult Fish Pilot Weir

The Okanogan adult fish pilot weir (herein referred to as the 'weir') was in its ninth year of design modifications and testing in 2020. Continued operation and improvements to
the weir are a central part of CCT's strategy for the successful implementation of the CJHP summer/fall Chinook Salmon (Oncorhynchus tshawytscha) programs. Pilot weir test results are essential for updating key assumptions, operations, and design of the weir.

Objectives for the pilot weir in 2020 included:

1. Install the weir in early July and operate until late September under allowable flow conditions ( $<3,000 \mathrm{cfs}$ ) and temperature ( $<22.5^{\circ} \mathrm{C}$ );
2. Document environmental effects of the weir through collection of physical and chemical data in the vicinity of the weir;
3. Test weir trapping operations and the Whooshh ${ }^{\text {TM }}$ fish transport system including live Chinook capture, handling and release;
4. Direct observations and fish counts for estimating species composition, abundance, health, and timing to inform management decisions and future programoperations;
5. Collect NOR and/or HOR brood stock at the weir and transport safely to the CJH;
6. Test the weir configuration, including the location of the trap box, to meet the program's biological and brood-take goals
7. Test fish entrainment through the trap entrance chute and into the trap box

The lower Okanogan fish weir was installed approximately 1.5 km downstream of Malott, WA ( $48^{\circ} 16^{\prime} 21.54 \mathrm{~N} ; 119^{\circ} 43^{\prime} 31.98 \mathrm{~W}$ ) in approximately the same location as previous years. Weir installation began on August 17th at a river flow of 1,900 cfs. and was completed with the underwater video system on August 21st. An aluminum trap was installed near the center of the channel at the downstream end of the deep pool in the thalweg of the channel. The trap was 3 m wide, 6 m long and 3 m high (Figure 1). A fifteen-foot aluminum accelerator chute was installed at the downstream trap gate. The wings of the weir stretched out from either side of the chute towards the river banks, angling downstream in a slight $V$ configuration. The wings consisted of steel tripods with aluminum rails that supported the 3 m long Acrylonitrile butadiene styrene (ABS) pickets. Each panel was zip-tied to the adjacent panel for strength and stability. Gravel bags were placed between panels when needed to fill gaps that exceeded the target picket spacing. Picket spacing ranged from 2.5 to 5.1 cm (1 to 2 inch) in 1.2 cm . (half-inch) increments (Figure 2). Pickets were manually forced into the river substrate upon deployment and then as needed to prevent fish passage under the weir.

The river-right wing consisted entirely of 2.5 cm . picket spacing (Figure 2). A 3 m gap between the last panel and the right shoreline remained to allow for portage of small vessels around the weir. This was a very shallow gravelly area and under most flow conditions it did not appear to be a viable path for adult salmon passage. However, a set up floating panels that were attached to the substrate extended from the last panel to the river-right shore to limit escapement via this route. The river left wing had variable picket spacing to accommodate non-Chinook fish passage through the pickets. The primary objective of the
wider picket spacing was to allow sockeye ( $O$. nerka) to pass through the weir and reduce the number of sockeye that would enter the trap. River left was selected for this spacing to better accommodate observation/data collection regarding successful passage of smaller fish through the panels.


Figure 5. Lower Okanogan adult fish pilot weir, 2020. Photo taken in late- August.


Figure 6. Conceptual diagram of picket (ABS pipe) spacing within each panel (or set of 5 panels) at the Lower Okanogan adult fish pilot weir. A 15 ft entrance chute was installed at the lower trap gate in 2020.

Physical and chemical data were collected in the vicinity of the weir including the water depth ( ft ) inside the trap, water velocity ( $\mathrm{ft} / \mathrm{sec}$ ) upstream, downstream and in the weir trap, dissolved Oxygen (mg/L), total dissolved solids (TDS)(ppm), turbidity (NTU), temperature $\left({ }^{\circ} \mathrm{C}\right.$ ), discharge (cfs) and head differential (cm). Temperature and discharge were taken from the online data for the USGS gauge at Malott (http://waterdata.usgs.gov/wa/nwis/uv?site no=12447200). When river temperature exceeded $22.5^{\circ} \mathrm{C}$, trapping operations ceased and weir pickets on panels adjacent to the trap on both sides were raised to allow for unrestricted fish passage.

Five-minute tower observations were conducted at least three times a day, in the morning (0600-0800), early afternoon (1200-1400) and evening (1700-1900) and an estimate of the number fish observed was recorded. Ten-minute bank observations were conducted about 0.8 river km . downstream of the weir, around two pools, at least twice a day, in the morning and afternoon. An estimate of the number of fish observed below the weir was recorded. Algae and debris were cleared off of the weir at least once per day generally during mid-day (1100-1300), depending on when fish were migrating. Dead fish on the upstream side of the weir were enumerated, identified to species and the presence and extent of injuries were noted. The tail was cut off of each mortality before they were tossed downstream of the weir so that they would not be double counted during surveys.

Weir efficiency, a measure of the proportion of total spawning escapement encountered by the weir, was calculated by the equation;

$$
X=\frac{W_{T}}{T}
$$

where $X$ was weir efficiency, $W_{T}$ was the number of adult summer/fall Chinook encountered in the weir trap including released fish, and $T$ was the total summer/fall Chinook spawning escapement for the Okanogan River Basin.

Weir effectiveness was a measure of the proportion of the adult hatchery Okanogan summer/fall Chinook run encountered in the weir trap, becoming available for removal from the population as a form of adult fish management. It was calculated by the equation;

$$
Y=\frac{W_{H}}{W_{H}+H O S}
$$

where $Y$ is weir effectiveness, $W_{H}$ is the number of adult hatchery origin fish encountered in the weir trap, and HOS is the total number of hatchery origin spawners.

Trapping operations were conducted 24 hours/day, 7 days/week, under allowable temperature conditions ( $\leq 22.5^{\circ} \mathrm{C}$ ) for the season. Trapping operations were suspended for the majority of the season, from July 22-August 25 and August 28-September 8. The last day of trapping was on September 11th. When fish entered the trap during an active trapping session, the downstream gate was closed, and fish were identified and either released or collected for brood.

Trapping operations were conducted under allowable temperature ( $\leq 22.5^{\circ} \mathrm{C}$ ) and head differential ( $<10 \mathrm{~cm}$.) conditions for the season. Trapping operations began on August 27 and continued until September 24. When fish entered the trap during an active trapping session, the downstream fyke was closed and fish were identified and either released or collected for brood.

Eighty-four natural-origin Chinook were collected from the weir trap from August 31 -September 22 and transported to a 2,500 gallon hatchery truck via a rubber boot. The fish were then transported approximately 32 km to Chief Joseph Hatchery where they were held in the brood stock raceways until spawning in October. The Whoooshh ${ }^{\text {TM }}$ fish transport system was not deployed in 2020 staffing was limited to effectively operate the system during the season.

## Spawning Ground Surveys

The objectives for spawning surveys were to:

1. Estimate total spawning escapement based on the number of Chinook redds per reach
2. Estimate the proportion of natural spawners composed of hatchery-origin recruits ( pHOS )
3. Estimate pre-spawn mortality and mean egg retention for wild- and hatchery-origin spawners
4. Determine the origin (rearing/release facility) of hatchery-origin spawners (HOS) in the Okanogan and estimate the spawner composition of out-of-population and out-ofESU strays (immigration)
5. Estimate out-of-population stray rate for Okanogan hatchery Chinook and estimate genetic contribution to out-of-basin populations (emigration)
6. Determine age composition of returning adults through scale analysis
7. Monitor status and trends of demographic and phenotypic traits of wild- and hatchery-origin spawners (age-at-maturity, length-at-age, run timing, SAR)

## REDD SURVEYS

A primary metric used to monitor the status and trends of salmonid populations is spawning escapement. Estimates of spawning escapement can be calculated based on redd counts and expanded by sex-ratios (Matthews and Waples 1991, Gallagher et al. 2007). This requires intensive visual survey efforts conducted throughout the spawning area and over the course of the entire spawning period. Visual redd surveys were conducted to estimate the number of redds per survey reach from the mouth of the Okanogan River to Zosel Dam (river km 124); the Similkameen River from its confluence with the Okanogan River upstream to Enloe Dam (river km 14); and in the mainstem Columbia River from the mouth of the Okanogan River upstream to Chief Joseph Dam (Table 2). Weekly surveys were timed to coincide with spawning in the basin, generally beginning the last week of September or the first week of October and ending approximately the second week of November. Redds were counted using a combination of fixed-wing aerial flight surveys and inflatable raft float surveys.

Aerial surveys occurred once weekly throughout the spawning season, each covering the entire survey area. Aerial surveys were flown at low elevation and at moderate speeds to accommodate visual identification of redds. From the aircraft, a trained observer recorded the number and GPS coordinates of all new redds as the plane passed overhead. All data were recorded directly into a Jupiter Mesa rugged computer tablet (Jupiter Systems, Inc.). Aerial surveys were primarily used to document redds in areas inaccessible to rafts, or in areas of low redd densities, such that they did not warrant weekly float surveys. All data points were
visualized in ArcGIS (ESRI, Inc.), and quality controlled to ensure that redd counts were not duplicated during float surveys. Aerial surveys also served a secondary function of informing research crews where to focus weekly carcass recovery efforts (see below section on Carcass Surveys).

Float surveys occurred once daily, 5 days per week throughout the spawning season. Float surveys consisted of three 2-person teams using inflatable rafts to count redds while floating downstream. Each team was responsible for covering one-third of the river width, (1) left bank, (2) center, and (3) right bank. Each individual redd was counted and its position recorded directly into a Jupiter Mesa rugged computer tablet (Jupiter Systems, Inc.).

Table 2. Reach names and locations for the Okanogan and Similkameen for summer/fall Chinook Salmon spawning and carcass surveys.

| Stream | Code | Reach Description | River km |
| :---: | :---: | :---: | :---: |
| Okanogan | O1 | Mouth to Malott Bridge | $0.0-27.0$ |
|  | O2 | Malott Bridge to Okanogan Bridge | $27.0-41.8$ |
|  | O3 | Okanogan Bridge to Omak Bridge | $41.8-49.1$ |
|  | O4 | Omak Bridge to Riverside Bridge | $49.1-65.1$ |
|  | O5 | Riverside Bridge to Tonasket Bridge | $65.1-90.9$ |
|  | O6 | Tonasket Bridge to Zosel Dam | $90.9-124.0$ |
| Similkameen | S1 | Mouth to Oroville Bridge | $0.0-8.0$ |
|  | S2 | Oroville Bridge to Enloe Dam | $8.0-14.0$ |
| Canada | Cx | TBD | TBD |

All redds were classified as either a:

1. Test-redd (disturbed gravel, indicative of digging by Chinook, but abandoned or without presence of Chinook; generally, this classification is reserved for early season redd counts before substantial post-spawn mortalities have occurred as indicated by egg-voidance analysis of recovered carcasses). Test-redds do not contribute to annual redd counts.
2. Redd (disturbed gravel, characteristic of successful Chinook redd construction and/or with presence of Chinook).
Redds per reach were calculated for each week as the combined number of new redds counted during aerial- and float-surveys for a given week. Post-season analysis consisted of summing the
combined aerial- and float-survey weekly redd totals to calculate annual redd totals per reach, and per total survey area. Estimated total spawning escapement was then calculated by multiplying the total redd count by the expansion factor for the current year (2.670 for 2020). The expansion factor $=1+$ the number of males per female as randomly collected for broodstock at Wells Dam (1670:1.000 in 2020). Assumptions include:

Assumption I - Each redd was constructed by a single female Chinook, and each female Chinook constructed only one redd

Assumption II - The male: female ratio on the spawning grounds was the same for wild- and hatchery-origin Chinook, and is equal to the male: female ratio as randomly collected for broodstock at Wells Dam

Assumption III - Every redd was observable and correctly enumerated

## Escapement into Canada

In previous years, video systems operated by OBMEP and located in the fishways of Zosel Dam allowed observation of salmonids passing over Zosel Dam and potentially into the British Columbia portion of the Okanagan River Basin. For detailed methods within a particular year please see the Okanogan Basin Monitoring and Evaluation Program (OBMEP) annual reports posted at (http://www.colvilletribes.com/obmep_publications.php). However, in 2020 no video monitoring occurred. Therefore, any information regarding Chinook passage at Zosel Dam and/or escapement into the Canadian portion of the Okanagan basin in 2020 is extremely limited and are based primarily on in-stream PIT array data and anecdotal observations.

## CARCASS SURVEYS

Carcass surveys provide important biological samples for evaluation of hatchery- and natural-origin fish on the spawning grounds, including:

1) Spawner composition
a. pHOS
b. out of population hatchery strays (immigration)
c. spatial distribution of natural- and hatchery origin spawners
2) Fish size
3) Sex-ratio
4) Age structure (CWT and scale analysis)
5) Pre-spawn mortality (i.e., egg retention)

The target sample size for carcass recovery efforts is $20 \%$ of the spawning population within each reach (Hillman et al. 2014). Carcass recovery efforts occurred simultaneously with redd float surveys, although carcass efforts also occurred once a week during August and September on the Similkameen River as well. These surveys assessed potential prespawn mortality that occurred for those fish that held in the cooler waters of the Similkameen River before spawning began in October. Recovered carcasses were transported within inflatable rafts downstream until a suitable site was found for processing. If a carcass was too degraded to sample for biological data, it was returned to the river without sampling. All adipose absent carcasses were assumed to be of hatchery-origin, and all carcasses displaying an intact adipose fin were assumed to be of natural-origin ${ }^{3}$. Origin was later verified by results from the WDFW scale lab analyses. Biological data collected from carcasses included sex, fork length (FL) and post-orbital hypural length ( POH ) to the nearest cm , and estimated egg retention for all females ( 0 to 5,000 max; visually estimated). All eggs that were not estimated to be within a carcass were assumed to have been successfully deposited. Any female carcass containing an estimated 5,000 eggs were considered a pre-spawn mortality. Forceps were used to remove five scale samples from all natural-origin Chinook. Scales were adhered to desiccant scale cards for preservation and identified by sample number and sample date. At the conclusion of spawning season, scales were sent to WDFW for post-hoc age analysis. Age analysis data were used to assess age-at-return (run-reconstruction) and combined with biological data to assess length-atage. All Chinook were scanned for passive integrated transponder (PIT) tags and all PIT detections were recorded and later uploaded to PTAGIS. Carcasses were scanned with a T-wand (Northwest Marine Technology, Inc., Shaw Island, WA USA) for coded wire tags (CWT). If present, the snout portion was removed and individually bagged and labeled with species, origin, FL, river of recovery and date. After sampling each carcass, the caudal fin was removed before the carcass was returned to the river to avoid resampling on subsequent surveys. All data collected in the field were input directly into a Jupiter Mesa rugged computer tablet (Jupiter Systems, Inc.). Weekly carcass recovery totals were summed post-season to calculate annual carcass recovery totals per reach and per survey area.

Some key assumptions for carcass surveys included:
Assumption I - All carcasses had the same probability of being recovered on the spawning grounds (despite differences in sex, origin, size, or spawning location)

Assumption II - The diagnostic unit in which a carcass is recovered is the same as the reach in which the fish spawned

[^2]Assumption III - Sampled carcasses are representative of the overall spawning composition within each reach

## pHOS and PNI

pHOS was first calculated using the straightforward method of calculation for the population-level pHOS by simply dividing the number of hatchery-origin spawners by the total spawners, such that:

$$
p H O S=\frac{H O S_{O}}{\mathrm{HOS}_{O}+\mathrm{NOS}_{O}}
$$

where $H_{O S}$ is the total recovered hatchery-origin carcasses and NOS ${ }_{o}$ is the total recovered natural-origin carcasses. This simple algorithm does not account for assumed deficiencies in hatchery fish effectiveness (i.e., relative reproductive success) nor does it account for spatial variation in pHOS and unequal sampling effort across reaches. For example, reach S 1 tends to have a higher pHOS than other reaches because the Similkameen acclimation site is located in the reach. Likewise, the probability of recovering carcasses in low density spawning reaches is lower than in reaches with high density spawning. We have attempted to account for each of these factors.

Relative reproductive success has not been estimated for summer/fall Chinook in the Okanogan. One of the key assumptions in the In-Season Implementation Tool was that firstgeneration hatchery fish are less effective natural spawners than natural-origin fish. Currently, the hatchery fish effectiveness assumption for the Okanogan population is that first-generation hatchery-origin spawners are $80 \%$ as effective as natural-origin fish as contributing genes to the next generation ${ }^{4}$ This assumption is based on research conducted by Reisenbichler and McIntyre (1977) and Williamson et al. (2010). Therefore, the pHOS calculation was amended in 2013 to account for the reduction in hatchery spawner effectiveness, such that:

$$
\text { Effective } p H O S=\frac{0.8 \operatorname{HOS}_{O}}{0.8 \operatorname{HOS}_{O}+\mathrm{NOS}_{O}}
$$

Further refinement of the pHOS calculation was needed to account for non-random sampling of carcasses and variable pHOS across reaches. This was done by weighting each reach's overall contribution to system-wide pHOS according to the overall proportion of summer/fall Chinook redds that occurred within that reach.

First, the proportion of redds that corresponded to each reach was calculated by the equation:

$$
r e d d_{p, r}=\frac{r e d d_{r}}{r e d d_{o}}
$$

[^3]where, $r e d d_{r}$ is the number of documented redds that occur within reach $r$, reddo is the total number of redds documented in the U.S. portion in the Okanogan River Basin, and redd ${ }_{p, r}$ is the proportion of total redds that were documented in reach $r$.

Next, Effective pHOS was calculated separately for each sampled reach, $r$, so that:

$$
p H O S_{r}=\frac{0.8 \mathrm{HOS}_{r}}{0.8 \mathrm{HOS}_{r}+\mathrm{NOS}_{r}}
$$

where $\mathrm{pHOS}_{r}$ is the Effective pHOS calculation for reach $r$, and $\mathrm{HOS}_{r}$ and $\mathrm{NOS}_{r}$ are the total recovered carcasses of hatchery- and natural-origin within that reach. Finally, Effective pHOS was corrected for the proportion of redds in each reach to determine an adjusted Effective pHOS, such that:

$$
\text { Effective } p H O S=\sum_{i=1}^{n} p \operatorname{HOS}_{r}\left(\text { redd }_{p, r}\right)
$$

where $n$ is the total number of sampled reaches that compose the Okanogan River Basin. These calculations assumed that sampled carcasses were representative of the overall spawning composition within each reach; that no carcasses were washed downstream into another reach; that all carcasses had an equal probability of recovery; and that all fish within origin types had equal fecundity. While it is unlikely that all of these assumptions were correct, the modified calculation results in a better representation of the actual census pHOS .

PNI was calculated as:

$$
P N I=\frac{p N O B}{\text { Effective } p H O S+p N O B}
$$

where $p N O B$ was the proportion of broodstock that were natural-origin Okanogan returns, and Effective pHOS was the reach weighted effective pHOS defined previously. To determine an Okanogan specific pNOB, we applied the results of a radio tracking study, which estimated that $90 \%$ of the natural-origin fish detected near the mouth of the Okanogan River in 2011 and 2012 ended up spawning in the Okanogan Basin (Mann and Snow 2013). Therefore, we assumed that $90 \%$ of the NOB collected in the purse seine (2010-2013) was of Okanogan origin.

In years prior to 2010 all of the broodstock for the Similkameen program were collected at Wells Dam. That program strived for $100 \%$ pNOB and did achieve $>95 \%$ pNOB in 7 of the last 8 years (Hillman et al. 2014). However, the Wells Dam broodstock collection efforts composited natural-origin fish from the Okanogan and Methow populations as well as fish originating from downstream populations ${ }^{5}$. We made a correction for non-Okanogan NOB for all years when Wells Dam was used for brood collection using the formula:

[^4]Adjusted Wells Dam pNOB = Wells Dam pNOB $*\left(\frac{\text { Okanogan NOS }}{\text { Okanogan NOS }+ \text { Methow NOS }}\right)$
where the Adjusted Wells Dam pNOB was estimated based on the proportion of natural-origin spawners (NOS) that were in the Okanogan compared to the Methow for that particular year. This correction was made for a portion of the broodstock in 2010 and 2011 and all of the broodstock previous to 2010. This correction did not account for stray NORs from downstream populations or NORs that would have remained in the Columbia River above Wells Dam. Although the radio tracking study provides an estimate of this for 2011 and 2012, there was uncertainty regarding the applicability of the radio tracking data for years prior.

## Origin of Hatchery Spawners

Snouts from adipose fin clipped fish were removed, individually labeled, frozen, and delivered to the Chief Joseph Hatchery coded wire tag lab for CWT extraction and reading. The Regional Mark Information System (RMIS; http://www.rmis.org/rmis) was queried in March 2021 to assess the rearing facility of hatchery-origin Chinook recovered on the Okanogan spawning grounds, the in-to-basin stray rate, and the out-of-basin stray rates. RMIS data queries are described in detail in the 2013 CJHP Annual Report (Baldwin et al. 2016).

## Smolt-to-Smolt Survival and Travel Time

Survival and travel time were assessed using the Data Acquisition in Real Time (DART) website analysis tools. DART calculates a survival estimate using a Cormack Jolly Seber mark recapture model, for full details on the analysis methods please see the DART website (http://www.cbr.washington.edu/dart/query/pit sum tagfiles). Each CJH release group with PIT tags were queried for survival from release to Rocky Reach Dam Juvenile bypass (RRJ) and McNary Dam Juvenile bypass (MCN). Although some recaptures were obtained further downstream than McNary Dam, survival through the entire hydropower system to Bonneville Dam could not be generated because there were not enough recaptures downstream to estimate the recapture probability. Survival estimates and travel time for nearby hatcheries and the wild summer Chinook captured in the RST and beach seine were also analyzed for comparison purposes.

Survival estimates are 'apparent survival’ because they were not adjusted for residuals, tag failure, tag loss (shedding), or other factors which could result in fish not dying but not being detected at a downstream location. Due to these factors, actual survival would be higher than the apparent survival estimates provided in this report.

Migration timing from release to the lower Okanogan River was determined using a query of the PTAGIS database (https://www.ptagis.org/data/quick-reports/small-scale-sitedetections) to determine the timing of PIT tag detections from releases of Summer Chinook at Omak Pond. No PIT tags were released from Similkameen Pond in 2020. The lower Okanogan

River PIT tag interrogation site (OKL) is located at rkm 25 and is within 2 km of the inundation effects of Wells Dam.

## Smolt-to-adult Return

The smolt to adult return rate (SAR) was calculated using two different methods, PIT tags and coded-wire tags (CWT). For PIT tags, SAR was calculated for adult fish (age 4-6) from release, back to Bonneville and Wells dams using the formula:

$$
S A R=\frac{\# \text { PIT tags detected in adult ladders at dam } x}{\# \text { PIT tags released }}
$$

A correction was then applied to the SAR to account for adult fish harvested before reaching each dam. Standard harvest rates for each return year were applied based on harvest summaries for indicator stocks generated by the Technical Advisory Committee of US v Oregon.

The SAR for CWT was estimated as:

$$
S A R=\frac{\text { expanded CWT recoveries }}{C W T \text { released }}
$$

where expanded CWT recoveries included estimated expanded recoveries on the spawning grounds, at hatcheries and in fisheries. Two expansions were applied. First the number of recoveries was expanded to account for the proportion of the release group that wasn't tagged. For example, with a $99 \%$ CWT mark rate the recoveries would be increased by $1 \%$. Second, the recoveries were expanded based on the proportion of the population that was sampled. For example, if carcass surveys recovered $20 \%$ of the estimated spawners, then the number of CWT recoveries was expanded by $80 \%$. The number of CWT fish released were simply the hatchery release data including all tag codes for CWT released fish (CWT + Ad Clip fish and CWT-only fish).

## Coded Wire Tag Lab Analysis

Coded wire tags (CWT) from broodstock, ladder surplus, purse seine harvest, creel and spawning ground surveys were extracted, read, and reported in the Chief Joseph Hatchery Lab from December 2020 to February 2021. The snouts were then interrogated for the presence of a CWT by using a V-reader or T-wand. After positive detection, the snout was cut bilaterally into symmetrical portions keeping the half that indicated detection and discarding the other half into the snout bag from which it came. This process was then repeated until only a small piece of tissue containing the CWT remains. The final piece of tissue was then smeared on a cutting mat exposing the CWT, then placed on its corresponding snout card and finally on to a cafeteria tray (groups of $\sim 25$ tags) to be read under a microscope.

Extracted tags were removed from the tray one-by-one to be cleaned, recorded, and read. The CWT was cleaned by wetting a lint free cloth and rolling the tag between a finger and cloth to remove all remaining tissue. The CWT was attached to a Northwest Marine Technologies (NMT) magnetic pencil and inserted into a jig to be read under a LCD microscope with the aid of an illuminator. Biological data was transcribed from the snout card to a final CWT datasheet. The CWT was attached to this datasheet with tape after the six-digit code was read. Information from the datasheet was transferred to an excel workbook which contains all applicable CWT code combinations.

CWTs were expanded based on their tag loss and sample rate to estimate total catch contribution for a specific fishery. For each fishery, every CWT recovered and decoded was grouped according to their tag code with the total number of CWTs recovered from that release group, (e.g., tag code 200108 was recovered 10 times for a fishery/location (tag group 1). (See formula 1 below). Tag group 1 is then divided by the sum of all recovered/decoded CWTs for that specific fishery. This value was multiplied by the sum of all lost and scratched tags with tag group 1 being added to the end of the calculation. This provides an adjustment factor for lost and scratched tags for every unique tag code by hatchery of origin. Mark rates are typically high ( $\sim 99 \%$ ) for most Upper Columbia River release groups, however it is important to account for missing tags or tags that were shed during the fish's lifecycle. (See formula 2 below). Taking the adjustment factor for lost and scratched tags and multiplying it by the tag loss rate (tag loss rate can be found at www.RMPC.ORG) provides an adjustment for missing tags. These adjustments (lost/scratched/missing) can be summed together to provide total catch contribution for a fishery that was sampled at 100 percent. (See formula 3 below). When sampling occurred at less than 100 percent the adjustment total is divided by the sample rate to calculate the expanded number of fish for each release group.
(1) Adjustment for Lost/scratched tags:
$C W T_{\text {Adjustment }=\left(\text { Tag }_{\text {group } 1} / \Sigma \text { Total tags }\right) *\left(\sum \text { Lost }+ \text { scratched } \text { Tags }\right)+\text { Tag }_{\text {group } 1}}$
(2) Adjustment for tag loss:

(3) CWT expansion
$C W T_{\text {Adjustment }=\left\{\left(\text { Tag }_{\text {group } 1} / \Sigma \text { Total tags }\right) *\left(\sum \text { Lost }+ \text { scratched } \text { Tags }\right)+\text { Tag }_{\text {group } 1}\right\} *(\text { Tag loss Rate })}$
Sample Rate

Finally, after accounting for the mark rate of each group, the remaining ad-clip, no-CWT fish were assigned to the CJH segregated group.

## Results

## Rotary Screw Traps

There are no results for 2020 due to COVID safety guidelines.

## Juvenile Beach Seine and Pit Tagging

In 2020, 20,340 natural-origin juvenile salmonids were collected over the course of 26 fishing days (Table 3.). Out of the juvenile summer/fall Chinook collected, 18,700 (92\%) subyearling Chinook were PIT tagged and released (Figure 7). Pre- and post-tag mortality was $1.0 \%$ and $7 \%$ respectively. Twelve-hundred eighteen shed tags were recovered from the net pens prior to release, twelve-hundred twelve of which were from post-tag mortalities, and the other six were ejected from fish that were later released alive, but without a tag. All recovered tags were removed from the tagging file before upload to PTAGIS. Fish size fluctuated and increased through time (Figure 8), peaking the week beginning July 06 at which point water temperatures began to rapidly rise above $14^{\circ} \mathrm{C}$ (Table 3). We suspect that sub-yearling Chinook may have migrated downstream, or to deeper, cooler water making it difficult to collect them via beach seine, as has presumably happened in past years. Fork length for tagged fish ranged from 43-125 mm, with an average of 69.9 mm (SD 10.2 mm ) and a median of 70 mm (Figure 9). Bycatch included hatchery-origin juvenile Chinook, three-spine stickleback, mountain whitefish, smallmouth bass, and sculpin.

Table 3. Summary of juvenile Chinook beach seining effort at Gebber's Landing in 2020. This table excludes Chinook salmon that were captured, PIT tagged, and then recaptured in the beach seine.

| Week start | Gebber's Fish <br> Collected | Gebber's Fish <br> Tagged | Proportion <br> Gebber's <br> Fish Tagged |
| :---: | :---: | :---: | :---: |
| $6 / 4 / 2020$ | 722 | 512 | $71 \%$ |
| $6 / 10 / 2020$ | 4,141 | 3123 | $75 \%$ |
| $6 / 16 / 2020$ | 4,577 | 3,371 | $73 \%$ |
| $6 / 23 / 2020$ | 6,866 | 6,303 | $91 \%$ |
| $6 / 29 / 2020$ | 6,114 | 5,215 | $85 \%$ |
| $7 / 8 / 2020$ | 2,270 | 1,524 | $67 \%$ |
| Total | $\mathbf{2 4 , 6 9 0}$ | $\mathbf{2 0 , 0 4 8}$ |  |
| Mean | $\mathbf{4 , 1 1 5}$ | $\mathbf{3 , 3 4 1}$ | $\mathbf{7 7 \%}$ |



Figure 7. Total mortality and number of released natural-origin sub-yearling Chinook in 2019. Primary y-axis shows number of juvenile Chinook; secondary $y$-axis (right hand side) shows water temperature (degrees Celsius (C)).


Figure 8. Size distribution of PIT tagged juvenile Chinook by release date from the beach seine effort in 2020. Boxes encompass the $25^{\text {th }}$ to $75^{\text {th }}$ percentiles of measured fish; the mid-line in the box is the median fish length. FL = fork length in millimeters (mm).


Figure 9. Size distribution of natural origin sub-yearling Chinook tagged during the beach seining effort in 2020.

The Rocky Reach juvenile bypass system detected 2,389 PIT tagged juvenile Chinook from the beach seining effort, which was $12.7 \%$ of total fish tagged and released. One hundred seventy-three ( $0.9 \%$ ), 128 ( $0.6 \%$ ) and 113 ( $0.6 \%$ ) were detected at the McNary, John Day and Bonneville Dams respectively. Detections for sub-yearlings occurred primarily from mid-June to mid-August at all downriver dams (Figure 10). Utilizing the mark-recapture model from DART, the apparent survival rate was 43\% (SE 3\%) to Rocky Reach and 37\% (34\% SE) to McNary.



Figure 10. Daily distribution of detections of PIT-tagged sub-yearling Chinook at Rocky Reach, McNary, John Day, and Bonneville Dams in 2020. Note differences in scale on the y-axis. The y axes denote the numbers of PIT-tagged fish encountered daily at each of the mainstem project arrays.

Travel time from release to Rocky Reach Dam was the slowest compared to travel time from release to the other lower river dams - on average, fish moved downstream more quickly the further downstream they travelled (Table 4). Larger fish travelled faster to Rocky Reach Dam (Figure 11). This is similar to what was reported in 2011-2013 by Douglas County PUD and observed in previous years by CCT.

Table 4. Mean travel time (d) and rate (km/d) for PIT tagged sub-yearling Chinook released near Gebber's Landing and detected at Columbia River dam PIT arrays.

| Locatio <br> n (River <br> KM) | Rocky Reach(762) |  | McNary (470) |  | John Day (347) |  | Bonneville (235) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Travel <br> Time (d) | Rate (km/d ) | Travel <br> Time (d) | Rate $(\mathrm{km} / \mathrm{d}$ ) | Travel <br> Time (d) | Rate $(\mathrm{km} / \mathrm{d}$ ) | Travel Time <br> (d) | Rate $(\mathrm{km} / \mathrm{d}$ J |
| Releas <br> e (856) | 36.3 <br> (Standar <br> d <br> Deviatio $\begin{aligned} & \mathrm{n}=14.7 \\ & \mathrm{n}=2,386 \end{aligned}$ | 2.6 | 47.9 <br> (Standar <br> d <br> Deviatio $\begin{aligned} & \mathrm{n}=15.8 ; \\ & \mathrm{n}=168) \end{aligned}$ | 8.1 | 42.9 <br> (Standar <br> d <br> Deviatio $\begin{aligned} & \mathrm{n}=10.6 ; \\ & \mathrm{n}=128) \end{aligned}$ | 11.8 | 39.3 <br> (Standard <br> Deviation= $8.4 ; n=112)$ | 15.8 |
| Rocky Reach (762) |  |  | 10.1 <br> (Standar <br> d <br> Deviatio $\begin{aligned} & \mathrm{n}=5.4 ; \\ & \mathrm{n}=61) \end{aligned}$ | 28.9 | 14.2 <br> (Standar <br> d <br> Deviatio $\begin{aligned} & \mathrm{n}=6.5 ; \\ & \mathrm{n}=41) \end{aligned}$ | 29.2 | 15.6 <br> (Standard <br> Deviation = <br> 4.6; $n=25$ | 33.8 |
| McNary $(470)$ |  |  |  |  | 3.5 <br> (Standar <br> d <br> Deviatio $\begin{aligned} & \mathrm{n}=1.0 ; \\ & \mathrm{n}=4) \end{aligned}$ | 35.1 | 4.5 <br> (Standard <br> Deviation=0. <br> 3; $n=2$ ) | 52.2 |
| John <br> Day <br> (347) |  |  |  |  |  |  | 2.1 (Standard Deviation = 0.5 ; n=13) | 53.3 |




Figure 11. Fish size (fork length) and travel time of tagged Chinook to Rocky Reach Dam.

## Lower Okanogan Adult Fish Pilot Weir

The Okanogan River (at Malott) discharge was above normal in 2020, which delayed deployment of the weir until August 17 when the river flow went below 2,000 cfs (Figure 3). Discharge continued to drop throughout the season and was approximately 1,000 cfs by the time the weir was removed for the season on September 24.

Migration of sockeye and summer/fall Chinook is generally affected by a thermal barrier that is caused by warm water temperatures $\left(\geq \sim 22^{\circ} \mathrm{C}\right)$ in the lower Okanogan River. The thermal barrier is dynamic within and between years, but it generally sets up in mid-July and breaks down in late August. In some years, the Okanogan River will temporarily cool off due to a combination of interrelated weather factors including rainstorms, cool weather, cloud cover and/or wildfire smoke. This 'break' in the thermal barrier can allow a portion of the fish holding in the Columbia River to enter the Okanogan and migrate up to thermal refuge in the Similkameen River or Lake Osoyoos. In 2020, temperatures were similar to the median daily temperatures from the last 13 years (Figure 4). Daily mean temperature was above $22.5^{\circ} \mathrm{C}$ from July 21 to August 24. Daily mean temperature dropped below $22.5^{\circ} \mathrm{C}$ on August 25th and stayed below this mark until the end of the season.


Figure 12. Discharge of the Okanogan River between July 1 and October 31, 2020. This figure was copied directly from the USGS website (http://nwis.waterdata.usgs.gov/wa).


Figure 13. Temperature of the Okanogan River between July 1 and October 31, 2020. This figure was copied directly from the USGS website (http://nwis.waterdata.usgs.gov/wa).

Dissolved Oxygen varied from 6.9 to 8.3 mg . /L, total dissolved solids varied from 129180 ppm and turbidity varied from 0.7 and 2.1 NTUs (Table 1). The head differential was measured only when pickets were down and ranged from 1.0-4.0 cm. The maximum water velocity measured was 3.4 ft . /sec. (Table 2).

Table 5. Water quality data at or near the lower Okanogan weir in 2020. Temperature and discharge were taken from the USGS gage at Malott. Minimum depth allowed for trap depth is 6 inches and optimal dissolved oxygen levels for adult Chinook should not drop below $6 \mathrm{mg} / \mathrm{L}$.

| Date | Trap Depth <br> (ft) | Dissolved <br> Oxygen <br> (mg/L) | Total Dissolved Solids (ppm) | Turbidity <br> (NTU) | Mean Temperatur e( $\left.{ }^{\circ} \mathrm{C}\right)$ | Mean Discharge (cfs) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 8/24 | 1.7 | 7.80 | 180 | 1.2 | 23 | 1,430 |
| 8/25 | 1.7 | 7.80 | 134 | 1.4 | 22 | 1,470 |
| 8/26 | 1.7 | 7.85 | 132 | 2.1 | 22 | 1,470 |
| 8/27 | 1.7 | 7.74 | 133 | 1.3 | 22 | 1,450 |
| 8/28 | 1.7 | 8.01 | 135 | 2.0 | 22 | 1,440 |
| 9/2 | 1.6 | 7.89 | 139 | 1.3 | 22 | 1,330 |
| 9/3 | 1.6 | 7.00 | 133 | 1.5 | 22 | 1,300 |
| 9/4 | 1.6 | 6.93 | 136 | 1.1 | 22 | 1,300 |
| 9/5 | 1.6 | 7.05 | 139 | 1.1 | 22 | 1,290 |
| 9/9 | 1.5 | 7.98 | 133 | 0.8 | 19 | 1,290 |
| 9/10 | 1.5 | 8.23 | 132 | 0.8 | 19 | 1,190 |
| 9/11 | 1.4 | 7.36 | 134 | 1.6 | 20 | 1,030 |
| 9/14 | 1.3 | 7.98 | 129 | 0.7 | 18 | 1,010 |
| 9/15 | 1.5 | 7.46 | 134 | 1.2 | 18 | 997 |
| 9/17 | 1.4 | 8.30 | 138 | 0.8 | 19 | 982 |
| 9/18 | 1.4 | 7.65 | 136 | 0.8 | 19 | 984 |
| 9/21 | 1.4 | 7.41 | 131 | 0.8 | 18 | 993 |
| 9/22 | 1.4 | 7.75 | 135 | 0.8 | 19 | 989 |
| 9/23 | 1.4 | 7.25 | 137 | 1.1 | 18 | 1020 |
| Min | 1.3 | 6.9 | 129 | 0.7 | 18 | 982 |
| Max | 1.7 | 8.3 | 180 | 2.1 | 23 | 1,470 |

Table 6. Water velocity upstream (US) and downstream (DS) of the weir and in the trap. Velocity should not exceed $3.5 \mathrm{ft} / \mathrm{sec}$ Measurements are in $\mathrm{ft} / \mathrm{sec}$.

| Date | River Left <br> US | US Center | River Right <br> US | River Left <br> DS | DS Center | River Right <br> DS | Trap <br> Velocity |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $8 / 24$ | 1.8 | 1.9 | 2.7 | 2.6 | 2.3 | 3.4 | 1.0 |
| $8 / 25$ | 1.8 | 1.9 | 2.7 | 2.5 | 2.3 | 2.7 | 1.3 |
| $8 / 26$ | 1.9 | 1.9 | 2.5 | 3.0 | 2.0 | 2.8 | 1.0 |
| $8 / 27$ | 1.6 | 1.6 | 2.5 | 2.4 | 2.5 | 3.1 | 1.2 |
| $8 / 28$ | 1.6 | 1.5 | 1.9 | 2.7 | 2.5 | 3.1 | 1.1 |
| $8 / 31$ | 2.4 | 1.8 | 2.0 | 2.8 | 2.9 | 2.9 | 1.1 |
| $9 / 2$ | 1.4 | 1.4 | 2.9 | 3.0 | 2.3 | 2.5 | 1.2 |
| $9 / 3$ | 1.7 | 1.4 | 2.8 | 3.2 | 2.3 | 2.3 | 1.2 |
| $9 / 4$ | 1.3 | 1.2 | 2.1 | 2.6 | 2.6 | 2.4 | 1.1 |
| $9 / 5$ | 1.8 | 1.4 | 2.5 | 2.3 | 2.1 | 2.3 | 1.4 |
| $9 / 10$ | 2.0 | 1.4 | 1.7 | 2.7 | 2.2 | 2.0 | 1.1 |
| $9 / 11$ | 1.3 | 1.4 | 1.5 | 2.9 | 2.3 | 2.1 | 1.0 |
| $9 / 14$ | 1.7 | 1.5 | 2.3 | 2.1 | 1.9 | 2.5 | 1.0 |
| $9 / 15$ | 1.7 | 1.4 | 1.8 | 2.3 | 2.4 | 3.2 | 1.5 |
| $9 / 17$ | 1.4 | 1.5 | 1.5 | 2.4 | 3.0 | 2.0 | 0.9 |
| $9 / 18$ | 1.1 | 1.4 | 1.6 | 2.3 | 2.2 | 3.0 | 1.0 |
| $9 / 21$ | 1.2 | 1.2 | 1.6 | 2.0 | 2.2 | 2.6 | 0.8 |
| $9 / 22$ | 1.3 | 1.2 | 2.1 | 2.0 | 2.4 | 2.6 | 0.9 |
| $9 / 23$ | 1.5 | 1.3 | 1.3 | 2.1 | 2.2 | 2.5 | 0.8 |
| Min | 1.3 | 1.2 | 1.5 | 2.1 | 1.9 | 2.0 | 0.9 |
| Max | 2.4 | 1.9 | 2.9 | 3.2 | 3.0 | 3.4 | 1.5 |

One hundred and forty-five dead fish were removed from the weir between August 28 and September 24. The majority of the mortalities (67) were sockeye and 23 of the mortalities were Chinook. All mortalities were impinged on the upstream side of weir indicating that they had most likely died upstream and floated down onto the weir.

Tower observations showed that about half the fish were milling in the center section of the river, just below the trap with the rest being equally split between the river right and river left sections (looking downstream). Estimates were highest during the last weekend of August when mean daily river temperatures dropped below $22.5^{\circ} \mathrm{C}$. Bank observations showed that the number fish observed holding in the lower pool, 0.8 km below the weir, increased about one week after the thermal barrier breakdown in late August (Figure 5). Trapping operations were conducted from August 27-September 24 when river temperature was $\leq 22.5^{\circ} \mathrm{C}$. The total fish trapped at the weir in 2020 was 931 with $93 \%$ of them being Chinook salmon (Figure 6). Seventy-two percent of the Chinook trapped were released back into the river (Figure 7). Twenty-seven steelhead were trapped in 2020.

Eighty-four natural-origin Chinook were transported to the hatchery and held in the brood stock ponds concurrently with the fish taken for brood stock from the purse seine. Adult Chinook were transported from the weir trap to the hatchery brood truck via a rubber boot. We were unable to assess the pre-spawn mortality of the weir brood because they were mixed with the rest of the integrated brood when they were transported to the hatchery. Past efforts have not indicated a problem with survival of brood fish collected at the weir. If we need to assess pre-spawn mortality in future years, we will need to mark these fish before they are transported to the hatchery or before they are mixed with the other brood at the hatchery.


Figure 14. Estimate of Chinook observed from the bank at the lower pool, 0.8 km downstream of the weir. Primary y-axis indicates number of Chinook observed; secondary $y$-axis (right hand side) indicates the mean stream temperature in degrees Celsius (C).


Figure 15. Total number of fish trapped at the Okanogan weir in 2020.


Figure 16. Final destination of Chinook adults captured in the weir trap during trapping operations in 2020.

In 2020, 0.065 (6.5\%) of total spawning escapement was detected in the trap (i.e., weir efficiency) (Table 7). The potential weir effectiveness (if we had been removing all of the HOR encountered) was 0.043 (4.3\%).

Table 7. The number of hatchery and natural origin Chinook Salmon encountered at the lower Okanogan weir in 2020. Weir efficiency and effectiveness were metrics for evaluating the potential for the weir to contribute to the CJHP population management goals in the future.

| Survey <br> Year | Number of Days Trapped | Chinook Adults Encountered in the Weir Trap |  | Chinook Spawning Escapement Estimates ${ }^{\text {c,d }}$ |  | Weir Metrics |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Natural <br> Origin <br> (NOR) | Hatchery <br> Origin <br> (HOR) | Natural <br> Origin <br> (NOS) | Hatchery Origin (HOS) | Weir Efficiencya | Weir Effectiveness ${ }^{\text {b }}$ |
| 2013 | 23 | 73 | 18 | 5,627 | 2,567 | 0.010 | 0.007 |
| 2014 | 34 | 2,006 | 318 | 10,407 | 1,756 | 0.147 | 0.140 |
| 2015 | 34 | 36 | 19 | 10,439 | 3,297 | 0.004 | 0.005 |
| 2016 | 30 | 135 | 34 | 8,700 | 1,905 | 0.014 | 0.016 |
| 2017 | 24 | 344 | 103 | 5,429 | 1,139 | 0.058 | 0.075 |
| 2018 | 38 | 32 | 16 | 3,266 | 1,594 | 0.009 | 0.009 |
| 2019 | 5 | 119 | 24 | 2,604 | 2,849 | 0.023 | 0.008 |
| 2020 | 27 | 709 | 161 | 7,957 | 3,062 | 0.066 | 0.045 |
| Avg. | 27 | 432 | 87 | 6,804 | 2,271 | 0.041 | 0.038 |

[^5]${ }^{\text {d NOS }}$ and HOS estimates determined by 'reach-weighted' pHOS calculations

## Redd Surveys

In 2020, 4,127 summer/fall Chinook redds were counted in the Okanogan and Similkameen rivers using a combination of ground and aerial surveys (Figure 22). The number of redds counted in 2020 was higher than the previous year (2019) - higher than the long-term average and 5-year average (Table 8). Consistent with previous years, the majority of Chinook redds were located in reaches S1 (35\%), 06 (30\%), and 05 (17\%; Table 15). These three reaches accounted for $82 \%$ of the total Chinook spawning in the basin. The overall redd distribution across reaches was similar to previous years with the majority of spawning taking place in the upper Okanogan reaches (05 and 06) and lower Similkameen (S1) (Figure 18).

Estimated spawning escapement was 11,019 ( 4,127 redds $\times 2.67$ fish per redd) (Table 10). Since 1989, the summer/fall Chinook spawning escapement within the U.S. portion of the Okanogan River Basin has averaged 5,861 and ranged from 473 to 13,857 (Table 10).

Summer/fall Chinook redds were counted during spawning ground surveys between October 6- Nov 5 (Table 11).

Table 8. Total number of redds counted in the Okanogan River Basin, 1989-2020 and the averages for the total time series and the most recent 5-year period.

| Survey <br> Year | Number of summer Chinook redds |  |  |
| :---: | :---: | :---: | :---: |
|  | Okanogan River | Similkameen River | Total Count |
| 1989 | 151 | 370 | 521 |
| 1990 | 99 | 147 | 246 |
| 1991 | 64 | 91 | 155 |
| 1992 | 53 | 57 | 110 |
| 1993 | 162 | 288 | 450 |
| 1994 | 375* | 777 | 1,152 |
| 1995 | 267* | 616 | 883 |
| 1996 | 116 | 419 | 535 |
| 1997 | 158 | 486 | 644 |
| 1998 | 88 | 276 | 364 |
| 1999 | 369 | 1,275 | 1,644 |
| 2000 | 549 | 993 | 1,542 |
| 2001 | 1,108 | 1,540 | 2,648 |
| 2002 | 2,667 | 3,358 | 6,025 |
| 2003 | 1,035 | 378 | 1,413 |
| 2004 | 1,327 | 1,660 | 2,987 |
| 2005 | 1,611 | 1,423 | 3,034 |
| 2006 | 2,592 | 1,666 | 4,258 |
| 2007 | 1,301 | 707 | 2,008 |
| 2008 | 1,146 | 1,000 | 2,146 |
| 2009 | 1,672 | 1,298 | 2,970 |
| 2010 | 1,011 | 1,107 | 2,118 |
| 2011 | 1,714 | 1,409 | 3,123 |
| 2012 | 1,613 | 1,066 | 2,679 |
| 2013 | 2,267 | 1,280 | 3,547 |
| 2014 | 2,231 | 2,022 | 4,253 |
| 2015 | 2,379 | 1,897 | 4,276 |
| 2016 | 3,486 | 1,790 | 5,276 |
| 2017 | 2,434 | 787 | 3,221 |
| 2018 | 1,554 | 558 | 2,112 |
| 2019 | 1,638 | 733 | 2,371 |
| 2020 | 2,386 | 1,741 | 4,127 |
| Average | 1,299 | 1,038 | 2,276 |
| 5-yr Avg. | 2,300 | 1,122 | 3,421 |



Figure 17. Distribution of summer/fall Chinook redds in 2020. Individual redds are identified by red circles. Horizontal coordinate information is referenced to the North American Datum of 1983 (NAD 83).

Table 9. Annual and average abundance of summer/fall Chinook redds in each reach of the Okanogan (01-06) and Similkameen (S1-S2) Rivers from 2006-2020.

| Return <br> Year | Number of Summer Chinook Redds |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Okanogan |  |  |  |  |  |  |  | O1 |
|  | O-2 | O-3 | O-4 | O-5 | O-6 | S-1 | S-2 |  |  |
| 2006 | 10 | 56 | 175 | 145 | 840 | 1,366 | 1,388 | 278 | 4,258 |
| 2007 | 3 | 16 | 116 | 63 | 549 | 554 | 652 | 55 | 2,008 |
| 2008 | 4 | 51 | 60 | 96 | 374 | 561 | 801 | 199 | 2,146 |
| 2009 | 3 | 32 | 91 | 138 | 621 | 787 | 1,091 | 207 | 2,970 |
| 2010 | 9 | 58 | 67 | 89 | 357 | 431 | 895 | 212 | 2,118 |
| 2011 | 3 | 20 | 101 | 55 | 593 | 942 | 1,217 | 192 | 3,123 |
| 2012 | 12 | 54 | 159 | 68 | 555 | 765 | 914 | 152 | 2,679 |
| 2013 | 3 | 2 | 158 | 46 | 397 | 1,661 | 1,254 | 26 | 3,547 |
| 2014 | 11 | 57 | 191 | 111 | 851 | 1,010 | 1,737 | 285 | 4,253 |
| 2015 | 36 | 113 | 284 | 79 | 1,008 | 859 | 1,611 | 286 | 4,276 |
| 2016 | 2 | 57 | 52 | 130 | 907 | 2,338 | 1,645 | 145 | 5,276 |
| 2017 | 2 | 62 | 192 | 111 | 830 | 1,237 | 710 | 77 | 3,221 |
| 2018 | 11 | 74 | 211 | 133 | 618 | 507 | 501 | 57 | 2,112 |
| 2019 | 12 | 154 | 275 | 92 | 600 | 505 | 694 | 39 | 2,371 |
| 2020 | 25 | 51 | 270 | 103 | 683 | 1,254 | 1,445 | 296 | 4,127 |
| Average | $\mathbf{1 0}$ | $\mathbf{5 7}$ | $\mathbf{1 6 0}$ | $\mathbf{9 7}$ | $\mathbf{6 5 2}$ | $\mathbf{9 8 5}$ | $\mathbf{1 , 1 0 4}$ | $\mathbf{1 6 7}$ | 3,232 |



Figure 18. Proportion of redds in each reach of the Okanogan and Similkameen Rivers from 2006 to 2020.

Table 10. Spawning escapements for summer/fall Chinook in the Okanogan and Similkameen Rivers for return years 1989-2020.

| Return Year | Fish/Redd Ratio | Spawning Escapement |  |  |
| :---: | :---: | :---: | :---: | :---: |
|  |  | Okanogan | Similkameen | Total |
| 1989* | 3.300 | 498 | 1,221 | 1,719 |
| 1990* | 3.400 | 337 | 500 | 837 |
| 1991* | 3.700 | 237 | 337 | 574 |
| 1992* | 4.300 | 228 | 245 | 473 |
| 1993* | 3.300 | 535 | 950 | 1,485 |
| 1994* | 3.500 | 1,313 | 2,720 | 4,033 |
| 1995* | 3.400 | 908 | 2,094 | 3,002 |
| 1996* | 3.400 | 394 | 1,425 | 1,819 |
| 1997* | 3.400 | 537 | 1,652 | 2,189 |
| 1998 | 3.000 | 264 | 828 | 1,092 |
| 1999 | 2.200 | 812 | 2,805 | 3,617 |
| 2000 | 2.400 | 1,318 | 2,383 | 3,701 |
| 2001 | 4.100 | 4,543 | 6,314 | 10,857 |
| 2002 | 2.300 | 6,134 | 7,723 | 13,857 |
| 2003 | 2.400 | 2,505 | 915 | 3,420 |
| 2004 | 2.300 | 2,986 | 3,735 | 6,721 |
| 2005 | 2.900 | 4,720 | 4,169 | 8,889 |
| 2006 | 2.020 | 5,236 | 3,365 | 8,601 |
| 2007 | 2.200 | 2,862 | 1,555 | 4,418 |
| 2008 | 3.250 | 3,725 | 3,250 | 6,975 |
| 2009 | 2.540 | 4,247 | 3,297 | 7,544 |
| 2010 | 2.810 | 2,841 | 3,111 | 5,952 |
| 2011 | 3.100 | 5,313 | 4,368 | 9,681 |
| 2012 | 3.070 | 4,952 | 3,273 | 8,225 |
| 2013 | 2.310 | 5,237 | 2,957 | 8,194 |
| 2014 | 2.860 | 6,381 | 5,783 | 12,164 |
| 2015 | 3.215 | 7,648 | 6,099 | 13,747 |
| 2016 | 2.010 | 7,007 | 3,598 | 10,605 |
| 2017 | 2.039 | 4,963 | 1,605 | 6,568 |
| 2018 | 2.301 | 3,576 | 1,284 | 4,860 |
| 2019 | 2.300 | 3,767 | 1,686 | 5,453 |
| 2020 | 2.670 | 6,371 | 4,648 | 11,019 |
| Average | 2.858 | 3248 | 2793 | 5861 |
| 5-Year Avg. | 2.324 | 4690 | 2299 | 6989 |

* Spawning escapement was calculated using the "Modified Meekin Method" (i.e., $3.1 \times$ jack multiplier).

Note: All values have been updated from previous reports to account for low sample rates (i.e., carcass recoveries). For any reach with carcass recoveries <5\%, the annual basin composition (i.e., HOS:NOS) was used to determine the number of HOS and NOS.

Table 11. Number and timing of summer Chinook redd counts in reaches of the Okanogan and Similkameen Rivers in 2020.

| Reach | River mile | $\begin{gathered} \text { Oct } 5 \\ 11 \end{gathered}$ | $\begin{gathered} \text { Oct } 12 \\ -18 \end{gathered}$ | $\begin{gathered} \text { Oct } 19 \\ -25 \end{gathered}$ | $\begin{gathered} \text { Oct } 26 \\ - \text { Nov } \\ 1 \\ \hline \end{gathered}$ | $\begin{gathered} \text { Nov } 2 \\ -8 \end{gathered}$ | Redd Count | Percent |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 01 | 0.0-16.9 | 0 | 5 | 18 | 2 | 0 | 25 | 1\% |
| 02 | $\begin{gathered} 16.9- \\ 26.1 \end{gathered}$ | 5 | 17 | 22 | 7 | 0 | 51 | 2\% |
| 03 | $\begin{gathered} 26.1- \\ 30.7 \end{gathered}$ | 24 | 107 | 134 | 5 | 0 | 270 | 11\% |
| 04 | $\begin{gathered} 30.7- \\ 40.7 \end{gathered}$ | 7 | 63 | 25 | 8 | 0 | 103 | 4\% |
| 05 | $\begin{gathered} \hline 40.7- \\ 56.8 \end{gathered}$ | 497 | 76 | 103 | 7 | 0 | 683 | 29\% |
| 06 | $\begin{gathered} 56.8- \\ 77.4 \end{gathered}$ | 651 | 522 | 67 | 14 | 0 | 1254 | 53\% |
|  |  | 1184 | 790 | 369 | 43 | 0 | 2386 | 100\% |
| Similkameen River |  |  |  |  |  |  |  |  |
| S1 | 0.0-1.8 | 880 | 557 | 0 | 8 | 0 | 1445 | 83\% |
| S2 | 1.8-5.7 | 142 | 152 | 0 | 2 | 0 | 296 | 17\% |
| Total |  | 1022 | 709 | 0 | 10 | 0 | 1741 | 100\% |

## Escapement into Canada

Methodological uncertainties have limited our confidence in Chinook escapement estimates into the Canadian portion of the Okanogan basin. Prior to 2018, insights into escapement into Canada had been based primarily on video counts of fish ascending the passageway at Zosel Dam, with the important caveat being that due to the variations in dam operations, there was uncertainty regarding the proportion of fish that are passing within range of the video system, and thus, available for counting. Additionally, fish fallback and re-ascension is known to occur (as indicated by limited PIT tag data), though the frequency of occurrence is poorly understood. With these uncertainties in mind, we present Canadian escapement information for years prior to 2018 (Table 18). No video count data exists for Chinook from 2018-2020. Average Chinook passage at Zosel Dam for years 2010 thru 2017 has been 1,315, with a minimum of 263 (2010) and a maximum of 2,276 (2013).

Table 12. Chinook escapement to Canada as estimated by Zosel Dam counts and Okanogan Nation Alliance area-under-the-curve (AUC) methods.

| Year | Zosel Dam <br> Video Count | Zosel Dam \% <br> Hatchery | ONA AUC <br> Spawner <br> Estimate | ONA AUC \% <br> Hatchery |
| :---: | :---: | :---: | :---: | :---: |
| $2006^{\text {a }}$ | 481 | $1 \%$ | 34 | $3 \%$ |
| 2007 | 455 | $40 \%$ | 7 | $0 \%$ |
| 2008 | 267 | $29 \%$ | 14 | $23 \%$ |
| $2009^{\text {a }}$ | 256 | $17 \%$ | 6 | $0 \%$ |
| 2010 | 359 | $29 \%$ | 5 | $0 \%$ |
| $2011^{\text {a }}$ | 1415 | $36 \%$ | 21 | $21 \%$ |
| $2012^{\text {a }}$ | 826 | $24 \%$ | 11 | $10 \%$ |
| 2013 | 2275 | $14 \%$ | 40 | $13 \%$ |
| $20144^{\text {b }}$ | 1188 | $10 \%$ | 52 | $13 \%$ |
| 2015 | 1206 | $7 \%$ | 61 | $8 \%$ |
| 2016 | 1823 | $13 \%$ | 40 | $5 \%$ |
| 2017 | 737 | $14 \%$ | 55 | $6 \%$ |
| 2018 | No Data | No Data | 10 | $20 \%$ |
| 2019 | No Data | No Data | 15 | $18 \%$ |
| 2020 | No Data | No Data | 79 | $7 \%$ |
| Average | $\mathbf{9 4 1}$ | $\mathbf{1 9 \%}$ | $\mathbf{3 0}$ | $\mathbf{1 0 \%}$ |

${ }^{a}$ AUC spawner estimates is based on the number of carcasses sampled so this is the minimum estimate. ${ }^{\text {b }} 2014$ data were adjusted for fallback/re ascension, down camera time, and differentiation of spring Chinook from summer/fall Chinook.

More recently, Okanagan Nation Alliance (ONA) has provided information on escapement estimates in Canada based on live counts of summer Chinook adjusted by the residency estimate for the area-under-the-curve (AUC). Using AUC estimation methods, the largest spawner estimate occurred more recently in 2020 with 79 spawners (Table 12).

## Carcass Surveys

In 2020, 2,604 carcasses were recovered on the spawning grounds, including 1,908 natural-origin and 696 hatchery-origin ${ }^{6}$. An additional 4 carcasses were recovered during pre-spawn surveys ( 1 ad-clipped, 3 ad-present). The spawning ground carcass recovery rate was $23.6 \%$ of the total spawning escapement. Similar to previous years, the majority of carcasses ( $n=2,229$; 86\%) were collected from reaches 05, 06 and S1 (Figure 19, also see

[^6]Appendix C). Regarding the distribution of carcasses throughout the basin, the proportions of natural-origin carcasses recovered in 2020 were similar in all reaches, compared to the average of the 10 years preceding Chief Joseph Hatchery (Figure 19, panel A). The proportions of hatchery-origin carcasses recovered in 2020 were significantly higher in reaches 01-04 and lower in reaches $05,06, \mathrm{~S} 1$ and S 2 compared to the average of the 10 years preceding Chief Joseph Hatchery (Figure 19, panel B).



Figure 19. Distribution of (A) natural-origin and (B) hatchery-origin summer/fall Chinook carcasses recovered in the Okanogan (reaches 01-06) and Similkameen (reaches S1-S2)

Rivers in 2020 compared to the average of the 10 years preceding Chief Joseph Hatchery (2006-2015). Error bars represent standard deviation (SD).

In the Okanogan basin, just 16 of the 1,389 sampled female carcasses were estimated to have retained all their eggs. Therefore, pre-spawn mortality, (for fish that survived to the spawn period) was estimated to be $0.99 \%$ for natural-origin females and $1.60 \%$ for hatchery-origin females (Table 13). Overall egg retention of all fish sampled (including fish that had expelled a portion of their eggs) was 3.15\%.

Table 13. Egg retention and pre-spawn mortality of sampled summer/fall Chinook carcasses in the Okanogan Basin.

| Year | Origin | Total carcasses sampled | Female carcasses sampled | Potential egg deposition | $\underset{\text { retained }}{\text { Eggs }}$ | aEgg retention rate | ${ }^{\text {b Pre- }}$ spawn mortality rate |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2013 | Natural | 613 | 326 | 1,630,000 | 6,152 | 0.40\% | 0.00\% |
|  | Hatchery | 297 | 237 | 1,185,000 | 10,970 | 0.90\% | 0.00\% |
|  | Total | 910 | 563 | 2,815,000 | 17,122 | 0.60\% | 0.00\% |
| 2014 | Natural | 2,123 | 1,136 | 5,680,000 | 373,708 | 6.60\% | 1.40\% |
|  | Hatchery | 329 | 166 | 830,000 | 81,105 | 9.80\% | 1.80\% |
|  | Total | 2,452 | 1,302 | 6,510,000 | 454,813 | 7.00\% | 1.50\% |
| 2015 | Natural | 2,554 | 981 | 4,905,000 | 609,869 | 12.40\% | 10.90\% |
|  | Hatchery | 738 | 340 | 1,700,000 | 96,354 | 5.70\% | 5.00\% |
|  | Total | 3,292 | 1,321 | 6,605,000 | 706,223 | 10.70\% | 9.40\% |
| 2016 | Natural | 2,171 | 1,370 | 6,850,000 | 300,046 | 4.38\% | 3.43\% |
|  | Hatchery | 584 | 434 | 2,170,000 | 66,254 | 3.05\% | 2.76\% |
|  | Total | 2,755 | 1,804 | 9,020,000 | 366,300 | 4.06\% | 3.27\% |
| 2017 | Natural | 997 | 592 | 2,960,000 | 17,345 | 0.59\% | 0.00\% |
|  | Hatcher | 204 | 129 | 645,000 | 24,997 | 3.88\% | 3.10\% |
|  | Total | 1,201 | 721 | 3,605,000 | 42,342 | 1.17\% | 0.55\% |
| 2018 | Natural | 374 | 251 | 1,255,000 | 3,075 | 0.25\% | 0.00\% |
|  | Hatchery | 173 | 123 | 615,000 | 16,024 | 2.61\% | 3.25\% |
|  | Total | 547 | 374 | 1,870,000 | 19,099 | 1.02\% | 1.07\% |
| 2019 | Natural | 229 | 122 | 610,000 | 5,680 | 0.93\% | 0.82\% |
|  | Hatchery | 244 | 161 | 805,000 | 22,149 | 2.75\% | 2.48\% |
|  | Total | 473 | 283 | 1,415,000 | 27,829 | 1.97\% | 1.77\% |
| 2020 | Natural | 1,908 | 826 | 4,045,568 | 84,432 | 2.04\% | 0.99\% |
|  | Hatchery | 696 | 252 | 1,260,000 | 53,552 | 4.25\% | 1.60\% |
|  | Total | 2,604 | 1,078 | 5,305,568 | 137,984 | 3.15\% | 1.30\% |

${ }^{\text {a Assuming fecundity of } 5,000 \text { eggs per female, egg retention rate is calculated as: (\# eggs estimated remaining in sampled }}$ female carcasses) / (\# female carcasses sampled * 5,000 eggs each)
${ }^{\mathrm{b}}$ A pre-spawn mortality is determined when a female retains the assumed 5,000 eggs on the spawning grounds.

## PHOS AND PNI

There was an increase in the proportion of hatchery-origin spawners (pHOS) across all lower reaches (01-04) in 2020 compared to the 10 years preceding Chief Joseph Hatchery (Figure 20). Hatchery-origin spawners comprised $28 \%$ of the spawn escapement estimate in the U.S. portion of the Okanogan, which was almost half the pHOS observed in 2019 (.52). After corrections for hatchery fish effectiveness assumptions ( 0.80 relative reproductive success rate for hatchery-origin spawners) the effective pHOS for 2020 was 0.24 , which was just below the five-year average (0.25) (Table 15). The five-year average is currently meeting the biological objective for $\mathrm{pHOS}(<0.3)$ (Figure 21).

The proportion of natural-origin broodstock (pNOB) in 2020 was 0.99 and the pNOB for Okanogan origin fish was 0.89 (Table 15). The resulting PNI for 2020 was 0.82 , with a 5-year average PNI of 0.75. The 5-year average is still meeting the Biological Objective ( $>0.67$ ) (Figure 22).


Figure 20. Okanogan (01-06) and Similkameen (S1-S2) river summer/fall Chinook pHOS (unadjusted for RSS) by reach. Red bars represent the average of the 10 years preceding Chief Joseph Hatchery (2006-2015), green bars represent the average of the years since Chief Joseph Hatchery operation (2016-2020), and blue bars represent the current year (2020). Reaches with <5\% carcasses recoveries were omitted. Error bars represent standard deviation.

Table 14. Natural- (NOS) and hatchery- (HOS) origin spawner abundance and composition for the Okanogan River Basin, brood years 1989-2020.

| Brood Year | Spawners |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
|  | NOS | HOS | pHOS | Effective pHOS^ |
| 1989 | 1,719 | 0 | 0 | 0 |
| 1990 | 837 | 0 | 0 | 0 |
| 1991 | 574 | 0 | 0 | 0 |
| 1992 | 473 | 0 | 0 | 0 |
| 1993 | 915 | 570 | 0.38 | 0.33 |
| 1994 | 1,323 | 2,710 | 0.67 | 0.62 |
| 1995 | 979 | 2,023 | 0.67 | 0.62 |
| 1996 | 568 | 1,251 | 0.69 | 0.64 |
| 1997 | 862 | 1,327 | 0.61 | 0.55 |
| 1998 | 600 | 492 | 0.45 | 0.4 |
| 1999 | 1,274 | 2,343 | 0.65 | 0.6 |
| 2000 | 1,174 | 2,527 | 0.68 | 0.63 |
| 2001 | 4,306 | 6,551 | 0.6 | 0.55 |
| 2002 | 4,346 | 9,511 | 0.69 | 0.64 |
| 2003 | 1,933 | 1,487 | 0.43 | 0.38 |
| 2004 | 5,309 | 1,412 | 0.21 | 0.18 |
| 2005 | 6,441 | 2,448 | 0.28 | 0.23 |
| 2006 | 6,787 | 1,814 | 0.21 | 0.18 |
| 2007 | 2,730 | 1,688 | 0.38 | 0.33 |
| 2008 | 2,820 | 4,155 | 0.60 | 0.54 |
| 2009 | 4,100 | 3,443 | 0.46 | 0.40 |
| 2010 | 3,178 | 2,773 | 0.47 | 0.41 |
| 2011 | 4,618 | 5,063 | 0.52 | 0.47 |
| 2012 | 4,521 | 3,704 | 0.45 | 0.40 |
| 2013a | 5,627 | 2,567 | 0.31 | 0.27 |
| 2014 | 10,407 | 1,756 | 0.14 | 0.12 |
| 2015 | 10,439 | 3,308 | 0.24 | 0.20 |
| 2016 | 8,700 | 1,905 | 0.18 | 0.15 |
| 2017 | 5,429 | 1,139 | 0.17 | 0.14 |
| 2018 | 3,266 | 1,594 | 0.33 | 0.28 |
| 2019 | 2,604 | 2,849 | 0.52 | 0.47 |
| 2020 | 7,957 | 3,062 | 0.28 | 0.24 |
| Average | 3,650 | 2,359 | 0.38 | 0.34 |
| 5-year Average | 5,591 | 2,110 | 0.30 | 0.26 |

${ }^{\text {a }} 2013$ data have been updated to reflect age and origin data acquired from scale reading since the publication of the 2013 annual report.
${ }^{\wedge}$ Effective pHOS assumes 0.80 HOS effectiveness
Note: All values have been updated from previous reports to account for low sample rates (i.e., carcass recoveries). For any reach with carcass recoveries $<5 \%$, the annual basin composition (i.e., HOS: NOS) was used to determine the number of HOS and NOS.


Figure 21. Annual and 5-year average proportion of hatchery-origin spawners (pHOS) in the Okanogan and Similkameen River (combined) from 1998-2020. pHOS values represent the effective pHOS (adjusted for RRS).

Table 15. Okanogan basin summer/fall Chinook spawn escapement, broodstock composition, pHOS, and PNI for Brood Years 1989-2020.

| Brood Year | Spawners |  |  | Broodstock |  |  |  |  | PNI | Okan. PNI |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | NOS | HOS | Effecti ve pHOS | NOB | Okan <br> NOB | HOB | pNOB | Okan pNOB |  |  |
| 1989 | 1,719 | 0 | 0.00 | 1,297 |  | 312 | 0.81 |  | 1.00 |  |
| 1990 | 837 | 0 | 0.00 | 828 |  | 206 | 0.80 |  | 1.00 |  |
| 1991 | 574 | 0 | 0.00 | 924 |  | 314 | 0.75 |  | 1.00 |  |
| 1992 | 473 | 0 | 0.00 | 297 |  | 406 | 0.42 |  | 1.00 |  |
| 1993 | 915 | 570 | 0.33 | 681 |  | 388 | 0.64 |  | 0.66 |  |
| 1994 | 1,323 | 2,710 | 0.62 | 341 |  | 244 | 0.58 |  | 0.48 |  |
| 1995 | 979 | 2,023 | 0.62 | 173 |  | 240 | 0.42 |  | 0.40 |  |
| 1996 | 568 | 1,251 | 0.64 | 287 |  | 155 | 0.65 |  | 0.50 |  |
| 1997 | 862 | 1,327 | 0.55 | 197 |  | 265 | 0.43 |  | 0.44 |  |
| 1998 | 600 | 492 | 0.40 | 153 | 77 | 211 | 0.42 | 0.21 | 0.51 | 0.35 |
| 1999 | 1,274 | 2,343 | 0.60 | 224 | 112 | 289 | 0.44 | 0.22 | 0.42 | 0.27 |
| 2000 | 1,174 | 2,527 | 0.63 | 164 | 82 | 337 | 0.33 | 0.16 | 0.34 | 0.21 |
| 2001 | 4,306 | 6,551 | 0.55 | 12 | 46 | 345 | 0.03 | 0.13 | 0.06 | 0.19 |
| 2002 | 4,346 | 9,511 | 0.64 | 247 | 124 | 241 | 0.51 | 0.25 | 0.44 | 0.29 |
| 2003 | 1,933 | 1,487 | 0.38 | 381 | 191 | 101 | 0.79 | 0.40 | 0.67 | 0.51 |
| 2004 | 5,309 | 1,412 | 0.18 | 506 | 253 | 16 | 0.97 | 0.48 | 0.85 | 0.73 |
| 2005 | 6,441 | 2,448 | 0.23 | 391 | 196 | 9 | 0.98 | 0.49 | 0.81 | 0.68 |
| 2006 | 6,787 | 1,814 | 0.18 | 500 | 250 | 10 | 0.98 | 0.49 | 0.85 | 0.74 |
| 2007 | 2,730 | 1,688 | 0.33 | 456 | 228 | 17 | 0.96 | 0.48 | 0.74 | 0.59 |
| 2008 | 2,820 | 4,155 | 0.54 | 359 | 202 | 86 | 0.81 | 0.45 | 0.60 | 0.46 |
| 2009 | 4,100 | 3,443 | 0.40 | 503 | 254 | 4 | 0.99 | 0.50 | 0.71 | 0.55 |
| 2010 | 3,178 | 2,773 | 0.41 | 484 | 242 | 8 | 0.98 | 0.49 | 0.71 | 0.54 |
| 2011 | 4,618 | 5,063 | 0.47 | 467 | 332 | 26 | 0.95 | 0.67 | 0.67 | 0.59 |
| 2012 | 4,521 | 3,704 | 0.40 | 107 | 96 | 0 | 1.00 | 0.90 | 0.72 | 0.69 |
| 2013 | 5,627 | 2,567 | 0.27 | 353 | 318 | 0 | 1.00 | 0.90 | 0.79 | 0.77 |
| 2014 | 10,407 | 1,756 | 0.12 | 499 | 449 | 5 | 0.99 | 0.89 | 0.89 | 0.88 |
| 2015 | 10,439 | 3,308 | 0.20 | 421 | 379 | 9 | 0.98 | 0.88 | 0.83 | 0.81 |
| 2016 | 8,700 | 1,905 | 0.15 | 584 | 526 | 0 | 1.00 | 0.90 | 0.87 | 0.86 |
| 2017 | 5,429 | 1,139 | 0.14 | 350 | 315 | 17 | 0.95 | 0.86 | 0.87 | 0.86 |
| 2018 | 3,266 | 1,594 | 0.28 | 193 | 174 | 212 | 0.48 | 0.43 | 0.63 | 0.60 |
| 2019 | 2,604 | 2,849 | 0.47 | 376 | 338 | 224 | 0.63 | 0.56 | 0.57 | 0.55 |
| 2020 | 7,957 | 3,062 | 0.22 | 530 | 477 | 5 | 0.99 | 0.89 | 0.82 | 0.80 |
| Average | 3,650 | 2,359 | 0.34 | 415 | 246 | 146 | 0.74 | 0.55 | 0.68 | 0.59 |
| 5-Year <br> Average | 5,591 | 2,110 | 0.25 | 408 | 367 | 88 | 0.81 | 0.73 | 0.75 | 0.74 |



Figure 22. Annual and 5-year average proportionate natural influence (PNI) in the Okanogan basin from 1998 to 2020.

## Age Structure

Attempts were made to age all carcasses recovered on the spawning grounds, either by microscopy of scale annuli for natural-origin fish or by extracting and reading coded wire tag information for hatchery-origin fish. Historically, most natural-origin summer Chinook out migrate as sub-yearlings, while the majority of hatchery-origin releases in the Okanogan River basin have been released as yearlings. To account for this difference, the number of winters a fish spent in the marine environment - salt age - is the format of reported data.

In 2020, male natural-origin spawners were comprised predominantly 3-year salt age fish, which is different than previous years (Figure 28-a). Natural-origin female spawner age structures were skewed towards 3-year salt age fish similar to previous years (Figure 23-b). With 1,013 natural-origin female Chinook collected on the spawning grounds in 2020, 102 were determined to be 4-year salt age. Hatchery-origin males were comprised by 2- and 3-year salt age fish. No 4-year hatchery-origin males were recovered.

Hatchery-origin females were also comprised of 2- and 3-year fish, and no 4-year fish were recovered.





Figure 23. The salt ages of carcasses collected on the spawning grounds of the Okanogan and Similkameen rivers in 2020 along with 10-year averages (2011-2020) for a) Naturalorigin males; b) Natural-origin females; c) Hatchery-origin males; and d) Hatchery-origin females.

## HATCHERY-ORIGIN STRAY RATES

Strays to the Okanogan - The majority (75\%) of hatchery-origin spawners recovered on the spawning grounds in 2020 were from the integrated CJH program Similkameen (48\%) and Okanogan acclimated (27\%) releases (Table 16). The majority of strays from outside the Okanogan were from the Chief Joseph Hatchery segregated program (19\%), whereas strays from other hatchery programs in the Methow River, Entiat River, Chelan River, and mainstem Columbia River releases comprised 6\% (Table 20). The
contribution of stray hatchery fish to total spawn escapement was 7\% (i.e., stray pHOS) (Table 17). This was above the recent (2006-2020) average of $4.21 \%$ and also above the biological target of $<5 \%$. Note that this includes those fish released from the Chief Joseph Hatchery segregated program which comprised $5.4 \%$ of the spawner composition.

Strays outside the Okanogan - With the caveat that data are likely to continue to be updated in future reports as more data becomes available through the RMIS database, the most recent brood year that could be fully assessed (through age 5) for stray rate of Okanogan fish to spawning areas outside the Okanogan was 2015. The 2015 brood year had a stray rate of $2.1 \%$ (includes straying to out of basin spawning grounds and hatcheries), which was slightly above the long term (1989-2015; 1.2\%) and recent fiveyear (2011-2015; 1.5\%) averages (Table 18). For return year 2020, RMIS queries revealed an estimate of 4 Okanogan hatchery-origin Chinook recovered on spawning grounds in non-target spawning areas in 2020 (Table 18). Okanogan basin hatchery program strays comprised $0.24 \%$ to Methow spawner composition in 2020 (Table 19). 5-year averages to Wenatchee, Methow, Chelan, and Entiat basins are all below 1\%.

Table 16. Estimated number (and percent of annual total) of hatchery-origin spawners from different release basins recovered on the Okanogan/Similkameen spawning grounds, based on CWT recoveries and expansions, for return years 20062020.

| Return Year |  |  | Release Site |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Summer Chinook Run |  |  |  |  |  | Spring and Fall Chinook Run |  |  |
|  | Homing Fish |  | Straying Fish |  |  |  |  |  |  |  |  |
|  | Okanogan River Basin |  | Within ESU Stray |  |  |  |  |  | Out of ESU Stray |  |  |
|  | Okanogan River ${ }^{\text {a }}$ | Similkamee n River ${ }^{\text {b }}$ | Methow River ${ }^{\text {c }}$ | Wenatchee River ${ }^{\text {d }}$ | Entiat River ${ }^{\text {e }}$ | Chelan River ${ }^{f}$ | Chief <br> Joseph <br> Hatchery <br> (Seg.) | Mainstem Columbia Rivers | Mainstem Columbia River ${ }^{\text {h }}$ | Snake <br> River ${ }^{\text {i }}$ | Other ${ }^{\text {i }}$ |
| 2006 | 0 (0\%) | 709 (87\%) | 12 (2\%) | 12 (2\%) | 0 (0\%) | 0 (0\%) |  | 81 (10\%) | 0 (0\%) | $\begin{gathered} 0 \\ (0 \%) \end{gathered}$ | $\begin{gathered} 0 \\ (0 \%) \end{gathered}$ |
| 2007 | 0 (0\%) | 1121 (95\%) | 17 (1\%) | 5 (0\%) | 0 (0\%) | 0 (0\%) |  | 42 (4\%) | 0 (0\%) | $\begin{gathered} 0 \\ (0 \%) \\ \hline \end{gathered}$ | $\begin{gathered} 0 \\ (0 \%) \\ \hline \end{gathered}$ |
| 2008 | 0 (0\%) | 3224 (95\%) | 11 (0\%) | 24 (1\%) | 0 (0\%) | 4 (0\%) |  | 133 (4\%) | 3 (0\%) | $\begin{gathered} 0 \\ (0 \%) \\ \hline \end{gathered}$ | $\begin{gathered} 0 \\ (0 \%) \\ \hline \end{gathered}$ |
| 2009 | 0 (0\%) | 2733 (95\%) | 14 (0\%) | 14 (0\%) | 0 (0\%) | 9 (0\%) |  | 99 (3\%) | 0 (0\%) | $\begin{gathered} 5 \\ (0 \%) \\ \hline \end{gathered}$ | $\begin{gathered} 4 \\ (0 \%) \\ \hline \end{gathered}$ |
| 2010 | 4 (0\%) | 2165 (89\%) | 44 (2\%) | 35 (1\%) | 0 (0\%) | $\begin{gathered} 110 \\ (5 \%) \end{gathered}$ |  | 75 (3\%) | 0 (0\%) | $\begin{gathered} 4 \\ (0 \%) \\ \hline \end{gathered}$ | $\begin{gathered} 0 \\ (0 \%) \end{gathered}$ |
| 2011 | 219 (5\%) | 4196 (93\%) | 44 (1\%) | 5 (0\%) | 0 (0\%) | $\begin{gathered} 34 \\ (1 \%) \end{gathered}$ |  | 22 (0\%) | 0 (0\%) | $\begin{gathered} 6 \\ (0 \%) \end{gathered}$ | $\begin{gathered} 0 \\ (0 \%) \end{gathered}$ |
| 2012 | 379 (13\%) | 2397 (83\%) | 29 (1\%) | 23 (1\%) | 0 (0\%) | $\begin{gathered} 17 \\ (1 \%) \\ \hline \end{gathered}$ |  | 52 (2\%) | 0 (0\%) | $\begin{gathered} 0 \\ (0 \%) \\ \hline \end{gathered}$ | $\begin{gathered} 0 \\ (0 \%) \\ \hline \end{gathered}$ |
| 2013 | 254 (14\%) | 1437 (81\%) | 10 (1\%) | 54 (3\%) | 0 (0\%) | 0 (0\%) |  | 10 (1\%) | 0 (0\%) | $\begin{gathered} 0 \\ (0 \%) \end{gathered}$ | $\begin{gathered} 0 \\ (0 \%) \\ \hline \end{gathered}$ |
| 2014 | 55 (5\%) | 1023 (90\%) | 16 (1\%) | 0 (0\%) | 6 (1\%) | $\begin{gathered} 12 \\ (1 \%) \\ \hline \end{gathered}$ |  | 29 (3\%) | 0 (0\%) | $\begin{gathered} 0 \\ (0 \%) \\ \hline \end{gathered}$ | $\begin{gathered} 0 \\ (0 \%) \\ \hline \end{gathered}$ |


| 2015 | $38(1 \%)$ | $2562(91 \%)$ | $70(3 \%)$ | $17(1 \%)$ | 19 <br> $(1 \%)$ | $33(1 \%)$ |  | $33(1 \%)$ | $4(0 \%)$ | 4 <br> $(0 \%)$ | 21 <br> $(1 \%)$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2016 | $81(4 \%)$ | $1963(91 \%)$ | $42(2 \%)$ | $7(0 \%)$ | $3(0 \%)$ | 31 <br> $(1 \%)$ |  | $14(1 \%)$ | $0(0 \%)$ | 0 <br> $(0 \%)$ | $17(1 \%$ <br> $)$ |
| 2017 | $249(20 \%)$ | $590(46 \%)$ | $0(0 \%)$ | $0(0 \%)$ | $0(0 \%)$ | $0(0 \%)$ | 428 <br> $(33 \%)$ | $9(1 \%)$ | $0(0 \%)$ | $3(0 \%)$ | 0 <br> $(0 \%)$ |
| 2018 | $357(24 \%)$ | $628(43 \%)$ | $27(2 \%)$ | $0(0 \%)$ | $0(0 \%)$ | $6(0 \%)$ | 396 <br> $(27 \%)$ | $28(2 \%)$ | $0(0 \%)$ | 0 <br> $(0 \%)$ | 36 <br> $(2 \%)$ |
| 2019 | $403(24 \%)$ | $1250(44 \%)$ | $68(2 \%)$ | $0(0 \%)$ | $9(0 \%)$ | 37 <br> $(1 \%)$ | 1021 <br> $(36 \%)$ | $25(1 \%)$ | $0(0 \%)$ | 7 <br> $(0 \%)$ | 0 <br> $(0 \%)$ |
| 2020 | $813(27 \%)$ | $1,470(48 \%)$ | $65(2 \%)$ | $5(0 \%)$ | 17 <br> $(1 \%)$ | 18 <br> $(1 \%)$ | 589 <br> $(19 \%)$ | $78(3 \%)$ | $0(0 \%)$ | 0 <br> $(0 \%)$ | 0 <br> $(0 \%)$ |
| Avg. | 1490 <br> $(10 \%)$ | 1831 <br> $(78 \%)$ | $31(1 \%)$ | $13(1 \%)$ | 4 |  |  |  |  |  |  |
| $(0 \%)$ | 21 <br> $(1 \%)$ | 609 <br> $(29 \%)$ | $49(3 \%)$ | $0(0 \%)$ | 2 <br> $(0 \%)$ | 5 <br> $(0 \%)$ |  |  |  |  |  |

${ }^{\text {a }}$ Includes releases from Omak Pond and Bonaparte Pond. Three spring Chinook recovered in 2008 from an Omak Creek release were excluded from analysis.
${ }^{\mathrm{b}}$ Includes releases from Similkameen Pond
c Includes releases from Carlton Acclimation Pond and Goat Wall Acclimation Pond
${ }^{d}$ Includes releases from Dryden Pond and Eastbank Hatchery
${ }^{e}$ Includes releases from Entiat NFH
${ }^{f}$ Includes releases from Chelan PUD Hatchery, Chelan River NFH, and Chelan Falls Hatchery
g Includes releases of summer Chinook from Wells Hatchery, Turtle Rock Hatchery, and Grant County PUD Hatchery
${ }^{\text {h }}$ Includes releases of fall Chinook from Hanford Reach
${ }^{\text {i }}$ Includes Releases from NPT Hatchery
j Includes releases from Marion Yakama Tribal, Cle Elum Hatchery, Irrigon, and Prosser Hatchery

Table 17. Percent of the total Okanogan spawning escapement comprised of various hatchery release groups, based on CWT recoveries and expansions for return years 2006-2020.

| Retur n Year |  | Release Site |  |  |  |  |  |  |  |  |  | HOS Stray Contributi on to Total Spawning Escapemen t | $\begin{gathered} \mathrm{pHO} \\ \mathrm{~S} \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Summer Chinook Run |  |  |  |  |  |  | Fall Chinook Run Out of ESU Stray |  |  |  |  |
|  | Okanogan River Basin |  | Within ESU Stray |  |  |  |  |  |  |  |  |  |  |
|  | Okanoga n River ${ }^{\text {a }}$ | Similkame en Riverb | Metho <br> w <br> River ${ }^{\text {c }}$ | Wenatch ee River ${ }^{\text {d }}$ | Entia <br> t <br> River <br> e | Chela <br> n <br> River ${ }^{f}$ | Chief Joseph Hatchery (Seg.) | Mainst em Colum bia Riverg | Mainste <br> m Columbi a Riverh | Snak <br> e <br> River | Other <br> j |  |  |
| 2006 | 0.00\% | 8.20\% | 0.10\% | 0.10\% | $\begin{gathered} 0.00 \\ \% \end{gathered}$ | 0.00\% |  | 0.90\% | 0.00\% | $\begin{gathered} 0.00 \\ \% \end{gathered}$ | $\begin{gathered} 0.00 \\ \% \end{gathered}$ | 1.10\% | 0.18 |
| 2007 | 0.00\% | 25.40\% | 0.40\% | 0.10\% | $\begin{gathered} 0.00 \\ \% \end{gathered}$ | 0.00\% |  | 1.00\% | 0.00\% | $\begin{gathered} 0.00 \\ \% \end{gathered}$ | $\begin{gathered} 0.00 \\ \% \end{gathered}$ | 1.50\% | 0.33 |
| 2008 | 0.00\% | 46.20\% | 0.20\% | 0.30\% | $\begin{gathered} 0.00 \\ \% \end{gathered}$ | 0.10\% |  | 1.90\% | 0.00\% | $\begin{gathered} 0.00 \\ \% \end{gathered}$ | $\begin{gathered} 0.00 \\ \% \end{gathered}$ | 2.50\% | 0.54 |
| 2009 | 0.00\% | 36.20\% | 0.20\% | 0.20\% | $\begin{gathered} 0.00 \\ \% \end{gathered}$ | 0.10\% |  | 1.30\% | 0.00\% | $\begin{gathered} 0.10 \\ \% \end{gathered}$ | $\begin{gathered} 0.10 \\ \% \end{gathered}$ | 2.00\% | 0.4 |
| 2010 | 0.10\% | 36.40\% | 0.70\% | 0.60\% | $\begin{gathered} 0.00 \\ \% \end{gathered}$ | 1.80\% |  | 1.30\% | 0.00\% | $\begin{gathered} 0.10 \\ \% \end{gathered}$ | $\begin{gathered} 0.00 \\ \% \end{gathered}$ | 4.50\% | 0.41 |
| 2011 | 2.30\% | 43.30\% | 0.50\% | 0.10\% | $\begin{gathered} 0.00 \\ \% \end{gathered}$ | 0.40\% |  | 0.20\% | 0.00\% | $\begin{gathered} 0.10 \\ \% \end{gathered}$ | $\begin{gathered} 0.00 \\ \% \end{gathered}$ | 1.30\% | 0.47 |
| 2012 | 4.60\% | 29.10\% | 0.40\% | 0.30\% | $\begin{gathered} 0.00 \\ \% \end{gathered}$ | 0.20\% |  | 0.60\% | 0.00\% | $\begin{gathered} 0.00 \\ \% \end{gathered}$ | $\begin{gathered} 0.00 \\ \% \end{gathered}$ | 1.50\% | 0.40 |
| 2013 | 3.10\% | 17.50\% | 0.10\% | 0.70\% | $\begin{gathered} 0.00 \\ \% \end{gathered}$ | 0.00\% |  | 0.10\% | 0.00\% | $\begin{gathered} 0.00 \\ \% \end{gathered}$ | $\begin{gathered} 0.00 \\ \% \end{gathered}$ | 0.90\% | 0.27 |
| 2014 | 0.50\% | 8.40\% | 0.10\% | 0.00\% | 0.00 | 0.10\% |  | 0.20\% | 0.00\% | 0.00 | 0.00 | 0.40\% | 0.12 |


|  |  |  |  |  | \% |  |  |  |  | \% | \% |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2015 | 0.30\% | 18.60\% | 0.50\% | 0.10\% | $\begin{gathered} 0.10 \\ \% \end{gathered}$ | 0.20\% |  | 0.20\% | 0.00\% | $\begin{gathered} 0.00 \\ \% \end{gathered}$ | $\begin{gathered} 0.20 \\ \% \end{gathered}$ | 1.30\% | 0.20 |
| 2016 | 0.10\% | 18.50\% | 0.40\% | 0.10\% | $\begin{gathered} 0.00 \\ \% \end{gathered}$ | 0.30\% |  | 0.10\% | 0.00\% | $\begin{gathered} 0.00 \\ \% \end{gathered}$ | $\begin{gathered} 0.20 \\ \% \end{gathered}$ | 1.10\% | 0.15 |
| 2017 | 3.80\% | 9.00\% | 0.00\% | 0.00\% | $\begin{gathered} 0.00 \\ \% \end{gathered}$ | 0.00\% | 6.50\% | 0.10\% | 0.00\% | $\begin{gathered} 0.00 \\ \% \end{gathered}$ | $\begin{gathered} 0.00 \\ \% \end{gathered}$ | 6.60\% | 0.14 |
| 2018 | 7.30\% | 12.90\% | 0.60\% | 0.00\% | $\begin{gathered} 0.00 \\ \% \end{gathered}$ | 0.10\% | 8.10\% | 0.60\% | 0.00\% | $\begin{gathered} 0.00 \\ \% \end{gathered}$ | $\begin{gathered} 0.70 \\ \% \end{gathered}$ | 10.10\% | 0.28 |
| 2019 | 7.39\% | 22.92\% | 1.25\% | 0.00\% | $\begin{gathered} 0.17 \\ \% \end{gathered}$ | 0.68\% | 18.72\% | 0.46\% | 0.00\% | $\begin{gathered} 0.13 \\ \% \end{gathered}$ | $\begin{gathered} 0.00 \\ \% \end{gathered}$ | 21.40\% | 0.47 |
| 2020 | 7.38\% | 13.34\% | 0.59\% | 0.05\% | $\begin{gathered} 0.15 \\ \% \end{gathered}$ | 0.16\% | 5.35\% | 0.71\% | 0.00\% | $\begin{gathered} 0.00 \\ \% \end{gathered}$ | $\begin{gathered} 0.00 \\ \% \end{gathered}$ | 7.01\% | 0.24 |
| Avg. | 2.46\% | 23.06\% | 0.40\% | 0.18\% | $\begin{gathered} 0.03 \\ \% \end{gathered}$ | 0.28\% | 9.67\% | 0.64\% | 0.00\% | $\begin{gathered} 0.03 \\ \% \end{gathered}$ | $\begin{gathered} 0.08 \\ \% \end{gathered}$ | 4.21\% | 0.31 |

${ }^{\text {a }}$ Includes releases from Omak Pond and Bonaparte Pond. Three spring Chinook recovered in 2008 from an Omak Creek release were excluded from analysis.
${ }^{\mathrm{b}}$ Includes releases from Similkameen Pond
${ }^{\text {c }}$ Includes releases from Carlton Acclimation Pond and Goat Wall Acclimation
Pond
${ }^{d}$ Includes releases from Dryden Pond and Eastbank Hatchery
${ }^{e}$ Includes releases from Entiat NFH
${ }^{f}$ Includes releases from Chelan PUD Hatchery, Chelan River NFH, and Chelan Falls Hatchery
g Includes releases of summer Chinook from Wells Hatchery, Turtle Rock Hatchery, and Grant County PUD Hatchery
${ }^{\text {h }}$ Includes releases of fall Chinook from Hanford Reach
${ }^{i}$ Includes Releases from NPT Hatchery
j Includes releases from Marion Yakama Tribal, Cle Elum Hatchery, Irrigon, and Prosser Hatchery

Table 18. Number and percent (\%) of hatchery-origin Okanogan summer/fall Chinook that were recovered at target spawning areas or were captured at en route hatcheries (Wells and Chief Joseph Hatchery), and number and percent that strayed to non-target spawning areas and non-target hatcheries, brood years 1989-2015. As fish continue to return through time and the RMIS database is continually updated, reported data from recent brood years may change.

| Brood <br> Year | Target Stream |  |  |  | En Route <br> Hatchery |  | Non-target <br> Streams | Non-target <br> Hatchery |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Number | $\%$ | Number | $\%$ | Number | $\%$ | Number | $\%$ |  |
|  | 3,132 | $69.7 \%$ | 1,328 | $29.6 \%$ | 2 | $0.0 \%$ | 31 | $0.7 \%$ |  |
| 1990 | 729 | $71.4 \%$ | 291 | $28.5 \%$ | 0 | $0.0 \%$ | 1 | $0.1 \%$ |  |
| 1991 | 1,125 | $71.3 \%$ | 453 | $28.7 \%$ | 0 | $0.0 \%$ | 0 | $0.0 \%$ |  |
| 1992 | 1,264 | $68.5 \%$ | 572 | $31.0 \%$ | 8 | $0.4 \%$ | 1 | $0.1 \%$ |  |
| 1993 | 54 | $62.1 \%$ | 32 | $36.8 \%$ | 0 | $0.0 \%$ | 1 | $1.1 \%$ |  |
| 1994 | 924 | $80.8 \%$ | 203 | $17.7 \%$ | 16 | $1.4 \%$ | 1 | $0.1 \%$ |  |
| 1995 | 1,883 | $85.4 \%$ | 271 | $12.3 \%$ | 52 | $2.4 \%$ | 0 | $0.0 \%$ |  |
| 1996 | 27 | $100.0 \%$ | 0 | $0.0 \%$ | 0 | $0.0 \%$ | 0 | $0.0 \%$ |  |
| 1997 | 11,659 | $97.1 \%$ | 309 | $2.6 \%$ | 35 | $0.3 \%$ | 2 | $0.0 \%$ |  |
| 1998 | 2,784 | $95.4 \%$ | 102 | $3.5 \%$ | 31 | $1.1 \%$ | 2 | $0.1 \%$ |  |
| 1999 | 828 | $96.7 \%$ | 18 | $2.1 \%$ | 10 | $1.2 \%$ | 0 | $0.0 \%$ |  |
| 2000 | 2,091 | $93.8 \%$ | 29 | $1.3 \%$ | 94 | $4.2 \%$ | 15 | $0.7 \%$ |  |
| 2001 | 105 | $98.1 \%$ | 2 | $1.9 \%$ | 0 | $0.0 \%$ | 0 | $0.0 \%$ |  |
| 2002 | 702 | $96.2 \%$ | 17 | $2.3 \%$ | 11 | $1.5 \%$ | 0 | $0.0 \%$ |  |
| 2003 | 1,580 | $96.2 \%$ | 47 | $2.9 \%$ | 16 | $1.0 \%$ | 0 | $0.0 \%$ |  |
| 2004 | 4,947 | $94.4 \%$ | 206 | $3.9 \%$ | 85 | $1.6 \%$ | 2 | $0.0 \%$ |  |
| 2005 | 1,699 | $92.7 \%$ | 96 | $5.3 \%$ | 37 | $2.0 \%$ | 0 | $0.0 \%$ |  |
| 2006 | 5,162 | $97.6 \%$ | 60 | $1.1 \%$ | 67 | $1.3 \%$ | 0 | $0.0 \%$ |  |
| 2007 | 1,384 | $97.7 \%$ | 23 | $1.6 \%$ | 9 | $0.7 \%$ | 0 | $0.0 \%$ |  |
| 2008 | 3,577 | $96.8 \%$ | 95 | $2.6 \%$ | 20 | $0.6 \%$ | 4 | $0.1 \%$ |  |
| 2009 | 1,102 | $79.9 \%$ | 260 | $18.9 \%$ | 14 | $1.1 \%$ | 2 | $0.2 \%$ |  |
| 2010 | 927 | $43.4 \%$ | 648 | $54.1 \%$ | 9 | $0.4 \%$ | 10 | $2.1 \%$ |  |
| 2011 | 3,028 | $76.7 \%$ | 881 | $22.3 \%$ | 16 | $0.4 \%$ | 26 | $0.7 \%$ |  |
| 2012 | 478 | $72.8 \%$ | 174 | $26.5 \%$ | 4 | $0.6 \%$ | 1 | $0.2 \%$ |  |
| 2013 | 1,111 | $62.0 \%$ | 666 | $37.1 \%$ | 7 | $0.4 \%$ | 9 | $0.5 \%$ |  |
| 2014 | 566 | $71.9 \%$ | 201 | $25.7 \%$ | 8 | $1.0 \%$ | 11 | $1.4 \%$ |  |
| 2015 | 1,097 | $95.8 \%$ | 24 | $2.1 \%$ | 4 | $0.4 \%$ | 19 | $1.7 \%$ |  |
| Total | 53,965 | $83.9 \%$ | 7,008 | $14.9 \%$ | 555 | $0.9 \%$ | 138 | $0.4 \%$ |  |
|  |  |  |  |  |  |  |  |  |  |

Table 19. Number of estimated spawners and percent (\%) of spawning escapements comprised of hatchery-origin Okanogan summer/fall Chinook within non-target basins, return years 1994-2020.

| Return Year | Wenatchee |  | Methow |  | Chelan |  | Entiat |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Number | \% | Number | \% | Number | \% | Number | \% |
| 1994 | 0 | 0.00\% | 0 | 0.00\% | - | - | - | - |
| 1995 | 0 | 0.00\% | 0 | 0.00\% | - | - | - | - |
| 1996 | 0 | 0.00\% | 0 | 0.00\% | - | - | - | - |
| 1997 | 0 | 0.00\% | 0 | 0.00\% | - | - | - | - |
| 1998 | 0 | 0.00\% | 0 | 0.00\% | 0 | 0.00\% | 0 | 0.00\% |
| 1999 | 0 | 0.00\% | 0 | 0.00\% | 0 | 0.00\% | 0 | 0.00\% |
| 2000 | 0 | 0.00\% | 6 | 0.50\% | 30 | 6.40\% | 0 | 0.00\% |
| 2001 | 12 | 0.10\% | 0 | 0.00\% | 10 | 1.00\% | 0 | 0.00\% |
| 2002 | 0 | 0.00\% | 3 | 0.10\% | 4 | 0.70\% | 5 | 1.00\% |
| 2003 | 0 | 0.00\% | 8 | 0.20\% | 22 | 5.30\% | 14 | 2.00\% |
| 2004 | 0 | 0.00\% | 0 | 0.00\% | 5 | 1.20\% | 0 | 0.00\% |
| 2005 | 5 | 0.10\% | 27 | 1.10\% | 36 | 6.90\% | 7 | 1.90\% |
| 2006 | 0 | 0.00\% | 5 | 0.20\% | 4 | 1.00\% | 7 | 1.80\% |
| 2007 | 0 | 0.00\% | 3 | 0.20\% | 4 | 2.10\% | 0 | 0.00\% |
| 2008 | 0 | 0.00\% | 9 | 0.50\% | 46 | 9.30\% | 4 | 1.90\% |
| 2009 | 15 | 0.20\% | 3 | 0.20\% | 11 | 1.80\% | 18 | 9.90\% |
| 2010 | 5 | 0.06\% | 0 | 0.00\% | 32 | 2.48\% | 0 | 0.00\% |
| 2011 | 0 | 0.00\% | 0 | 0.00\% | 49 | 4.79\% | 0 | 0.00\% |
| 2012 | 7 | 0.09\% | 5 | 0.22\% | 17 | 0.36\% | 0 | 0.00\% |
| 2013 | 0 | 0.00\% | 0 | 0.00\% | 0 | 0.00\% | 0 | 0.00\% |
| 2014 | 0 | 0.00\% | 0 | 0.00\% | 0 | 0.00\% | 0 | 0.00\% |
| 2015 | 0 | 0.00\% | 0 | 0.00\% | 4 | 0.37\% | 0 | 0.00\% |
| 2016 | 0 | 0.00\% | 4 | 0.20\% | 4 | 0.35\% | 0 | 0.00\% |
| 2017 | 0 | 0.00\% | 0 | 0.00\% | 11 | 1.17\% | 0 | 0.00\% |
| 2018 | 0 | 0.00\% | 4 | 0.34\% | 4 | 0.53\% | 0 | 0.00\% |
| 2019 | 0 | 0.00\% | 0 | 0.00\% | 8 | 0.23\% | 0 | 0.00\% |
| 2020 | 0 | 0.00\% | 4 | 0.24\% | 0 | 0.00\% | 0 | 0.00\% |
| Total | 44 | 0.02\% | 81 | 0.15\% | 301 | 2.00\% | 55 | 0.80\% |
| 5-year <br> Total | 0 | 0.00\% | 12 | 0.16\% | 27 | 0.46\% | 0 | 0.00\% |

## Homing Fidelity within the Okanogan Basin

The 469 coded-wire tags recovered during spawning grounds surveys in fall of 2020 expanded to 814 and 1,471 spawners originated from Omak Pond and Similkameen Pond acclimation sites, respectively. The majority (83\%) of the spawners originating from the Omak Pond acclimation site spawned in the Okanogan River and $16 \%$ in the Similkameen River (Table 20). The Omak Pond fish tended to spawn in habitat downstream and upstream of the Omak Pond site, with the majority in reaches 03 (24\%) and 05 (32\%). Only Omak Pond CWT's were recovered below reach 03 (Figure 24). . Most fish acclimated at Similkameen Pond spawned in the Similkameen River (65\%) (Table 20). Of the Similkameen-origin fish that spawned in the Okanogan River, most used reaches 05 and 06 ( $32 \%$ combined; Figure 24). However, some of the CWT recoveries in reach 05 could have been fish that spawned upstream in S1 and swam or drifted downstream after spawning.

Table 20. Spawning distribution by river, for fish acclimated at Omak Pond and Similkameen Pond acclimation sites for 2018-2020.

| 2018 | Acclimation site (origin) |  |
| :---: | :---: | :---: |
| Spawning location | Omak <br> Pond | Similkameen <br> Pond |
| Okanogan River | $92 \%$ | $60 \%$ |
| Similkameen River | $8 \%$ | $40 \%$ |


| 2019 | Acclimation site (origin) |  |
| :---: | :---: | :---: |
| Spawning location | Omak <br> Pond | Similkameen <br> Pond |
| Okanogan River | $90 \%$ | $49 \%$ |
| Similkameen River | $10 \%$ | $51 \%$ |


| 2020 | Acclimation site (origin) |  |
| :---: | :---: | :---: |
| Spawning location | Omak <br> Pond | Similkameen <br> Pond |
| Okanogan River | $83 \%$ | $35 \%$ |
| Similkameen River | $17 \%$ | $65 \%$ |



Figure 24. 2020 spatial distribution of CJHP integrated program summer/fall Chinook spawners originally reared at the Similkameen Pond and Omak Pond acclimation sites and CJHP segregated program strays to Okanogan spawning grounds.

## Smolt Survival and Travel Time

Apparent survival of yearlings to RRJ in 2020 was $66 \%$ (SE 6\%) for the segregated program released from CJH and 56\% (SE 4\%) for integrated fish released from Omak Pond (Table 25). PIT tagged fish were not released from Similkameen pond in 2020. Apparent survival of yearlings to MCN was $22 \%$ (SE 5\%) for the segregated program released from CJH and $37 \%$ (SE 8\%) for the integrated fish released from Omak Pond (Table 25). Both programs had lower survival than the recent average and Carlton Pond (Table 26).

Apparent survival of subyearlings to RRJ in 2020 was $49 \%$ (SE 8\%) for the segregated program released from CJH and $45 \%$ (SE 5\%) for integrated fish released from Omak Pond (Table 25). Apparent survival of subyearlings to MCN was 23\% (SE 8\%) for the segregated program released from CJH and $27 \%$ (SE 8\%) for the integrated fish released from Omak Pond (Table 25). Both programs had very similar survival compared to their recent averages and lower survival than Wells Hatchery subyearlings (Table 26). The reduced survival compared to Wells Hatchery is to be expected considering the shorter migration distance; however Wells is the only other subyearling program in the area.

Table 21. Apparent survival estimates for PIT tagged summer/fall Chinook released in 2020 from Chief Joseph Hatchery (CJH), Omak Pond and other nearby hatcheries.

| Summer Chinook Release Group | \# PIT tags |  | Reach | Survival | Survival <br> Standard <br> Error (SE) | Capture Prob. | Capture Prob. (SE) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Released | Recap. |  |  |  |  |  |
| Yearlings released at CJH | 4081 | 923 | Release to RRJ | 0.66 | 0.06 | 0.34 | 0.03 |
|  |  | 104 | Release to MCN | 0.22 | 0.05 | 0.11 | 0.03 |
| Yearlings released at Omak Pond | 4532 | 802 | Release to RRJ | 0.56 | 0.04 | 0.31 | 0.03 |
|  |  | 129 | Release to MCN | 0.37 | 0.08 | 0.08 | 0.02 |
| Yearlings released at Similkameen Pond | 0 | 0 | Release to RRJ | NA | NA | NA | NA |
|  |  | 0 | Release to MCN | NA | NA | NA | NA |
| Yearlings released at Carlton Pond | 5052 | 2054 | Release to RRJ | 0.82 | 0.03 | 0.50 | 0.02 |
|  |  | 123 | Release to MCN | 0.60 | 0.14 | 0.04 | 0.01 |
| Yearlings released at Dryden Pond | 20725 |  |  |  |  |  |  |
|  |  | 577 | Release to MCN | 0.68 | 0.08 | 0.04 | 0.01 |
| Yearlings Released at Wells Hatchery | 4933 | 1409 | Release to RRJ | 0.82 | 0.05 | 0.35 | 0.02 |
|  |  | 117 | Release to MCN | 0.64 | 0.15 | 0.04 | 0.01 |
| Summer Chinook Release Group | \# PIT tags |  |  |  | Survival <br> Standard | Capture | Capture Prob. |
| Subyearlings released at CJH | 4785 | 378 | Release to RRJ | 0.49 | 0.08 | 0.16 | 0.03 |
|  |  | 38 | Release to MCN | 0.23 | 0.08 | 0.04 | 0.01 |
| Subyearlings released at Omak | 5085 | 449 | Release to RRJ | 0.45 | 0.05 | 0.20 | 0.02 |
|  |  | 60 | Release to MCN | 0.27 | 0.08 | 0.04 | 0.01 |
| Wells Fish Hatchery Subyearlings | 4976 | 796 | Release to RRJ | 0.59 | 0.05 | 0.27 | 0.02 |
|  |  | 58 | Release to MCN | 0.39 | 0.12 | 0.03 | 0.01 |
| Wild subyearlings from Col. R. | 17261 | 2386 | Release to RRJ | 0.43 | 0.03 | 0.32 | 0.02 |
|  |  | 165 | Release to MCN | 0.37 | 0.15 | 0.03 | 0.01 |

Table 22. PIT tag survival estimates for juvenile summer/fall Chinook from release to Rocky Reach and McNary dams from 2015 to 2020.

| Release Year | Summer Chinook Yearling Release Group |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Survival to Rocky Reach Dam |  |  |  |  |  |  |  | Survival to McNary Dam |  |  |  |  |  |  |  |
|  | CJH segr. |  | Omak Pond |  | Similk. |  | Carlton Pond |  | CJH segr. |  | Omak Pond |  | Similk. |  | Carlton Pond |  |
|  | Surv. | StdEr | Surv. | StdEr | Surv. | StdEr | Surv. | StdEr | Surv. | StdEr | Surv. | StdEr | Surv. | StdEr | Surv. | StdEr |
| 2015 | 0.71 | 0.04 | NA | NA | NA | NA | 0.63 | 0.02 | 0.68 | 0.14 | NA | NA | NA | NA | 0.55 | 0.10 |
| 2016 | 0.78 | 0.04 | 0.57 | 0.04 | NA | NA | 0.81 | 0.04 | 0.53 | 0.04 | 0.44 | 0.05 | NA | NA | 0.63 | 0.06 |
| 2017 | 0.77 | 0.06 | 0.80 | 0.06 | NA | NA | NA | NA | 0.82 | 0.14 | 0.63 | 0.10 | NA | NA | NA | NA |
| 2018 | 0.83 | 0.04 | 0.54 | 0.04 | NA | NA | 0.76 | 0.04 | 0.60 | 0.06 | 0.42 | 0.06 | NA | NA | 0.59 | 0.07 |
| 2019 | 0.67 | 0.04 | 0.69 | 0.03 | 0.63 | 0.03 | 0.79 | 0.04 | 0.45 | 0.10 | 0.50 | 0.08 | 0.53 | 0.10 | 0.56 | 0.11 |
| 2020 | 0.66 | 0.06 | 0.56 | 0.04 | NA | NA | 0.82 | 0.03 | 0.22 | 0.05 | 0.37 | 0.08 | NA | NA | 0.60 | 0.14 |
| Average | 0.74 |  | 0.63 |  | 0.63 |  | 0.76 |  | 0.55 |  | 0.47 |  | 0.53 |  | 0.59 |  |


| Release Year | Summer Chinook Sub-Yearling Release Group |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Survival to Rocky Reach Dam |  |  |  |  |  |  |  | Survival to McNary Dam |  |  |  |  |  |  |  |
|  | CJH segr. |  | Omak Pond |  | Wells Hatchery |  | Wild |  | CJH segr. |  | Omak Pond |  | Wells Hatchery |  | Wild |  |
|  | Surv. | StdEr | Surv. | StdEr | Surv. | StdEr | Surv. | StdEr | Surv. | StdEr | Surv. | StdEr | Surv. | StdEr | Surv. | StdEr |
| 2015 | 0.28 | 0.08 | 0.37 | 0.09 | 0.43 | 0.06 | 0.26 | 0.06 | 0.20 | 0.20 | 0.23 | 0.15 | 0.77 | 0.76 | NA | NA |
| 2016 | 0.44 | 0.08 | 0.35 | 0.05 | 0.51 | 0.05 | 0.24 | 0.03 | 0.14 | 0.05 | 0.14 | 0.06 | 0.25 | 0.05 | NA | NA |
| 2017 | 0.65 | 0.05 | 0.70 | 0.05 | 0.48 | 0.06 | 0.46 | 0.02 | 0.34 | 0.06 | 0.48 | 0.07 | 0.22 | 0.05 | 0.18 | 0.02 |
| 2018 | 0.65 | 0.06 | NA | NA | 0.79 | 0.07 | 0.44 | 0.04 | 0.53 | 0.09 | NA | NA | 0.53 | 0.11 | 0.12 | 0.03 |
| 2019 | NA | NA | NA | NA | 0.59 | 0.03 | 0.36 | 0.02 | NA | NA | NA | NA | 0.29 | 0.20 | 0.18 | 0.05 |
| 2020 | 0.49 | 0.08 | 0.45 | 0.05 | 0.59 | 0.05 | 0.43 | 0.03 | 0.23 | 0.08 | 0.27 | 0.08 | 0.39 | 0.12 | 0.37 | 0.15 |
| Average | 0.50 |  | 0.47 |  | 0.56 |  | 0.36 |  | 0.29 |  | 0.28 |  | 0.41 |  | 0.21 |  |

Wild subyearlings had a survival to RRJ of 43\% (SE 3\%) and 37\% (SE 15\%) to MCN (Table 21). The survival of wild summer Chinook from release to RRJ and MCN was higher than the recent average but, at least for estimates to McNary the uncertainty was relatively high due to the large standard error (Table 22).

Releases of yearling Summer Chinook smolts began on April 15, 2020. Of the 4,532 PIT tagged yearling summer Chinook released from Omak Pond (rkm 52), only 26 were detected at the Lower Okanogan PIT detection array. Fifty percent passed OKL within 3 days and $90 \%$ passed within 28 days. Travel time data for subyearlings revealed that 50\% had negative travel times (arriving at OKL before the release date), indicating that fish were released earlier than the reported release date or some other issues were present in the data set (as per reported on DART and PTAGIS). We had no way of figuring out the correct release date or travel time for those or what proportion of the release it affected, therefore travel times will not be reported subyearling summer Chinook in 2020. The mean travel time of yearling summer Chinook released from CJH facilities to RRJ in 2020 was very similar for CJH segregated yearlings ( 12 days; $10.1 \mathrm{~km} /$ day) and Omak Pond yearlings (13 days; $11.6 \mathrm{~km} /$ day) (Table 23). Although travel times and speeds could not be calculated for subyearlings due to the apparent early release of some portion of the programs, the $90 \%$ arrival dates at RRJ, MCN, and BON were 10-15 days longer than the 5 year average (Table 28). Travel times to RRJ, MCN and BON were 5-9 days shorter CJH and Omak yearling programs in 2020 compared to the five-year means (Table 29). The majority of yearling Summer Chinook from CJH and Omak Pond arrived at RRJ from late April to early May, with $90 \%$ passage dates of May 4 and May 6, respectively (Figure 25). The programs appeared to be successfully releasing actively migrating smolts and the migration speed increased substantially in reaches downstream of Rocky Reach Dam for all release groups (Table 23).

Table 23. Travel time (days), migration speed (km/day) and the number of days to $90 \%$ passage for summer/fall Chinook release groups in 2020.

| Release Group | Release timing | Release <br> Strategy | Mean Travel Time (days) |  |  | 90\% Passage (days) |  |  | Travel Rate (km/day) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Release to RRJ | Release to MCN | Release to BON | RRJ | MCN | BON | Release to RRJ | RRJ to <br> MCN | MCN to BON |
| CJH Summer subs | May 27 | Volitional | c | C | C | 56 | 61 | 62 | c | a | a |
| Omak Pond subs | May 27 | Forced | c | c | c | 51 | 58 | 61 | c | 29.1 | 71.6 |
| Wells FH subs | Apr 29-Jun 1 | ? | 28 | 38 | 47 | 47 | 57 | 59 | 2.4 | 34.2 | 69.1 |
| Wild subs | Jun 4-Jul 11 | NA | 49 | a | a | 54 | 64 | 51 | 2.0 | a | a |
| CJH Summer yearlings | Apr 15 | Volitional | 12 | 23 | 25 | 17 | 32 | 30 | 10.1 | 28.0 | a |
| Omak Pond yearlings | Apr 15 | Volitional | 13 | 24 | 25 | 19 | 29 | 30 | 11.6 | 31.2 | 61.4 |
| Carlton yearlings | Apr 20 | Forced | 18 | 26 | 33 | 26 | 37 | 42 | 7.1 | 29.0 | 61.4 |
| Dryden yearling | Apr 14 | Volitional | NA | 30 | 35 | NA | 42 | 46 | NA | 10.4 b | 51.2 |
| ${ }^{\text {a }}$ sample size too small ( $<10$ ) to calculate an estimate |  |  |  |  |  |  |  |  |  |  |  |
| ${ }^{\text {b }}$ Release to McNary, not Rocky Reach to McNary |  |  |  |  |  |  |  |  |  |  |  |
| ${ }^{\text {c }}$ Some negative travel time values were in the data, suggesting fish were released earlier than reported, or some escaped before the official release date |  |  |  |  |  |  |  |  |  |  |  |

Table 24. Travel time (days) and the number of days to $90 \%$ passage for subyearling summer/fall Chinook release groups from 2015 to 2020.

| Release Group | Year | Rocky Reach Dam |  | McNary Dam |  | Bonneville Dam |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Mean <br> Travel Time (d) | $90 \%$ <br> Passage <br> (d) | Mean <br> Travel Time (d) | 90\% <br> Passage <br> (d) | Mean <br> Travel Time (d) | 90\% <br> Passage <br> (d) |
| CJH Segregated Summer Subyearling | 2015 | 35 | 54 | 48 | 63 | 55 | 65 |
|  | 2016 | 18 | 31 | 27 | 38 | 31 | 44 |
|  | 2017 | 21 | 32 | 32 | 43 | 36 | 46 |
|  | 2018 | 15 | 32 | 27 | 46 | 30 | 43 |
|  | 2019 | NA | NA | NA | NA | NA | NA |
|  | 2020 | c | 56 | c | 61 | C | 62 |
|  | Average | 22 | 41 | 33 | 50 | 38 | 52 |
|  |  |  |  |  |  |  |  |
| Omak Pond Integrated Summer Subyearlings | 2015 | 27 | 44 | 40 | 52 | 45 | 57 |
|  | 2016 | 13 | 27 | 21 | 37 | 24 | 34 |
|  | 2017 | 14 | 22 | 24 | 33 | 28 | 37 |
|  | 2018 | NA | NA | NA | NA | NA | NA |
|  | 2019 | NA | NA | NA | NA | NA | NA |
|  | 2020 | c | 51 | C | 58 | C | 61 |
|  | Average | 18 | 36 | 28 | 45 | 32 | 47 |
|  |  |  |  |  |  |  |  |
| Wild Subyearlings | 2015 | 22 | 35 | 42 | 44 | a | a |
|  | 2016 | 28 | 55 | 35 | 59 | 36 | 69 |
|  | 2017 | 20 | 66 | 34 | 65 | 30 | 61 |
|  | 2018 | 31 | 56 | 44 | 71 | 45 | 53 |
|  | 2019 | 36 | 51 | 50 | 62 | 49 | 62 |
|  | 2020 | 49 | 54 | a | 64 | a | 51 |
|  | Average | 31 | 53 | 41 | 61 | 40 | 59 |

a) Sample size too small (<10) for a reliable estimate
c) Some negative travel time values were in the data, suggesting fish were released earlier than reported, or some escaped before the official release date

Table 25. Travel time (days) and the number of days to $90 \%$ passage for yearling summer Chinook release groups from 2015 to 2020.

| Release Group | Year | Rocky Reach Dam |  | McNary Dam |  | Bonneville Dam |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Mean <br> Travel Time (d) | 90\% <br> Passage <br> (d) | Mean <br> Travel Time (d) | 90\% <br> Passage <br> (d) | Mean <br> Travel Time (d) | 90\% <br> Passage <br> (d) |
| CJH <br> Segregated Yearlings | 2015 | 30 | 41 | 41 | 55 | 42 | 53 |
|  | 2016 | 15 | 26 | 25 | 36 | 28 | 42 |
|  | 2017 | 15 | 26 | 24 | 37 | 26 | 38 |
|  | 2018 | 13 | 27 | 24 | 36 | 29 | 50 |
|  | 2019 | 32 | 61 | 43 | 69 | 58 | 75 |
|  | 2020 | 12 | 17 | 23 | 32 | 25 | 30 |
|  | Average | 19 | 33 | 30 | 44 | 35 | 48 |
|  |  |  |  |  |  |  |  |
| Omak Pond <br> Integrated <br> Yearlings | 2015 | NA | NA | NA | NA | NA | NA |
|  | 2016 | 16 | 30 | 25 | 36 | 27 | 39 |
|  | 2017 | 22 | 37 | 30 | 44 | 32 | 44 |
|  | 2018 | 22 | 42 | 31 | 47 | 39 | 58 |
|  | 2019 | 23 | 44 | 36 | 62 | 47 | 68 |
|  | 2020 | 13 | 19 | 24 | 29 | 25 | 30 |
|  | Average | 19 | 34 | 29 | 44 | 34 | 48 |



Figure 25. Arrival timing at Rocky Reach Juvenile bypass (RRJ) of PIT tagged Summer Chinook released from the Chief Joseph Hatchery and Omak Pond in 2020.

## Smolt-to-Adult Return (SAR)

SAR was estimated using two methods, PIT tags and coded-wire tags.

PIT based estimate of SAR—The most recent brood year that could be fully assessed with PIT tags (through age 5) for SAR was 2015. For CJH segregated Summer Chinook from brood year 2015 (outmigration year 2017), 28 adult fish (age 4\&5) returned to Bonneville Dam with a PIT tag, resulting in SAR estimates of $0.6 \%$ before harvest and $0.7 \%$ with harvested fish added back in (Table 26). For brood year 2015, the SAR back to Wells Dam was $0.4 \%$ before harvest and $0.7 \%$ with harvested fish added back in (Table 26).

For the brood year 2014 integrated yearling program released from Omak Pond, 56 adult fish (age 4-5) returned to Bonneville Dam with a PIT tag, resulting in SAR estimates of $1.2 \%$ before harvest and $1.6 \%$ with harvested fish added back in (Table 26). For brood year 2015, the SAR back to Wells Dam was $0.9 \%$ before harvest and $1.5 \%$ with harvested fish added back in (Table 26).

The subyearling program showed considerably worse SARs, with no adult PIT tagged fish returning from the segregated program thus far, resulting in an SAR estimate of $0 \%$. For the brood year 2015 integrated sub yearling program at Omak Pond, zero adult fish returned in 2020 resulting in an SAR of 0\% (Table 27).

Table 26. Estimate of the smolt to adult return rate (SAR) for yearling Summer Chinook from Chief Joseph Hatchery and Omak Pond. Adult return data were available through 2020, therefore the most recent brood year that could be assessed through age 5 was 2015.

| CJH Segregated <br> Yearling Summer Chinook |  | PIT tag Detections at Bonneville Dam |  |  |  |  | Excluding Jacks |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |
| Brood Year | Number of PIT tags | Age 2 <br> Mini- <br> Jack | Age 3 | Age 4 | Age 5 | Age 6 | $\begin{aligned} & \text { Raw } \\ & \text { SAR } \\ & \hline \end{aligned}$ | Harvest Corrected SAR |
|  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |
| 2013 | 5017 | 17 | 16 | 28 | 24 | 0 | 1.0\% | 1.4\% |
| 2014 | 4951 | 1 | 7 | 35 | 29 | 0 | 1.3\% | 1.7\% |
| 2015 | 5024 | 27 | 3 | 18 | 10 | NA | 0.6\% | 0.7\% |
| 2016 | 4921 | 4 | 2 | 40 | NA | NA |  |  |
| 2017 | 4945 | 0 | 0 | NA | NA | NA |  |  |
|  |  | PIT Tag Detections at Wells Dam |  |  |  |  |  |  |
| 2013 | 5017 | 5 | 12 | 16 | 15 | 0 | 0.6\% | 1.0\% |
| 2014 | 4951 | 0 | 4 | 20 | 22 | 0 | 0.8\% | 1.4\% |
| 2015 | 5024 | 5 | 2 | 13 | 7 | NA | 0.4\% | 0.7\% |
| 2016 | 4921 | 2 | 1 | 24 | NA | NA |  |  |
| 2017 | 4945 | 0 | 0 |  |  |  |  |  |


| Integrated Yearling Summer Chinook from Omak Pond |  | PIT tag Detections at Bonneville Dam |  |  |  | Excluding Jacks |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Brood Year | Number of PIT tags | Age 2 <br> Mini- <br> Jack | Age 3 | Age 4 | Age 5 | Age 6 | $\begin{aligned} & \text { Raw } \\ & \text { SAR } \\ & \hline \end{aligned}$ | Harvest Corrected SAR |
| 2013 | 1204 | 0 | 0 | 0 | 0 | 0 |  |  |
| 2014 | 4193 | 28 | 4 | 19 | 9 | 0 | 0.7\% | 0.9\% |
| 2015 | 4830 | 4 | 8 | 22 | 34 | NA | 1.2\% | 1.6\% |
| 2016 | 5326 | 0 | 0 | 15 | NA | NA |  |  |
| 2017 | 4987 | 2 | 1 | NA | NA | NA |  |  |
| PIT Tag Detections at Wells Dam |  |  |  |  |  |  |  |  |
| 2013 | 1204 | 0 | 0 | 0 | 0 | 0 |  |  |
| 2014 | 4193 | 3 | 3 | 12 | 6 | 0 | 0.4\% | 0.7\% |
| 2015 | 4830 | 1 | 5 | 17 | 26 | NA | 0.9\% | 1.5\% |
| 2016 | 5326 | 0 | 0 | 11 | NA | NA |  |  |
| 2017 | 4987 | 2 | 0 | NA | NA | NA |  |  |

Table 27. Estimate of the smolt to adult return rate (SAR) for subyearling Summer Chinook from Chief Joseph Hatchery and Omak Pond. Adult return data were available through 2019, therefore the most recent brood year that could be assessed through age 5 was 2015.

| CJH Segregated Subyearling Summer |  | PIT tag Detections at Bonneville Dam |  |  |  | Excluding Jacks |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Brood Year | Number of PIT tags | Age 2 <br> Mini- <br> Jack | Age 3 | Age 4 | Age 5 | Age 6 | $\begin{aligned} & \text { Raw } \\ & \text { SAR } \end{aligned}$ | Harvest Corrected SAR |
| 2013 | NA | NA | NA | NA | NA | NA |  |  |
| 2014 | 4967 | 0 | 0 | 0 | 0 | 0 | 0.0\% | 0.0\% |
| 2015 | 4983 | 0 | 0 | 0 | 0 | NA | 0.0\% | 0.0\% |
| 2016 | 5029 | 0 | 0 | 7 | NA | NA |  |  |
| 2017 | 5027 | 1 | 17 | 0 | NA | NA |  |  |
| PIT Tag Detections at Wells Dam |  |  |  |  |  |  |  |  |
| 2013 | NA | NA | NA | NA | NA | NA |  |  |
| 2014 | 4967 | 0 | 0 | 2 | 0 | NA | 0.0\% | 0.1\% |
| 2015 | 4983 | 0 | 0 | 0 | 0 | NA | 0.0\% | 0.0\% |
| 2016 | 5029 | 0 | 0 | 5 | NA | NA |  |  |
| 2017 | 5027 | 0 | 0 | 0 |  |  |  |  |



CWT-based estimate of SAR—Based on expanded CWTs, the 2015 brood year had a SAR of $0.77 \%$, which was below the long-term and 5 -year averages. (Table 28).

Table 28. Smolt-to-adult return rate (SARs) for Okanogan/Similkameen summer/fall Chinook, brood years 1989-2015.

| Brood Year | Number of tagged smolts released ${ }^{\text {a }}$ | Estimated adult captures ${ }^{\text {b }}$ | SAR |
| :---: | :---: | :---: | :---: |
| 1989 | 202,125 | 4,293 | 2.1\% |
| 1990 | 367,207 | 972 | 0.3\% |
| 1991 | 360,380 | 975 | 0.3\% |
| 1992 | 537,190 | 2,282 | 0.4\% |
| 1993 | 379,139 | 117 | 0.0\% |
| 1994 | 212,818 | 1,526 | 0.7\% |
| 1995 | 574,197 | 2,842 | 0.5\% |
| 1996 | 487,776 | 32 | 0.0\% |
| 1997 | 572,531 | 18,570 | 3.2\% |
| 1998 | 287,948 | 7,742 | 2.7\% |
| 1999 | 610,868 | 2,782 | 0.5\% |
| 2000 | 528,639 | 6,765 | 1.3\% |
| 2001 | 26,315 | 424 | 1.6\% |
| 2002 | 245,997 | 1,979 | 0.8\% |
| 2003 | 574,908 | 3,503 | 0.6\% |
| 2004 | 676,222 | 12,960 | 1.9\% |
| 2005 | 273,512 | 1,662 | 0.6\% |
| 2006 | 597,276 | 13,605 | 2.3\% |
| 2007 | 610,379 | 4,943 | 0.8\% |
| 2008 | 516,533 | 14,894 | 2.9\% |
| 2009 | 522,295 | 7,119 | 1.4\% |
| 2010 | 610,927 | 10,666 | 1.7\% |
| 2011 | 625,234 | 18,757 | 3.0\% |
| 2012 | 157,390 | 3,643 | 2.3\% |
| 2013 | 677,483 | 5,580 | 0.82\% |
| 2014 | 749,546 | 6,047 | 0.81\% |
| 2015 | 474,928 | 3,667 | 0.77\% |
| Total | 12,459,763 | 158,347 | 1.27\% |
| 5-year Total | 536,916 | 7,539 | 1.40\% |
| ${ }^{\text {a }}$ Includes all tag codes and CWT released fish (CWT + Ad Clip fish and CWT-only fish). |  |  |  |
| ${ }^{\mathrm{b}}$ Includes estimated recoveries (spawning grounds, hatcheries, all harvest - including the ocean and Columbia River basin, etc.) and observed recoveries if estimated recoveries were unavailable. |  |  |  |

## DISCUSSION

## Juvenile Beach Seine

The CJHP took over the beach seining effort in 2014, adopting methods used by Douglas County PUD and Biomark in 2011-2013. Given the low catch rate of taggable summer/fall Chinook from the RST, beach seining appeared to be a more reliable opportunity to capture large numbers of taggable summer/fall Chinook juveniles. Again in 2018, PIT tags deployed at the beach seine far outnumbered tags deployed at the RST.

Mortality related to capture, handling and tagging was similar to what it has been in previous years. Maintaining water temperatures below $18{ }^{\circ} \mathrm{C}$, reducing MS-222 concentrations in the anaesthetizing solution, and further limiting handling time during tagging and capture likely contributed to this low pre- and post-tagging mortality. The hope for future years is to continue to reduce overall mortality associated with our PIT tagging efforts.

Fish size increased through the tagging period, but the number of fish captured and CPUE began to decrease in early July, which is later than what had been observed previous to that. Interestingly, dates of detection at downstream PIT arrays occurred about the same as they had in 2017 and 2019.

We do not have absolute certainty regarding natal stream for any of the juvenile Chinook fitted with a PIT tag, but assume the vast majority, especially of fish captured at the Gebber's location, are of Okanogan origin. However, juvenile summer Chinook in the Wells Pool originate from the Methow and Columbia Rivers as well. Therefore, future analyses of returning adults must recognize that some fish may not be destined for the Okanogan. Results from the stable isotope analysis conducted in 2018 indicated that most fish collected from the Gebber's location are of Okanogan River origin (See 2018 Annual Report, Appendix E).

## Lower Okanogan Adult Fish Pilot Weir

Discharge conditions on the Okanogan River in 2020 were quite a bit higher than those in previous years, restricting installation of the weir until mid- August, which was a month later than 2019. Temperatures on the Okanogan River were fairly normal, compared to the 13 year median. Temperature was not a factor for trapping operations once it began on August 27th. Tower observations were relatively low for the majority of the season outside of the last week in August. Observations of fish from the bank of the downstream pool increased after the water temperature stayed below $20^{\circ} \mathrm{C}$ in midSeptember. In September, fish observations 0.8 km . below the weir, at the lower pool, were higher than observations at the weir. When river temperature was lower and gage height was less than 4 feet, Chinook were more likely to mill in deeper pools. In previous years tower observations were much higher in September, so it's reasonable that there were more fish milling in the lower pool than there were milling around the weir in September. Continued monitoring of Chinook passage through the weir with respect to temperatures should continue in order to better refine weir operations and future expectations for weir effectiveness.

The number of Chinook handled at the weir $(\mathrm{n}=870)$ was more than the average ( $\mathrm{n}=468$ ). Configuration of the weir was similar to that in 2019 with the trap installed downstream, on the edge of the thalweg, and below the deep pool. The fish entrance chute was included with the trap gate again to test whether it would increase entrainment to the trap box. We evaluated the water conditions as it relates to discharge and stage height and think that we should continue to install the trap at the same location as 2020 to continue testing it with the chute.

None of the water quality parameters monitored were at a level that would cause concern regarding an environmental effect of the weir on water quality. The number (23) of dead fish at the weir was similar to other years with similar run sizes. There were no fish impinged between pickets (head upstream) in 2020.

There were thirteen sockeye trapped in 2020. When pickets were down and the trap was operating, there were no observations of jack or small adult Chinook escaping through the 2" weir panels, but we did observe several sockeye pass through the panels during the day. We will continue to use the 2 -inch weir panels again next year to increase the efficiency of Chinook trapping without causing too many sockeye to also use the trap.

There was no way to know exactly how many fish escaped past the weir before it was installed or how many fish swam through while the pickets were up or jumped over the sealing aprons after it was installed. The potential weir effectiveness measure of $4.3 \%$ was the third highest to date. Although the barrier broke down in late August,
this did not affect fish management objectives in 2020. With a higher adult return, CCT was able to collect their full brood stock quota (84) at the weir and remove about $4 \%$ of the hatchery-origin returns. In the future, with larger returns of hatchery fish due to CJH releases we anticipate a much higher pHOS at the weir resulting in higher weir effectiveness. Continuing these evaluations in future years will be critical to determining the long-term viability of the weir as a fish management tool for summer/fall Chinook.

In 2020 CCT F\&W staff were able to safely and successfully deploy, operate, and monitor the weir and add to the multi-year evaluation of the weir as a fish management tool for the CJH program. The program experience a larger run than the previous two years which allowed the program to successfully collect brood stock for the hatchery's integrated program and remove a portion of the hatchery-origin returns to manage pHOS. The weir's importance to successful management of the Okanogan summer/fall Chinook population should continue in the coming years with larger hatchery returns resulting from the increased production at CJH. Experiencing a broad range of environmental conditions spanning the extremely high summer flows of 2012 to the very low and warm flows in 2015 or high algal densities in 2019, is important for understanding the range of challenges and resulting weir effectiveness that can be expected through time.

## Redd Surveys

Summer/fall Chinook spawning consisted of 4,127 redds in 2020, which was above the long-term average (2,276 for 1989-2020) and above the more recent 5-year average of 3,421 . Redd counts were below average in only one reach in the Okanogan River (02) and above average in all other reaches in the Okanogan and Similkameen rivers (Table 15).

The redd count in reach 06 - which most years, supports the largest proportion of natural-origin spawners - was the fourth highest count on record, which dates back to 2006. Likewise, reach S1 in the Similkameen River - which generally supports the highest proportion of hatchery-origin spawners had its fourth highest redd count going back to 2006. These two adjacent reaches, along with reach still provide the primary spawning habitat for summer/fall Chinook in the Okanogan/Similkameen basin, comprising $82 \%$ of the total spawning in 2020. One objective of the CJHP is to increase the spatial distribution of spawning into the lower reaches of the Okanogan. Historically, a low proportion of the spawning activity has occurred in these reaches (01-04), likely due to lower quality spawning habitat (increased fine substrate, reduced gradient, increased pool habitat). The 2020 redd counts showed an increase in the proportion of redds in reaches 01 and 04. Although the changes are modest, they represent progress towards a goal that will likely take a long time to fully achieve. CJHP Chinook reared at the Omak Pond acclimation site (located around the break between reach 03 and 04) may be contributing to increased
spawning in lower reaches through natal homing. Continued monitoring of redd and carcass distribution will be critical to evaluate this metric.

Chinook spawning in the Okanogan generally begins as water temperatures drop below $15^{\circ} \mathrm{C}$. Conditions in 2020 were characterized with average discharge and increased stream temperatures going into the spawning period (Figure 2). However, despite these challenging holding conditions for Chinook, conditions improved (discharge increased and stream temperatures cooled to below average) at approximately the beginning of the spawning period (October). The greatest single week count of redds occurred between October 5 to October 11. Spawning lower in the Okanogan Basin (reaches 01, 02, and 03) appears to have peaked slightly later, with peak counts occurring the week of October 1925. Few redds were recorded in November, as most spawning was complete by then (Table 19). Although aerial surveys contribute a relatively small portion of the observed redds compared to ground or float surveys, they remain an important tool for documenting spawning, or lack of, in areas not accessible by ground crews.

As in previous years, the fish per redd expansion is based on the sex ratio of fish passing Wells Dam. This method has been used since at least 1998 (Hillman et al. 2014) and is still being applied to both the Methow and Okanogan Chinook populations. However, there is uncertainty that the combined sex ratio of hatchery- and natural-origin summer/fall Chinook at Wells Dam is representative of the Okanogan population because it also includes Methow returns, mainstem released hatchery-origin Chinook, as well as roaming downstream hatchery- and natural-origin Chinook. If the Okanogan has a different ratio of precocial males (jacks) than that of the Wells count, then the Okanogan abundance estimate could be biased. We suggest exploring other approaches to estimating the number of fish per redd in the Okanogan and Similkameen Rivers. Until then, the annual spawning escapement will continue to be calculated using the sex ratio of fish at Wells Dam.

## EsCAPEMENT INTO CANADA

Escapement of summer/fall Chinook into Canada had been largely overlooked until recent years, aided by video counts of Chinook passing over Zosel Dam. Spawning escapement to Canada has still been difficult to assess, as the video counts represent run escapement and the relationship between run escapement and spawn escapement is not clear. In 2018, video monitoring at Zosel Dam was discontinued, so we are now further limited in our ability to assess Chinook spawning escapement into Canada. In recent years, a substantial number of Chinook have been counted passing Zosel Dam, ranging from a low of 737 to a high of 2275 between 2013 and 2017 (Table 20), so there is the potential for Canada-bound Chinook to have a significant contribution to the trans-boundary Okanogan summer/fall Chinook population. No formal Chinook spawning grounds surveys are currently being conducted in Canada, but surveys for Sockeye (O. nerka) occur annually.

Biologists in Canada have observed small numbers (i.e., substantially fewer than the Zosel Dam video counts) of Chinook spawners building redds in the Canadian portion of the Okanogan River (R. Bussanich, ONA, pers. comm., 2014). There is a clear need for increased collaboration between agencies to better monitor and manage this trans boundary population. Researchers and managers for CCT and ONA have begun to discuss research and monitoring needs as well as potential strategies for accomplishing monitoring goals.

Research \& monitoring needs may include:

1. Organization of protocols and methods for formal Chinook spawning grounds surveys in Canada
2. Increased PIT array systems to better assess PIT-tagged fish passage into Canada

## Carcass Surveys

Spawning ground monitoring efforts resulted in an 23.6\% carcass recovery rate, which was well above the target carcass recovery rate of $20 \%$. However, it is unclear if $20 \%$ is necessary to obtain reliable biological-data or what the implications of reduced sampling rates may be. Zhou (2002) reported fish length as a significant factor in carcass recovery probability, with larger fish recovered at a higher rate than smaller fish. This is especially important as it relates to precocious males, or jacks, which are expected to occur with higher frequencies in hatchery-origin Chinook. Failing to assess and correct for biases and population discrepancies could lead to potential underestimation of hatchery-origin Chinook survival (resulting in inflated hatchery production) or over-estimation of wildorigin Chinook survival (masking potentially negative effects of the hatchery program) (Murdoch et al. 2010).

Spawning grounds surveys beginning in mid-August and lasting through November 8 revealed very few carcasses attributable to pre-spawn mortality, or PSM. Of the 283 female Chinook carcasses recovered, only $1.30 \%$ were determined to have expired prespawn. Also, few female carcasses had retained a significant portion of their eggs, with an egg retention rate of just $3.15 \%$. In other words, it appears that if a significant pre-spawn mortality event takes place, it occurs prior to the spawning period in October, or even late September, as the carcasses we recover on the spawning grounds are nearly all void of eggs. Given the challenging thermal conditions encountered by Chinook in the Okanogan River, it is likely that the majority of PSM occurs earlier in the season while water temperatures are higher and are a greater risk to fish attempting to travel to or hold near the spawning grounds. If this were true, the current design of our redd/carcass surveys would provide an underrepresentation of actual PSM. Therefore, egg retention and prespawn mortality results should be interpreted cautiously. One thing of note is that
carcasses that are collected during spring Chinook spawning ground surveys in August and September are assessed via coded wire tag recovery to determine spring or summer run. During the 2020 surveys, 4 summer Chinook pre-spawn mortality carcasses were collected, in the Similkameen River. The carcasses of fish that died prior to the onset of spawning and before sampling began may have been carried downstream of recovery floats, consumed by scavengers, or covered with sediment, making them unavailable for sampling or harder to detect and collect. This could result in an underestimation of pre-spawn mortality. The protocol assumes that each female may contain up to 5,000 eggs and were only considered pre-spawn mortality if they retained $>4500$ eggs. A static fecundity assumption may not be the best approach because younger and smaller females will likely have fewer eggs. We expanded the assessment to include an evaluation of fish that retained greater than 1,000 eggs as an attempt to capture some of the variability in fecundity and situations where fish died before depositing a biologically important portion of their eggs. However, even when considering any female that retained $\geq 1000$ eggs, the estimated PSM remained unchanged. We suggest continued review and modification of the egg retention estimation methods/protocol in the future.

## PHOS AND PNI

The biological target for CJHP is to maintain a 5-year average pHOS <0.3. 2015 was the first year since the CJHP began monitoring the population that the 5-year average (0.30) met this objective. 2020 pHOS ( 0.24 ) was below the biological target and the 5 -year average. The program met the biological target for PNI ( $>0.67$ ) in 2020 (0.82) after failing to meet it in 2018 and 2019. The 5-year mean PNI (0.75) remains above objective. There was a reduction in hatchery-origin spawners, including CJH segregated fish, on the spawning grounds in 2020. In the future, we suggest that continued aggressive removal of hatchery-origin fish through selective fisheries and adult management at the weir and hatchery ladder given the uncertainty regarding the adequacy of the objectives to meet long-term population conservation goals. Exceeding the targets whenever possible also provides a buffer for years when goals may not be achieved due to low run size or challenging environmental conditions.

## Origin of Hatchery Spawners

Hatchery-origin Chinook recovered on the spawning grounds in the Okanogan Basin were predominantly (75\%) from Okanogan and Similkameen acclimated, CJH Integrated Program releases. CJH Segregated fish made up 19\% of the hatchery-origin spawners, and $5 \%$ of the total spawning escapement. In order to stay under the $5 \%$ segregated pHOS goal on the spawning grounds there will need to be more removals of segregated Chief Joseph Hatchery fish before they reach the Okanogan. More aggressive operations of the Chief Joseph Hatchery ladder could help the program obtain this goal. Stray hatchery-origin fish
originating from outside the Okanogan made up $1.7 \%$ of the total estimated spawners, which was less than the goal of $5 \%$, although if we include the CJH segregated spawners, the stray rate increases to $7.0 \%$. Okanogan Basin hatchery-origin fish strayed to other areas at a low rate ( $0.4 \%$ to non-target basins and $1.7 \%$ to non-target hatcheries, based on RMIS queries of the 2015 BY ) and were a small percentage of the spawner composition in other Upper Columbia tributaries in 2020 (less than 2\% in any stray basin). Fish released within the Okanogan Basin have consistently homed to their natal stream, and 2020 was not an exception. One of the goals of the CJHP is to redistribute Chinook spawners to the middle and lower portion of the Okanogan River instead of inundating the already saturated Similkameen River with additional spawners. Juvenile Chinook releases from the Omak Pond acclimation site are primarily spawning in the Okanogan River ( $92 \%$ in 2018, $90 \%$ in 2019, and $83 \%$ in 2020) instead of the Similkameen River. Specifically, the Omak Pond-reared Chinook have spawned almost exclusively in the lower (03 reach) and middle ( 05 reach) sections of the Okanogan River.

## Smolt Survival and Travel Time

The survival results for each release group provide a useful index of annual survival for comparison between release groups and, in the future, between years. Targets for post release survival have not been established, but it was noteworthy that yearling survival from CJH and Omak Pond were $7-8 \%$ less than the recent average and $16-26 \%$ less than yearlings from Carlton Pond in 2020. This was a particularly surprising result considering that the travel times for CJH and Omak Pond were faster than normal. In the future, with more years of smolt migration data, the program should develop a statistical framework for evaluating smolt-to-smolt survival and establish targets that could be used to help adaptively manage the release strategies, if it is determined that survival or travel time are not adequate to meet program goals. Similar to previous years, the hatchery fish migrated out of the system relatively quickly in 2020 , with $90 \%$ passage at OKL within 28 days and only one detection after May 15. The assessment of sub-yearling travel time was invalid because half of the detections at OKL and some of the detections at RRJ occurred before the reported release date (5/27/2020). It is unclear if this was due to an undocumented early release or some other cause. Regardless, we were still able to evaluate the outmigration timing of subyearlings from the Okanogan and note that the last detection of a subyearling at OKL was June 11, 2020. These assessments suggest that the program was successful at releasing actively migrating smolts. This analysis did not attempt to account for detection probability at OKL and sample size was relatively small for the Omak Pond release ( $\mathrm{n}=26$ ). It is likely that the detection rate was different throughout the time period when smolts were detected. However, detection rates at large river arrays generally increase with decreased flow, so late arriving fish would have a better chance of being detected at OKL than fish out-migrating during high flows from April to June. Therefore, it is not likely that a meaningful number of late migrating smolts or residual hatchery fish would have crossed

OKL when compared to what was detected during peak migration. Although the OKL PIT detection site is 25 km from the confluence with the Columbia River, it is very close ( $\sim 2 \mathrm{~km}$ ) to the inundated zone of Wells Pool. Therefore, we can assume that smolts crossing OKL do represent fish leaving the Okanogan River system, or at least they are entering a more reservoir-like environment where interspecific competition for food and space is likely to be less than in the river. Unfortunately, it is not possible to evaluate juvenile outmigration (or movement within the Columbia River) in the winter months because juvenile bypass facilities do not operate year round.

## Smolt-TO-ADULT RETURN

The 2015 is the earliest brood year that a PIT-based estimate of SAR could be calculated. The data set for PIT-based estimates of SAR is too short to evaluate trends, but it was discouraging to see that the SAR for the segregated program dropped considerably for BY2015 compared to BY2013 and BY2014. Also discouraging was that there were zero returns of any PIT tagged subyearlings from BY2015. Clearly the poor ocean conditions in recent years have been affecting the program. It's unclear if CJH programs have been affected more or less than nearby programs. However, the fact that zero fish returned from the segregated subyearling program and only three adults returned from the integrated program suggests that PIT tags may not be a good tool for evaluating the SAR of subyearling Summer Chinook. PIT tagging resources may be better utilized increasing the sample size of yearling release groups. In future years the program will have more years of data to assess smolt to adult survival differences that can be used to provide insight on two options for the program: 1.) continue PIT tagging the subyearlings or 2.) rear fewer integrated subyearlings and, if possible, convert some of the integrated subyearlings to yearlings.

The CWT based SAR for the most recent full brood returns (2015) was less than the 5-year and long-term averages.

## Adaptive Management and Lessons Learned

## The Annual Program Review (APR)

Each year the CJHP hosts a workshop to review and present findings from the previous year and plan for the upcoming fish production and science monitoring cycle. The APR was convened in April 2021 with the purpose of reviewing data collection efforts and results from 2020 and developing the hatchery implementation and monitoring plan for 2021 (Figure 26). This effort is focused on using adaptive management to guide the program. After a series of presentations highlighting the data collection activities and results, the group (CJHP staff and invited guests from Federal, State, PUD, and other organizations) used the In-Season Implementation Tool (ISIT) during the "Analysis" step (Figure 27). The group reviewed the ISIT input parameters for key assumptions, status and trends and decision rules to be sure that the best available information was included in the model. ISIT then used the pre-season Upper Columbia summer/fall Chinook Salmon forecast to provide an estimate of how the program could be implemented with respect to broodstock collection, harvest, weir, and hatchery ladder operations to achieve biological targets for 2021. APR materials with more details than what is provided within this report can be found at https://www.cct-fnw.com/annual-program-review/.

## Key Management Questions

Answering key management questions is an essential function of the CJHP and is central to the analysis and reporting steps in both the APR and this annual report. Management questions inform the development of the RM\&E activities, the CJHPs Key Management Questions (KMQs) are:

1. What is the current status and recent historical trend of the naturally spawning population in terms of Viable Salmonid Population (VSP) parameters? ${ }^{7}$
2. What is the current status and recent historical trends for hatchery returns and harvest?
3. Is the hatchery program meeting target in-hatchery performance standards?
4. Are the hatchery post-release targets met for survival, catch contribution and straying?
5. Are targets for total catch contribution and selectivity for HORs met?
6. Are there negative effects of the hatchery on the natural population?
7. Are assumptions about natural production potential valid?

[^7]8. How should the program be operated in the coming year?

## Annual Planning Workflow



Figure 26. The Chief Joseph Hatchery's annual planning process and workflow.


Figure 27. The Chief Joseph Hatchery's analytical workflow.

## 2021 Run Size Forecast and Biological Targets

Run-size forecasts and updates are an early indicator for the biological targets for the coming season, through the Decision Rules outlined in the ISIT. The preseason forecast is based on brood year escapement and juvenile survival indicators and is generated through the Technical Advisory Committee (TAC) to the U.S. v. Oregon fish management agreement. As the season nears, this information is supplemented with return data from downstream dam counts. The pre-season forecast for Upper Columbia summer Chinook Salmon was 59,600. The pre-season forecast, and subsequent run updates from early dam counts, were used to predict the NOR and HOR run size for the Okanogan population. Hatchery broodstock and selective harvest targets are determined based on these estimates and the objectives for $\mathrm{pHOS}(<0.30)$ and $\mathrm{PNI}(>0.67)$. A regression analysis conducted within ISIT in preparation for the APR predicted that the pre-season forecast of 59,600 upper Columbia would yield 4,871 NORs and 4,674 HORs (Figure 28). The harvest and broodstock collection goals were established from this prediction. With a NOR run size just under 5,000 the broodstock collection recommendation for the integrated program was full production ( 791 NOB) with $100 \%$ pNOB (Figure 28). Likewise, the segregated program should achieve full production with 670 HOB. The model predicted that 1,465 HORs would be captured in the terminal (above Wells Dam) fisheries and that 66 HORs could be removed at the weir. These efforts could result in 3,511 NOS and 1,043 HOS for a pHOS of $19 \%$ and a PNI of 0.84 . Under this modeling scenario the biological targets will likely be met in 2021. As run size updates become available (through TAC) the ISIT outputs will be double checked until the final in-season check point on July 15, 2021. At that time the run size at Wells Dam will be input into ISIT and the final plan for broodstock and harvest will be updated. If the July 15 update includes more hatchery and natural fish than predicted, then harvest and removal of surplus fish at the weir and the hatchery ladder will be implemented by CCT and WDFW (through their mark-selective sport fishery). If the July 15 update includes less hatchery and natural fish than predicted, then CCT and WDFW will manage the harvest and removal of surplus fish in a way that will allow enough natural and hatchery-origin fish to escape to the Okanogan basin spawning grounds (NOS $\geq 5,250$, total escapement $\geq 7,500$ ) and also meet the pHOS objective of $<.30$

## ANNUAL MANAGEMENT TARGETS



Figure 28. The in-season updates management worksheet used to set biological targets for the upcoming year (2020) in the In-Season Implementation Tool.

## 2021 Key Assumptions

The CJHP reviews the key assumptions (working hypothesis) each year at the APR workshop. These assumptions directly affect the decision rules used to guide in-season management decisions. The program documents the changes and uses this information for future review and analysis (Figure 29).

KEY ASSUMPTIONS

| Natural Production |  | Biological |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Baseline | Targets | Transition 1 | Transition 2 | Long-term | Segregated Prog |
|  | 1307 |  | 1307 | 1307 | 1307 |  |
| Capacity (Smolts) | 3,672,603 |  | 3,672,603 | 3,672,603 | 3,672,603 |  |
| Juv Passage Survival | 27\% |  | 27\% | 27\% | 27\% |  |
| Ocean Survival (BON to BON) | 1.98\% |  | 1.98\% | 1.98\% | 1.98\% |  |
| Adult Passage Survival | 83\% |  | 83\% | 83\% | 83\% |  |
| Fitness | 0.87 |  | 0.87 | 0.87 | 0.87 |  |
| PNI | 0.72 | < 0.67 | 0.72 | 0.72 | 0.72 |  |
| Total pHOS | 38\% | $>30 \%$ | 38\% | 38\% | 38\% |  |
| Segr.pHOS | 2\% | < 5\% | 2\% | 2\% | 2\% |  |
| Ocean Harvest Rate | 24\% |  | 24\% | 24\% | 24\% |  |
| Lower Columbia Harvest Rate (Zones 1-6, Mouth to MCN) | 1\% |  | 1\% | 1\% | 1\% |  |
| Upper Columbia Harvest Rate (MCN to Wells) | 22\% |  | 22\% | 22\% | 22\% |  |
| Terminal Harvest Rate (Post Wells) | 3\% |  | 3\% | 3\% | 3\% |  |
| Natural Origin Spawners | 5,747 | < 5,250 | 5,747 | 5,747 | 5,747 |  |
| Hatchery Production |  |  |  |  |  |  |
| Local Brood | 398 |  | 791 | 791 | 791 | 275 |
| Yearling Release | 250,000 |  | 800,000 | 800,000 | 800,000 | 500,000 |
| Sub-yearling Release | 300,000 |  | 300,000 | 300,000 | 300,000 | 400,000 |
| SAR (yearling) | 1.66\% |  | 1.66\% | 1.66\% | 1.66\% | 1.66\% |
| SAR (sub-yearling) | 0.30\% |  | 0.30\% | 0.30\% | 0.30\% | 0.30\% |
| Return Rate to Okanogan | 89\% |  | 89\% | 89\% | 89\% | 20\% |
| pNOB | 100\% |  | 100\% | 100\% | 100\% |  |
| NOB | 709 |  | 709 | 709 | 709 |  |
| Relative Reproductive Success | 80\% |  | 80\% | 80\% | 80\% | 80\% |
| Ocean Harvest Rate | 24\% |  | 24\% | 24\% | 24\% | 24\% |
| Lower Columbia Harvest Rate (Zones 1-6, Mouth to MCN) | 3\% |  | 3\% | 3\% | 3\% | 3\% |
| Upper Columbia Harvest Rate (MCN to Wells) | 27\% |  | 27\% | 27\% | 27\% | 27\% |
| Pre-terminal Harvest Rate (Ocean to Wells) | 47\% |  | 47\% | 47\% | 47\% | 47\% |
| Terminal Harvest Rate (Post Wells) | 36\% |  | 36\% | 36\% | 36\% | 24\% |
| Hatchery Surplus | 362 |  | 362 | 362 | 362 | 2,721 |
| Average Terminal HOR Run | 2,211 |  | 7,075 | 7,075 | 7,075 | 4,422 |
| Expected HOS | 1,217 |  | 3,895 | 3,895 | 3,895 | 622 |
| Fisheries and Weirs |  |  |  |  |  |  |
| Weir Factor | 3\% |  | 3\% | 3\% | 3\% |  |
| NOR Harvest Release Mortality | 11\% |  | 11\% | 11\% | 11\% |  |

Figure 29. The key assumptions worksheet used in the 2021 In-Season Implementation Tool for the CJHP planning at the Annual Program Review

## 2021 Status and Trends

The recent performance of the population is a primary driver for determining how the hatchery program should be operated in the future. This was accomplished by updating and reviewing the status and trend information within five categories: (1) natural production, (2) hatchery production, (3) harvest, (4) migration, and (5) habitat (Figure 30).

| Return year |  | FPC Reported Dam Count at Wells thru 07/15 (excludes jacks) | $\begin{aligned} & \% \text { of } \\ & \text { final } \\ & \text { count } \end{aligned}$ | PUD Counts at WellsDam |  | Estimated Return of OkanoganOrigin Fish to Wells Dam |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | NOR All Origins (excludes jacks) | $\begin{gathered} \text { HOR All } \\ \text { Origins } \\ \text { (excludes } \\ \text { facks) } \end{gathered}$ | Okan. NORs | Okan. HORs | CJH HORs |
| 1998 | 3 | 1,060 | 0.25 | 970 | 5,519 | 841 | 833.44 |  |
| 1999 | 4 | 999 | 0.11 | 2,708 | 4,580 | 1,562 | 2,686 |  |
| 2000 | 5 | 2,266 | 0.26 | 2,726 | 7,398 | 1,213 | 2,291 |  |
| 2001 | 6 | 9,766 | 0.24 | 10,266 | 19,195 | 4,632 | 7,141 |  |
| 2002 | 7 | 23,221 | 0.34 | 24,138 | 42,035 | 5,207 | 11,801 |  |
| 2003 | 8 | 20,564 | 0.40 | 9,194 | 7,373 | 2,693 | 2,948 |  |
| 2004 | 9 | 14,762 | 0.40 | 23,227 | 13,989 | 8,004 | 2,599 |  |
| 2005 | 10 | 14,449 | 0.42 | 18,911 | 15,164 | 8,615 | 3,404 |  |
| 2006 | 11 | 12,563 | 0.43 | 20,262 | 8,730 | 10,047 | 2,749 |  |
| 2007 | 12 | 5,532 | 0.37 | 7,088 | 7,789 | 4,480 | 3,154 |  |
| 2008 | 13 | 8,838 | 0.35 | 11,244 | 13,779 | 4,337 | 6,554 |  |
| 2009 | 14 | 13,753 | 0.46 | 15,184 | 14,187 | 5,751 | 5,782 |  |
| 2010 | 15 | 12,264 | 0.41 | 5,671 | 7,167 | 4,791 | 5,409 |  |
| 2011 | 16 | 3,912 | 0.12 | 12,139 | 19,164 | 5,256 | 6,184 | - |
| 2012 | 17 | 10,082 | 0.24 | 14,424 | 27,716 | 5,974 | 7,793 |  |
| 2013 | 18 | 25,571 | 0.38 | 34,965 | 30,179 | 8,559 | 5,842 |  |
| 2014 | 19 | 26,010 | 0.46 | 36,060 | 21,015 | 12,803 | 4,251 |  |
| 2015 | 20 | 25,153 | 0.32 | 46,030 | 31,625 | 14,294 | 8,246 |  |
| 2016 | 21 | 21,479 | 0.43 | 28,467 | 21,542 | 12,065 | 4,766 | 3 |
| 2017 | 22 | 15,124 | 0.44 | 15,729 | 18,479 | 7,778 | 2,406 | 1,344 |
| 2018 | 23 | 11,886 | 0.48 | 6,533 | 18,347 | 3,730 | 2,398 | 2,212 |
| 2019 | 24 | 12,950 | 0.47 | 8,499 | 18,800 | 3,215 | 3,586 | 2,323 |
| 2020 | 25 | 25,965 | 0.46 | 22,243 | 36,309 | 9,271 | 3,500 | 1,763 |


|  | Terminal Harvest Above Wells |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Tribal Harvest |  |  |  |  |  | Recreational Harvest |  |  |  |  |  | Terminal Harvest Rates |  |  |
|  | $\begin{gathered} \text { Total } \\ \text { Tribal } \\ \text { Haryest } \end{gathered}$ | Total NORs | Total HORs | Okan. NORs | Okan. HORs | CJH HORs | Total <br> Rec. Harvest | Total NORs | Total HORs | Okan. NORs | Okan HORs | CIH HORs |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  | NOR | Int HOR | Seg HOR |
| 0.50 | - | 0 | 0 | - | - | - | - | - | - | - | - | - | 0\% | 0\% | - |
| 0.37 | . | 0 | 0 | . | . | . | . | - | . | . | - | - | 0\% | 0\% | - |
| 0.35 | . | 0 | 0 | - | - | - | . | - | . | . | - | - | 0\% | 0\% | - |
| 0.39 | . | 0 | 0 | - | . | - | . |  | - | - | - | - | 0\% | 0\% | - |
| 0.31 | 1,753 | 653 | 1100 | 118 | 990 | . | . | - | - | . | - | - | 2\% | 8\% | . |
| 0.48 | 2,130 | 785 | 1345 | 141 | 1,211 | - | . | - | - | - | - | - | 5\% | 41\% | - |
| 0.75 | 242 | 0 | 242 | - | 218 | - | 2,803 | 1,895 | 908 | 1,706 | 817 | - | 21\% | 40\% | - |
| 0.72 | 784 | 392 | 392 | 71 | 353 | - | 1,419 | 1,025 | 394 | 923 | 355 | - | 12\% | 21\% | - |
| 0.79 | 1,389 | 563 | 826 | 101 | 743 | - | 2,119 | 1,809 | 310 | 1,628 | 54 | - | 17\% | 29\% | - |
| 0.59 | 1,078 | 467 | 611 | 84 | 550 | - | 1,803 | 887 | 916 | 798 | 726 | - | 20\% | 40\% | - |
| 0.40 | 2,299 | 588 | 1711 | 106 | 1,540 | - | 1,665 | 698 | 967 | 628 | 561 | - | 17\% | 32\% | - |
| 0.50 | 2,598 | 363 | 2235 | 65 | 2,012 | . | 1,062 | 648 | 414 | 583 | 244 | - | 11\% | 39\% |  |
| 0.47 | 2,912 | 354 | 2558 | 64 | 2,174 | . | 1,019 | 612 | 407 | 551 | 208 | . | 13\% | 44\% | . |
| 0.46 | 1,097 | 449 | 648 | 81 | 577 | - | 1,017 | 200 | 817 | 180 | 286 | - | 5\% | 14\% | - |
| 0.43 | 3,184 | 656 | 2528 | 118 | 2,250 | . | 2,470 | 829 | 1,641 | 746 | 1,559 | - | 14\% | 49\% | - |
| 0.59 | 3,176 | 832 | 2344 | 150 | 1,781 | - | 2,107 | 179 | 1,928 | 161 | 713 | - | 4\% | 43\% | - |
| 0.75 | 2,963 | 1508 | 1455 | 271 | 1,164 | . | 1,383 | 321 | 1,062 | 289 | 382 | . | 4\% | 36\% | - |
| 0.63 | 9,729 | 6257 | 3472 | 1,126 | 2,639 | - | 1,660 | 289 | 1,371 | 260 | 494 | - | 10\% | 38\% | - |
| 0.72 | 3,141 | 1889 | 1252 | 340 | 989 | 3 | 1,784 | 237 | 1,547 | 213 | 665 | - | 5\% | 35\% |  |
| 0.67 | 1,397 | 746 | 651 | 134 | 143 | 91 | 1,568 | 591 | 977 | 532 | 479 | 59 | 9\% | 26\% | 11\% |
| 0.45 | 1,238 | 484 | 754 | 87 | 128 | 204 | 993 | 28 | 965 | 25 | 280 | 68 | 3\% | 17\% | 12\% |
| 0.35 | 1,363 | 129 | 1234 | 23 | 457 | 234 | 1,924 | 91 | 1,833 | 82 | 733 | 110 | 3\% | 33\% | 15\% |
| 0.64 | 1,731 | 95 | 1636 | 17 | 360 | 327 | 1,150 | 101 | 1,049 | 91 | - | - | 1\% | 10\% | 19\% |



Figure 30. The status and trends worksheet in the In-Season Implementation Tool for CJHP planning at the Annual Program Review.

## 2021 Decision Rules

The decision rules determine the targeted size of the hatchery program and the management of natural escapement abundance and composition. The purpose of the Decision Rules is to assure that the CJHP manages the hatchery, terminal fisheries, and weir to meet the guidelines for abundance, spawner composition, and distribution of the natural spawning escapement (Figure 31).

*Median, minimum and maximum values from 2021-2045 based on a single model run.
Figure 31. Screen shot of the decision rules in the In-Season Implementation Tool for CJHP planning at the Annual Program Review.

## Data Gaps and Research Needs

In a partnership with USGS, WDFW and the ONA, the CJHP is working to identify data gaps and applied research needs within the Okanogan Basin that would better inform hatchery management, increase available data for resource management decision making, and benefit overall salmonid recovery in the greater Columbia River basin. If funded in the future, the tasks identified could directly inform CJHP and other natural resource managers and aid in the decision-making process. Some of the data gaps and applied research needs that have been identified include:

1. Refined estimates (extent, fate, timing, and location) of summer/fall Chinook using the mainstem Columbia River above Wells Dam for spawning (i.e., straying), rather than returning to their natal Okanogan River using radio or acoustic telemetry.
2. Extent, fate, timing, and location of spawning Chinook in the Canadian portion of the Okanogan Basin.
3. Development and testing of a panel of microsatellites and/or single nucleotide polymorphisms (SNPs) for genotyping genetic stocks of Chinook salmon in the Okanogan Basin and upper-Columbia River, upstream of Wells dam, to identify and differentiate Okanogan summer- vs. fall- vs. spring-Chinook, as well as hatchery $\times$ hatchery, hatchery $\times$ wild, and wild $\times$ wild crosses of these various life-history types.
4. Utilization of advancements in thermal imaging/LiDAR or other remote sensing technologies combined with in-stream temperature loggers and ArcGIS/R Statistical Program (STARS \& FLoWs toolsets \& SSN package) to map current thermal refugia in the Okanogan basin and model potential changes resulting from climate change scenarios.
5. Development and/or adaptation of existing methods for better estimation of fine sediment loads per reach length in the Okanogan River to quantify effects on Chinook salmon spawning redds and productivity.
6. Design for testing fish tagging rate assumptions. PIT, radio, and genetic tagging emphasis.
7. Post-release mortality for various capture techniques including the purse seine, hatchery ladder, sport fishing, the weir, etc.

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## APPENDIX A: Summer/Fall Chinook

## Hatchery Operations and Production

The CJH's central facility is a 15-acre facility located immediately below Chief Joseph Dam along the right bank of the Columbia River at rkm 872 near Bridgeport, WA. There is one summer/fall Chinook CJH acclimation facility on the Okanogan River, Omak (rkm 51) acclimation pond. There is an additional acclimation facility on the Similkameen River (rkm 6.4) that is part of the CJH program but is operated by WDFW and funded by the CPUD.

Construction of the hatchery was completed in 2013 and broodstock were brought on station for the first time. The goal of the CJHP is to contribute to the increased abundance, productivity, temporal-spatial diversity, re-colonization of Chinook in the Okanogan Basin, and provide increased harvest for all fishers.

## Production Objectives

Full program production totals 2 million summer/fall Chinook. The summer/fall Chinook program incorporates both an integrated program ( 1.1 million smolts) supported by Okanogan River natural-origin broodstock and a segregated program (900,000 smolts) supported by hatchery-origin adults returning from the integrated program.

In 2020, the summer/fall Chinook program production level did not meet full production as planned, due to higher than expected pre-spawn mortality on both the integrated and segregated summer/fall brood as well as lower than anticipated fecundity.

## Summer/Fall Chinook Salmon

## BY 2019 Summer/Fall Chinook Salmon Rearing and Release

The yearling summer/fall Chinook rearing began February 19, 2020. Marking was completed, for both the integrated and the segregated programs, on July 28, 2020. The segregated summer/fall Chinook were $100 \%$ ad-clipped, with a 100 k CWT group tagged. The integrated summer/fall Chinook were $100 \%$ AD/CWT. As shown in Table A 1 and Table A 2, ponding and rearing mortality for the segregated program was elevated between Dec. 2020 and Feb. 2021 due to irritated gills and some delayed dropout from earlier issues. The segregated fish were marked into rearing Pond B, while the integrated fish were marked into the lower raceways, and reared until transfer to the acclimation ponds in late October. The segregated group was released on April 21st. Approximately 5,000 PIT tags were added to each group in October 2020. After subtracting shed tags and mortality, a total of 4,972 PIT tags were released from the segregated group (3,204 were detected at release).

Table A 1. Chief Joseph Hatchery brood year 2019 segregated summer/fall yearling rearing summary.

|  | Month | Total on hand | Mortality | Feed Fed | Fish per pound | Cumulative Survival (\%) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $N^{2}$ | 3/31/2020 | 601,237 | 2,584 | 142 | 888 | 99.57\% |
|  | 4/30/2020 | 596,069 | 5,168 | 886 | 494 | 98.72\% |
|  | 5/31/2020 | 595,117 | 952 | 1,723 | 139 | 98.56\% |
|  | 6/30/2020 | 594,955 | 162 | 1,285 | 113 | 98.53\% |
|  | 7/31/2020 | 594,690 | 265 | 2,014 | 87 | 98.49\% |
|  | 8/31/2020 | 594,495 | 195 | 2,346 | 57 | 98.46\% |
|  | 9/30/2020 | 620,624 | 450 | 3,191 | 41 | 98.45\% |
|  | 10/31/2020 | 620,115 | 509 | 3,016 | 35 | 98.37\% |
|  | 11/30/2020 | 619,486 | 629 | 3,476 | 25 | 98.27\% |
|  | 12/31/2020 | 595,272 | 24,214 | 9,460 | 21 | 94.43\% |
|  | 1/31/2021 | 584,080 | 11,084 | 8,032 | 21 | 92.67\% |
|  | 2/28/2021 | 570,910 | 13,170 | 8,844 | 14 | 90.58\% |
|  | 3/31/2021 | 568,911 | 1,999 | 8,448 | 14 | 90.26\% |
|  | 4/21/2021 | 568,625 | 286 | 1,546 | 14 | 90.22\% |
|  | Tot SEG | 568,625 | 61,667 | 54,409 | 14 | 90.22\% |

The integrated summer/fall Chinook were shipped to the Omak Acclimation Pond and the Similkameen Acclimation Pond between October 10th and October 29th. Reporting for the Similkameen Pond will reside with WDFW through release.

## Omak Acclimation Pond

On October 29, 2020 Chief Joseph Hatchery staff transferred 308,729 Integrated BY 19 summer/fall Chinook from Chief Joseph Hatchery to the Omak Acclimation Pond. Approximately 5,000 PIT tags were added to the group in October 2019. At the time of transfer, the fish were approximately 37 fpp , and were programmed to be reared over winter, with a target size at release of 10 fpp . These fish were forced released April 22, 2021. After subtracting shed tags and mortality, a total of 4,715 PIT tags were released from this integrated group ( 3,638 were detected at release). Table A 2 illustrates feed fed, feeding rate, and mortality to date for the integrated summer/fall Chinook transferred to the Omak Acclimation pond.

Table A 2. Omak Acclimation Pond BY 19 integrated yearling summer/fall Chinook rearing summary.

| Month | Total on hand | Mortality | Feed Fed | Fish per pound | Cumulative <br> Survival (\%) |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 10/31/2020 | 308,550 | 179 | - | 32 | 99.94\% |
| 11/30/2020 | 307,281 | 1,269 | 792 | 35 | 99.53\% |
| R. 12/31/2020 | 306,796 | 485 | - | 35 | 99.37\% |
| +o 1/31/2021 | 305,649 | 1,147 | - | 35 | 99.00\% |
| 2/28/2021 | 303,173 | 2,476 | 110 | 35 | 98.20\% |
| 3/31/2021 | 299,841 | 3,332 | 2,288 | 26 | 97.12\% |
| 4/22/2021 | 298,988 | 853 | 2,838 | 23 | 96.84\% |
| Cumulative: | 298,988 | 9,741 | 6,028 | 23 | 96.84\% |

## Riverside Acclimation Pond

Riverside Acclimation Pond was not used to rear BY 2018 summer/fall Chinook but was utilized to rear BY 19 10j Spring Chinook.

## Similkameen Acclimation Pond

Similkameen Pond was used to rear yearling summer/fall Chinook per the WDFW program funded by CPUD. Adult broodstock used to generate the juveniles for BY 2019 were collected via the CCT purse seine as part of the transition to the collaborative CJH program. On October 12, 2020, Chief Joseph Hatchery staff transferred 417,184 summer/fall Chinook to the Similkameen Pond, with the assistance of WDFW's Eastbank Hatchery staff. At the time of transfer, the fish were approximately 37 fpp , and were programmed for over winter acclimation, with a target size at release of 10 fpp . These fish began volitional release on April 15th, with an end release date of April 30, 2021. Cumulative survival, at the date of transfer, was $96.55 \%$. Survival from transfer to release was $96.84 \%$.

## Cumulative egg to smolt survival

The target egg to smolt survival identified in the original summer/fall Chinook HGMP was $77.5 \%$ for sub-yearlings and $73.5 \%$ for yearlings (CCT 2008b). The cumulative egg to smolt survival, for the BY 2019 yearlings, was 84.5\% (83.1\% for the segregated program and $85.8 \%$ for the integrated program.

## BY 2020 Summer/FALL Chinook SALMON

## 2020 Broodstock collection

Collection of summer/fall Chinook for BY 2020 occurred between July 2nd and August 17th via the CCT purse seine operation at the mouth of the Okanogan River. Both hatchery-origin and natural-origin brood were collected to supply the integrated and segregated production programs at CJH. As the seine was being pursed, 9-meter transport barges approached the seine vessel and tied off on the opposite side. The broodstock transport barges have two transport tanks, a 300 gallon for HORs and a 600 gallon for NORs. Brood fish were removed from the seine and placed headfirst in a rubber tube, or boot, containing some water and handed to the staff on the barges for placement in the holding tanks. A maximum of 14 HOR and 28 NOR brood could be loaded per barge. Once full, or at the commencement of the purse seine haul, the barges returned to the offload area at Mosquito Park approximately 2 km away. The brood was then removed from the tanks by hand, placed into a boot, then delivered to one of two 2,500 gallon tanker trucks and transported 16 km to the hatchery.

Water temperatures were of major concern during these operations and monitored to minimize trauma to the adult brood. Okanogan River temperatures during July ranged from $67.5^{\circ} \mathrm{F}\left(19.7^{\circ} \mathrm{C}\right)$ to $78.4^{\circ} \mathrm{F}\left(25.8^{\circ} \mathrm{C}\right)$. In order to limit the effects of the temperature changes we monitored the temperature of all transport vessels and strived to not expose brood to changes greater than $8^{\circ} \mathrm{F}$. We accomplish this by utilizing both well water and surface water when filling the barges and transport tankers, and monitoring our raceway temperatures.

A weekly quota was developed to ensure that brood collections occurred across as much of the summer/fall run timing as possible (Table A 3). If brood collection failed to meet the weekly quota it was adjusted the following week. However, due to low returns and to ensure overall broodstock goals were met, this quota was not followed and broodstock was collected as early as possible. The purse seine is only effective when there is a thermal barrier at the mouth of the Okanogan, therefore broodstock can only be collected there until late August or early September. Once at the hatchery, broodstock were offloaded 6 at a time into totes in order to inject with Draxxin and LA200 (liquamycin), with females receiving both while males only receiving LA200. Broodstock were then separated by program and sex and put into their designated raceways. The receiving water was approximately $57^{\circ} \mathrm{F}$. The adult ponds had a flow rate of 500 gpm , and an exchange rate of 54 minutes, representing a Flow Index (FI) of 0.56 and a Density Index (DI) of 0.08 at max capacity. Upon arrival, adult ponds were put on well water.

All adult ponds were treated a minimum of three days per week with formalin to control fungus at a rate of 1:6000, for one exchange. Additionally, brood fish were treated
twice per week with Chloramine-T at 12 ppm for one exchange to control Columnaris bacteria. Diquat was also used under an INAD once Columnaris was detected in broodstock, which was from mid-August through October.

Table A 3. Chief Joseph Hatchery summer/fall Chinook weekly broodstock collection objectives and results for brood year 2020.

## Chief Joseph Hatchery BY 20 Summer Chinook Weekly Broodstock Collection Objectives

|  | Weekly Quota $^{1}$ |  |  | Cumulative Collection |  |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Week | Natural $^{\text {Origin }^{2}}$ | Hatchery $^{\text {Origin }^{3}}$ | Cumulative <br> Proportion | Natural <br> Origin | Hatchery <br> Origin |
| July 6 - July 12 | 22 | 22 | 0.04 | 22 | 22 |
| July 13 - July 19 | 22 | 22 | 0.08 | 44 | 44 |
| July 20 - July 26 | 108 | 104 | 0.27 | 152 | 148 |
| July 27 - Aug 2 | 108 | 104 | 0.46 | 260 | 252 |
| Aug 3 - Aug 9 | 132 | 126 | 0.69 | 392 | 378 |
| Aug 10 - Aug 16 | 132 | 126 | 0.92 | 524 | 504 |
| Aug 17 - Aug 23 | 36 | 36 | 0.98 | 560 | 540 |
| Aug 24 - Aug 30 | 12 | 12 | 1.00 | 572 | 552 |
| *Sept 15 - Oct 15 | 84 |  |  | 656 |  |

*NOR weir collection
${ }^{1}$ Weekly collection short-fall to be added to following week's collection
${ }^{2}$ Combined collection strategies in priority order: purse seine, tangle-net, Okanogan weir beach seine and CJH ladder
${ }^{3}$ Combined collection strategies in priority order: purse seine, tangle-net, CJH ladder, Okanogan weir and beach seine

A total of 611 HOB were collected including 270 females, 277 adult males and 64 jacks (Table A 4). However, it's believed some of these were mis-sexed as they entered the hatchery, so these numbers do not match exactly to the table. A total of 676 NOB were collected including 324 females, 330 adult males, and 22 jacks (Table A 4). Some of the fish initialed classified as jacks were actually adult males, thus the difference from Table A 5.

Through the month of November 2020, there were 75 adult male and 32 female mortalities in the HOR brood, representing $72.2 \%$ and $88.4 \%$ cumulative pre-spawn survival to date, respectively. For the same time frame, 93 adult NOR summer/fall Chinook males died, and 43 females died, representing $71.3 \%$ and $87.0 \%$ cumulative pre-spawn survival, respectively. (Table A 4) Brood fish, particularly males, suffered higher than anticipated mortality due to Columnaris disease, which affected us particularly hard once the well water in which these fish are held reached $>60^{\circ} \mathrm{F}$.

The cumulative pre spawn holding survival, for all summer/fall brood collected, was 81.2\% (including jack) for HOB and 79.3\% (including jacks) for NOB (Table A 4); neither program achieving program survival goal of $90 \%$.
Table A 4. Chief Joseph Hatchery summer/fall Chinook Hatchery (HOB) and Natural (NOB) origin broodstock holding survival summary for brood year 2020. ( $\mathrm{M}=$ adult males, $\mathrm{J}=$ jacks and $\mathrm{F}=$ adult females). The survival standard for this life stage was $90 \%$.

|  | Month | Beginning of Month |  |  | End of Month |  |  | Mortality |  |  | Monthly Survival (\%) |  |  | Cumulative Survival (\%) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | M | J | F | M | J | F | M | J | F | M | J | F | M | J | F |
| $\mathrm{p}^{2}$ | July | 0 | 0 | 0 | 270 | 64 | 175 | 0 | 0 | 0 | 100.0\% | 100.0\% | 100.0\% | 100.0\% | 100.0\% | 100.0\% |
|  | August | 270 | 64 | 175 | 268 | 64 | 277 | 2 | 0 | 0 | 99.3\% | 100.0\% | 100.0\% | 99.3\% | 100.0\% | 100.0\% |
|  | Sept | 268 | 64 | 277 | 268 | 64 | 277 | 0 | 0 | 0 | 100.0\% | 100.0\% | 100.0\% | 99.3\% | 100.0\% | 100.0\% |
|  | Oct | 268 | 64 | 277 | 12 | 7 | 9 | 68 | 8 | 32 | 74.6\% | 87.5\% | 88.4\% | 74.1\% | 87.5\% | 88.4\% |
|  | Nov | 12 | 7 | 9 | 0 | 0 | 0 | 5 | 0 | 0 | 58.3\% | 100.0\% | 100.0\% | 72.2\% | 87.5\% | 88.4\% |
|  | Total | 270 | 64 | 277 | 12 | 7 | 9 | 75 | 8 | 32 |  |  |  | 72.2\% | 87.5\% | 88.4\% |
| $a^{p}$ | July | 0 | 0 | 0 | 249 | 22 | 220 | 0 | 0 | 0 | 100.0\% | 100.0\% | 100.0\% | 100.0\% | 100.0\% | 100.0\% |
|  | August | 249 | 22 | 220 | 271 | 22 | 290 | 1 | 0 | 3 | 99.6\% | 100.0\% | 99.0\% | 99.6\% | 100.0\% | 99.0\% |
|  | Sept | 271 | 22 | 290 | 322 | 22 | 327 | 1 | 0 | 0 | 99.7\% | 100.0\% | 100.0\% | 99.4\% | 100.0\% | 99.1\% |
|  | Oct | 322 | 22 | 327 | 63 | 14 | 62 | 75 | 2 | 30 | 76.7\% | 90.9\% | 90.8\% | 76.2\% | 90.9\% | 90.0\% |
|  | Nov | 63 | 14 | 62 | 0 | 0 | 0 | 16 | 2 | 10 | 74.6\% | 85.7\% | 83.9\% | 71.3\% | 81.8\% | 87.0\% |
|  | Total | 324 | 22 | 330 | 0 | 0 | 0 | 93 | 4 | 43 |  |  |  | 71.3\% | 81.8\% | 87.0\% |

Hatchery staff began collection of NOR brood from the weir on August 31, 2020 and concluded on September 22, 2020 with 43 wild males and 41 wild females caught and transferred to the hatchery for broodstock. Fish were transferred from the weir trap manually.

The fish were then transported approximately 32 km to Chief Joseph Hatchery where they were held in the broodstock raceways until the first spawn date the first week in October. We recognize that fish collected late may have arrived at any point in their run timing; however, the efforts to collect fish in late August into September at least offer the opportunity to include fish that arrive later in the run timing.

## Spawning

Spawning of summer/fall Chinook began on October 6, 2020 with the segregated program, and continued through November 4, 2020 for segregated program and November 10,2020 for the integrated program. Beginning with the 2018 brood year, the segregated and integrated programs will be spawned on separate days. As with the spring Chinook, the summer/fall Chinook program is also 100\% ELISA sampled. For the 2020 brood, there were no eggs culled for either program in 2020.

Total NOB spawned included 234 males, 10 jacks, and 281 females. (Table A 5) Total HOR spawn included 221 males, 25 jacks, and 245 females. Total eyed egg take for the season was 1,639,336. Egg survival from green egg to eyed egg for NOB averaged 80.4\% (Table A 5). Egg survival for HOB averaged 81.4\%. Survival was lower than the key assumption of ( $90 \%$ ) for this life stage.

Table A 5. Chief Joseph Hatchery brood year 2020 summer/fall Chinook spawning and incubation results.

| $\lambda^{0}$ | Spawn Date | Total Adults Spawned |  |  | Eyed Eggs On <br> Hand | Mortality (Pick off) | Culledeggs | Adjusted Total Egg Take | Cumulative <br> Survival (\%) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | M | J | F |  |  |  |  |  |
|  | 10/6/2020 | 21 | 0 | 21 | 74,615 | 2,321 | - | 76,936 | 97.0\% |
|  | 10/13/2020 | 49 | 2 | 51 | 147,800 | 33,212 | - | 181,012 | 81.7\% |
|  | 10/20/2020 | 83 | 12 | 94 | 305,617 | 58,052 | - | 363,669 | 84.0\% |
|  | 10/27/2020 | 58 | 10 | 70 | 188,613 | 63,331 | - | 251,944 | 74.9\% |
|  | 11/4/2020 | 10 | 1 | 9 | 16,588 | 10,456 | - | 27,044 | 61.3\% |
|  | Total: | 221 | 25 | 245 | 733,233 | 167,372 | - | 900,605 | 81.4\% |
|  |  |  | ults |  | Eyed Eggs On | Mortality |  | Adjusted Total | Cumulative |
|  | Spawn Dat | M | J | F | Hand | (Pick off) |  | Egg Take | Survival (\%) |
|  | 10/6/2020 | 10 | 0 | 10 | 39,419 | 5,063 | - | 44,482 | 88.6\% |
|  | 10/14/2020 | 39 | 2 | 41 | 137,724 | 29,373 | - | 167,097 | 82.4\% |
|  | 10/21/2020 | 79 | 4 | 83 | 267,300 | 62,459 | - | 329,759 | 81.1\% |
|  | 10/28/2020 | 53 | 0 | 104 | 327,275 | 93,081 | - | 420,356 | 77.9\% |
|  | 11/4/2020 | 41 | 1 | 38 | 118,530 | 27,129 | - | 145,659 | 81.4\% |
|  | 11/10/2020 | 12 | 3 | 5 | 15,855 | 4,163 | - | 20,018 | 79.2\% |
|  | Total: | 234 | 10 | 281 | 906,103 | 221,268 | - | 1,107,353 | 80.4\% |

## Broodstock origin

Broodstock were interrogated for coded-wire tags by program throughout October and the first week of November. Beginning October 6 th , segregated fish were spawned on Tuesday of each week, while integrated fish were spawned on Wednesdays. A total of eight spawning events occurred in 2020. All ad-clipped chinook incorporated in the integrated and segregated programs were sampled at $100 \%$ for CWTs regardless of electronic detection via T-wand. Collected samples were then frozen until mid-December where CWTs are extracted and analyzed in the laboratory during winter months. .

All broodstock collected for the summer/fall Chinook segregated program came from an Upper Columbia River hatchery program. The CJH integrated program was the largest contributor to segregated brood with ( $\mathrm{n}=326$ ) 67\% of adults coming from either the Similkameen or Omak Pond (Table A 6). Other Upper Columbia River Hatcheries contributed ( $\mathrm{n}=92$ ) 19\%, most of which were from DCPUD releases near Wells, (6.1\%), Chelan Falls, (5.7\%) and Wells Hatchery (5\%). A portion of snouts ( $n=52$ ) indicated detection during spawning events but a coded-wire tag was not found during extraction. Reasons for this include but are not limited to rapidly shaking a Northwest Marine Technologies (NMT) T-Wand when scanning for a CWT (false positive in the field), failure to detect a tag in the lab (false negative), metals in the soil that transfer to a fish during handling or hooks or other metal debris in the fish's head. The unknown component represents $10.7 \%$ of the 2020 segregated brood (Table A 6). The mark rate for brood year 2013-2017 segregated releases vary between 99\% and 24\% ad-clipped + CWT however, overall mark rate between these brood years averages 68.5\%. All summer/fall Chinook programs upstream of Priest Rapids Dam are expected to have a 100\% tag rate (except for CJH segregated). We would expect a portion of no CWT detection in the lab with the CJH segregated adult returns after adjusting for tag loss, the number of estimated non-CWT recoveries ( $\mathrm{n}=52$ ) can be assigned to the segregated CJH program. The overall composition of the segregated program (tagged and non-tagged) to the segregated brood was $14 \%$.

Table A 6. Composition of hatchery-origin brood, by program, collected for the CJH segregated program in 2020.

| Category | Hatchery Program | Brood | \% of Brood |  |
| :---: | :---: | :---: | :---: | :---: |
| Okanogan Integrated | Similkameen <br> Omak Pond | $\begin{aligned} & 196 \\ & 130 \end{aligned}$ | $\begin{aligned} & 40.3 \text { \% } \\ & \\ & 26.7 \% \end{aligned}$ | 67\% |
| CJH Segregated | Chief Joseph <br> Chief Joseph (nontagged) | $16$ $52$ | $\begin{aligned} & 3.3 \% \\ & 10.7 \% \end{aligned}$ | 14\% |
| Other UCR summer/fall Chinook hatchery | Chelan Falls <br> DCPUD <br> Entiat <br> WDFW <br> Wells | 28 <br> 30 <br> 5 <br> 4 <br> 25 | 5.7\% <br> 6.1\% <br> $1 \%$ <br> <1\% <br> 5\% | 19\% |
| Total |  | 486 | 100 |  |

*Brood values are adjusted to account for segregated no tag fish and are rounded to the nearest whole number.

## Integrated Program Broodstock Age Structure

Scales are taken from summer/fall Chinook integrated program broodstock in order to capture the age of successfully spawned fish. In 2020, the integrated and segregated programs were comprised of mostly four and five-year old male and female fish (Figure A 1).


Male Integrated Summer Chinook Broodstock Salt Age


Female Integrated Summer Chinook Broodstock Total Age


Figure A 1. The total and salt ages of the 2020 broodstock, males and females, collected for the Okanogan summer/fall Chinook integrated program.

## Segregated Program Broodstock Age Structure

Coded wire tags are extracted from summer/fall Chinook segregated program broodstock and later read in order to capture the age of successfully spawned fish (Figure A 2).


Male Segregated Summer Chinook Broodstock Salt Age



Figure A 2. The total and salt ages of the 2020 broodstock, males and females, collected for the Chief Joseph Hatchery segregated program.

## Incubation

Eggs from each female summer/fall Chinook were placed in individual incubators (Heath Trays) and remained individually incubated until ELISA results were obtained. Once eye-up occurred, eggs from any moderate and high ELISA would be removed; No eggs were discarded from either program. The cull rate for this production plan allows for a rate of $5 \%$ for segregated and $3 \%$ for integrated. After eye-up, egg mortality was removed, and the eggs were inventoried and put back into their individual trays for hatching. Incubation water temperatures were initially manipulated to the level necessary to synchronize the hatching and ponding of the spawn takes throughout October and November 2020 and to achieve the size-at-release target for both yearling and sub-yearling summer/fall Chinook programs.

## Rearing

The first group of brood year 2020 sub-yearlings were brought out of incubation and transferred into early rearing troughs in early February 2021 (Table A 7). During this time, the group was introduced to feed in the early rearing troughs and reared for a period of two weeks. After the initial rearing period inside, they were transferred outside to the standard raceways via the fry transfer line. No inventories were taken during transfers, to prevent excess handling stress.

The first group of integrated yearlings were brought out of incubation and transferred into early rearing troughs in late March 2021 while the first group of segregated yearlings was brought out in early April 2021 (Table A 8). During this time, the group was introduced to feed in the early rearing troughs and reared for a period of two weeks. After the initial rearing period inside, they were transferred outside to the standard raceways via the fry transfer line. No inventories were taken during transfers, to prevent excess handling stress.

Table A 6. Chief Joseph Hatchery brood year 2020 summer/fall Chinook sub-yearling rearing summary.

|  | Month | Total on hand | Mortality | Feed Fed | Fish per pound | Cumulative <br> Survival(\%) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\mathbf{N}^{2 R}$ | 3/31/2021 | 180,061 | 3,396 | 188 | 448 | 98.1\% |
|  | 4/30/2021 | 178,715 | 1,346 | 700 | 137 | 97.4\% |
|  | 5/31/2021 | 178,366 | 349 | 594 | 97 | 97.2\% |
|  | Subtotal: | 178,366 | 5,091 | 1,482 | 97 | 97.2\% |
| $p^{0}$ | 2/28/2021 | 90,648 | 5,437 | 25 | 685 | 94.34\% |
|  | 3/31/2021 | 88,731 | 1,917 | 148 | 219 | 92.35\% |
|  | 4/30/2021 | 0** | 90 | 203 | 140 | 92.25\% |
|  | Subtotal: | 0** | 7,444 | 376 | 140 | 92.25\% |
|  | Cumulative: | - | 12,535 | 1,858 | NA | 95.51\% |

**88,641 fish were transferred to Omak on April 26, 2021

| pr Month | Total on hand | Mortality | Feed Fed | Fish per pound | Cumulative <br> Survival (\%) |
| :---: | :---: | :---: | :---: | :---: | :---: |
| +0 4/30/2021 | 88,583 | 58 | 66 | 140 | 99.93\% |
| 5/31/2021 | 88,474 | 109 | 726 | 47 | 99.94\% |
| Cumulative: | 88,474 | 167 | 792 | 47 | 99.94\% |

*88,474 were released from the Omak Pond on May 27, 2021

Table A 7. Chief Joseph Hatchery brood year 2020 summer/fall Chinook yearling rearing summary.

|  | Month | Total on hand* | Mortality | Feed Fed | Fish per pound | Cumulative <br> Survival(\%) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\mathrm{N}^{0}$ | 4/30/2021 | 248,285 | 595 | 6 | 792 | 99.76\% |
|  | 5/31/2021 | 480,970 | 5,594 | 262 | 462 | 98.73\% |
|  | Subtotal: | 480,970 | 6,189 | 268 | 462 | 98.73\% |
| $p^{R}$ | 3/31/2021 | 353,042 | 3,577 | - | 1,200 | 99.00\% |
|  | 4/30/2021 | 677,323 | 5,719 | 534 | 410 | 98.65\% |
|  | 5/31/2021 | 675,480 | 1,843 | 1,738 | 156 | 98.38\% |
|  | Subtotal: | 675,480 | 11,139 | 2,272 | 156 | 98.38\% |
|  | Cumulative: | 675,480 | 17,328 | 2,540 | NA | 98.52\% |

## Chief Joseph Hatchery Ladder

The CJH ladder is operated with the primary purpose of adult management (reducing pHOS) but can also be utilized to collect brood for the segregated program in years of low abundance or if the purse seine is not effective due to environmental conditions. The CJH fish ladder began operation on May 18, 2020, for spring chinook broodstock collection, with the first summer/fall chinook adult management activities occurring on August 11th. All hatchery Chinook and Sockeye were removed from the ladder and utilized for Tribal subsistence and ceremonial food purposes. All steelhead and NOR Chinook were returned to the river via a water-to-water transfer.

From August $11^{\text {th }}$ thru August 18th, 1,053 hatchery-origin summer/fall Chinook were removed at the CJH ladder and were utilized for tribal subsistence purposes (Table A 8). A total of 62 natural-origin summer/fall Chinook ( 59 adults, 3 jacks), no NOR steelhead or HOR steelhead were trapped during summer/fall Chinook ladder operations (Tables A 9 and A 10).

Table A 8. Chief Joseph Hatchery adult summer/fall Chinook ladder operations from June to August 2020.

| Month | \# of Ladder <br> Trap Checks | HOR Adults <br> Surplussed | HOR Jacks <br> Surplussed | NOR <br> Adults <br> RTS | NOR Jacks <br> RTS | HOR <br> Adults <br> RTS | HOR Jacks <br> RTS |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| June | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| July | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Aug | 2 | 1,053 | 88 | 59 | 3 | 0 | 0 |
| Total | $\mathbf{2}$ | $\mathbf{1 , 0 5 3}$ | $\mathbf{8 8}$ | $\mathbf{5 9}$ | $\mathbf{3}$ | $\mathbf{0}$ | $\mathbf{0}$ |

RTS= Return to stream

Table A 9. Chief Joseph Hatchery adult spring Chinook, Sockeye, and steelhead ladder operations from May to August 2020.

| Month | \# of Ladder <br> Trap Checks | HOR Spring <br> Chinook <br> Surplussed | HOR Spring <br> Chinook Jacks <br> Surplussed | NOR Spring <br> Chinook <br> RTS | NOR Spring <br> Chinook <br> Jacks RTS | Sockeye <br> Surplussed | AD Present <br> Steelhead <br> RTS | AD Absent <br> Steelhead <br> RTS |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| May | 4 | 0 | 0 | 5 | 0 | 0 | 2 | 32 |
| June | 7 | 0 | 0 | 52 | 4 | 0 | 3 | 3 |
| July | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Aug | 2 | 0 | 0 | 13 | 2 | 0 | 0 | 0 |
| Total | $\mathbf{1 0}$ | $\mathbf{0}$ | $\mathbf{0}$ | $\mathbf{6 5}$ | $\mathbf{6}$ | $\mathbf{0}$ | $\mathbf{3}$ | $\mathbf{3}$ |

RTS= Return to stream

Table A 10. Chief Joseph Hatchery annual summer/fall Chinook, Sockeye, and steelhead collected during summer/fall Chinook ladder operations.

| Date | HOR Chinook <br> surplussed | HOR jacks (1) <br> surplussed | NOR <br> Chinook <br> RTS | NOR jack <br> RTS | HOR <br> Chinook <br> Brood | Sockeye | AD <br> Present <br> Steelhead <br> RTS | AD Absent <br> Steelhead <br> RTS | Coho RTS |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |

[^8]The ladder was closed and dewatered mid-September for the season. The protocol was to sample $20 \%$ (one of five) of the adipose-clipped summer/fall Chinook for code-wire tags (CWT). Snouts with positive CWT detection were held frozen until December 2020 when CWT extraction and reading took place in the Chief Joseph Hatchery lab. Recovery data were expanded by the tag rate at the hatchery of origin and the sample rate at the ladder. Please refer to the Methods section for details on the expansion process for recovered tags. Beginning with jacks in 2016, snouts without a tag were assumed to be from the CJH segregated program.

Seven summer/fall Chinook hatchery programs were encountered at the CJH ladder in 2020, with the majority coming from the CJH segregated program (69\%), Wells Hatchery (16\%) and Douglas County PUD (7\%) (Table A 12). Approximately half of the recoveries were from ad-clipped, non-coded wire tagged (CWT) fish and are presumed to be from the CJH segregated program since this is the only program above Priest Rapids that releases ad-clipped, non-CWT fish.

Table A 11. Summary of summer/fall Chinook coded-wire tags encountered and expansions for the CJH ladder in 2020.

| Category | Hatchery <br> Program | \# Tags | Expanded <br> Abundance | \% of Ladder <br> Surplus |
| :---: | :---: | :---: | :---: | :---: |
|  | Omak Yearlings | 0 | 0.0 | $0.0 \%$ |
|  | Omak <br> Subyearlings | 0 | 0.0 | $0.0 \%$ |
|  | Similkameen | 3 | 11.7 | $1.1 \%$ |
| CJH Segregated | Segregated <br> yearlings | 61 | 232.1 | $22.0 \%$ |
|  | Segregated <br> subyearlings | 3 | 11.4 | $1.1 \%$ |
|  | No CWT, <br> presumed Segr | 132 | 485.7 | $46.1 \%$ |
| Other UCR summer/fall <br> Chinook hatchery | Whells | 44 | 167.5 | $15.9 \%$ |
|  | DCPUD | 15 | 57.0 | $5.4 \%$ |
|  | WDFW | 3 | 76.2 | $7.2 \%$ |
|  | Entiat NFH | 0 | 0.0 | $0.0 \%$ |
| Total |  | $\mathbf{2 8 1}$ | $\mathbf{1 0 5 3 . 0}$ | $\mathbf{1 0 0 \%}$ |

Table A 12. Percent of CJH ladder surplus adult (age 4+) summer/fall Chinook each year estimated to be from various facilities based on CWT assessment. Similkameen includes some returns from Bonaparte Pond releases (2010 and 2011). Chelan includes returns from the Turtle Rock program (2010 and 2011). 2017 was the first year of adults (4-year-olds) to CJH. 2018 was the first return year with a full complement of brood years in the return (through age 5).

|  | Surplus Fish | Facility/Program |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | CJH Seg. ${ }^{\text {a }}$ | Omak | Similk ${ }^{\text {b }}$ | Wells | Chelan ${ }^{\text {c }}$ | Carlton | Entiat | Dryden ${ }^{\text {d }}$ | Priest | Other |
| 2013 | 1,061 | 0\% | 0\% | 10\% | 22\% | 33\% | 8\% | 0\% | 26\% | 1\% | 1\% |
| 2014 | 2,008 | 0\% | 0\% | 10\% | 28\% | 26\% | 8\% | 2\% | 11\% | 0\% | 0\% |
| 2015 | 6,802 | 1\% | 0\% | 13\% | 34\% | 29\% | 6\% | 4\% | 12\% | 0\% | 0\% |
| 2016 | 5,788 | 5\% | 2\% | 3\% | 50\% | 26\% | 2\% | 2\% | 8\% | 0\% | 0\% |
| 2017 | 4,310 | 21\% | 7\% | 1\% | 35\% | 28\% | 2\% | 1\% | 5\% | 0\% | <1\% |
| 2018 | 2,249 | 58\% | 4\% | 3\% | 16\% | 13\% | 2\% | 1\% | 3\% | 0\% | <1\% |
| 2019 | 1,404 | 53\% | <1\% | 3\% | 30\% | 8\% | <1\% | 1\% | 4\% | 0\% | 0\% |
| 2020 | 1,053 | 69\% | 0\% | 1\% | 21\% | 5\% | 2\% | 0\% | 1\% | 0\% | 0\% |
| Avg. | 3,084 | 26\% | 2\% | 6\% | 30\% | 21\% | 4\% | 1\% | 9\% | <1\% | <1\% |

${ }^{\text {a }}$ Includes recoveries with 'no coded wire tags' in 2013-present: 2013 (47), 2014 (152), 2015 (71), 2016(45), 2017(76), 2018 (177), 2019 (130); starting in 2017 recoveries with 'no coded wire tags' were classified as CJH segregated fish which was the first year of adults ( $4+$ ) returned back to the CJH
${ }^{\mathrm{b}}$ Includes Bonaparte Pond releases, all years

${ }^{d}$ Includes releases by the Eastbank Hatchery into the Wenatchee R. (2013)
eIncludes releases from DCPUD(5\%) and WDFW(3\%) 2020.

## Appendix B

## 2021 Production Plan

Table B 1. Summer/Fall Chinook - Integrated Program

| Chief Joseph Hatchery Production Plan |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Brood Year: | 2021 |  |  |  |  |  | Planting Goal: | 1,100,000 |  |  |
| Species: | Summer Chinook |  |  |  |  |  | Pounds: | 86,000 |  |  |
| Stock: | Okanogan |  |  |  |  |  |  |  |  |  |
| Origin: | Wild |  |  |  |  |  |  |  |  |  |
| Program: | Integrated |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Egg Take Goal: | 1,485,000 |  |  |  |  | Adult Goal: |  | 656 |  |  |
|  |  |  |  |  |  | Assumed Fecundity |  | 5,000 |  |  |
| Estimated Release Data: |  |  |  |  |  | Average Fecundity (BY15-BY20) |  | 4,059 |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Start Date: | End Date: | Num Released | fish per lb. | Wt. grams | Total weight (lb.) | Total weight (kg) | Life Stage | Release Site | Mark Type | Tagged |
| 05/15/22 | 06/01/22 | 300,000 | 50.0 | 9.1 | 6,000 | 2,722 | Sub-Yearlings | Omak | Ad Clipped | 100\% CWT |
| 04/15/23 | 04/30/23 | 400,000 | 10.0 | 45.4 | 40,000 | 18,144 | Yearlings | Similkameen | Ad Clipped | 100\% CWT |
| 04/15/23 | 04/30/23 | 400,000 | 10.0 | 45.4 | 40,000 | 18,144 | Yearlings | Omak | Ad Clipped | 100\% CWT |
| Notes: | Egg take goal includes 3\% for culling. |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
|  | Adult Goal includes 10\% pre-spawn mortality |  |  |  |  |  |  |  |  |  |
|  | 10\% Green to Eyed egg mortality |  |  |  |  |  |  |  |  |  |
|  | Rearing mortality $10.7 \%$ for all groups |  |  |  |  |  |  |  |  |  |
| Rearing Summary: |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Species | Source | Date | Number Green Eggs | Number Eyed Eggs | Number Ponded | Fed Fry | Released | Location |  |  |
| EA SU Chinook Sub | Okanogan | June | 392,850 | 353,565 | 335,887 | 319,092 | 300,000 | Omak |  |  |
| EA SU Chinook YR | Okanogan | April | 523,800 | 471,420 | 447,849 | 425,457 | 400,000 | Similkameen |  |  |
| EA SU Chinook YR | Okanogan | April | 523,800 | 471,420 | 447,849 | 425,457 | 400,000 | Omak |  |  |

Table B 2. Summer/Fall Chinook - Segregated Program (CJH Site Release)

| Chief Joseph Hatchery Production Plan |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Brood Year: | 2021 |  |  |  |  |  | Planting Goal: | 900,000 |  |  |
| Species: | Summer Chinook |  |  |  |  |  | Pounds: | 58,000 |  |  |
| Stock: | Okanogan |  |  |  |  |  |  |  |  |  |
| Origin: | Hatchery |  |  |  |  |  |  |  |  |  |
| Program: | Segregated |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Egg Take Goal: | 1,240,000 |  |  |  |  |  | Adult Goal: | $552^{7}$ |  |  |
|  |  |  |  |  |  | Assu | Amed Fecundity | 5,000 |  |  |
| Estimated Release Data: |  |  |  |  |  | Average Fecundity (BY15-BY20) |  | 3,873 |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Start Date: | End Date: | Num Released | fish per lb. | Wt. grams | Total weight (lb.) | Total weight (kg) | Life Stage | Release Site | Mark Type | Tagged |
| 05/15/22 | 06/01/22 | 400,000 | 50.0 | 9.1 | 8,000 | 3,629 | Sub-Yearlings | CJ Hatchery | Ad Clipped | 100k CWT |
| 04/15/23 | 04/30/23 | 500,000 | 10.0 | 45.4 | 50,000 | 22,680 | Yearlings | CJ Hatchery | Ad Clipped | 100k CWT |
|  |  |  |  |  |  |  |  |  |  |  |
| Notes: | Egg take goal | includes 5\% for c | culling. |  |  |  |  |  |  |  |
|  | Adult Goal inc | ludes 10\% pre-sp | pawn mortality |  |  |  |  |  |  |  |
|  | 10\% Green to | Eyed egg mortal |  |  |  |  |  |  |  |  |
|  | Rearing mort | ality is 9.7\% for $y$ | yearlings, 11.7 | \% for sub-ye | earlings. |  |  |  |  |  |
| Rearing Summary: |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Species | Source | Date | Number Green Eggs | Number Eyed Eggs | Number Ponded | Fed Fry | Released | Location |  |  |
| EA SU Chinook Sub | Okanogan | June | 530,100 | 477,090 | 453,236 | 430,574 | 400,000 | CJ Hatchery |  |  |
| EA SU Chinook YR | Okanogan | April | 647,900 | 583,110 | 553,955 | 526,257 | 500,000 | CJ Hatchery |  |  |

## Appendix C

pHOS and Effective pHOS

Table C 1. Annual Chinook spawning grounds data for the Okanogan Basin from 2006 to 2020, including pHOS and effective pHOS values per reach.

| 2020 |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| reach | redds | fish per redd | spawners per reach | \% sampled | total carcasses | hatchery carcasses | wild carcasses | \% hatchery | \% wild | HOS | NOS | pHOS |
| 01 | 25 | 2.67 | 67 | 10.5\% | 7 | 5 | 2 | 71.4\% | 28.6\% | 48 | 19 | 0.71 |
| 02 | 51 | 2.67 | 136 | 18.4\% | 25 | 17 | 8 | 68.0\% | 32.0\% | 93 | 44 | 0.68 |
| 03 | 270 | 2.67 | 721 | 19.7\% | 142 | 84 | 58 | 59.2\% | 40.8\% | 426 | 294 | 0.59 |
| 04 | 103 | 2.67 | 275 | 15.6\% | 43 | 24 | 19 | 55.8\% | 44.2\% | 153 | 122 | 0.56 |
| 05 | 683 | 2.67 | 1824 | 21.6\% | 393 | 110 | 283 | 28.0\% | 72.0\% | 510 | 1313 | 0.28 |
| 06 | 1254 | 2.67 | 3348 | 25.2\% | 843 | 117 | 726 | 13.9\% | 86.1\% | 465 | 2883 | 0.14 |
| S1 | 1445 | 2.67 | 3858 | 25.7\% | 993 | 295 | 698 | 29.7\% | 70.3\% | 1146 | 2712 | 0.30 |
| S2 | 296 | 2.67 | 790 | 20.0\% | 158 | 44 | 114 | 27.8\% | 72.2\% | 220 | 570 | 0.28 |
| Totals | 4127 |  | 11019 | 23.6\% | 2604 | 696 | 1908 |  |  | 3062 | 7957 | 0.28 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  | pHOS | 0.28 |
|  |  |  |  |  |  |  |  |  |  |  | effective pHOS | 0.24 |


| 2019 |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| reach | redds | fish per redd | spawners per reach | \% sampled | total carcasses | hatchery carcasses | wild carcasses | \%hatchery | \%wild | HOS | NOS | pHOS |
| O1* | 12 | 2.3 | 28 | 0.0\% | 0 | 0 | 0 | 51.6\% | 48.4\% | 14 | 13 | 0.52 |
| O2* | 154 | 2.3 | 354 | 0.3\% | 1 | 1 | 0 | 51.6\% | 48.4\% | 183 | 171 | 0.52 |
| 03 | 275 | 2.3 | 633 | 11.4\% | 72 | 52 | 20 | 72.2\% | 27.8\% | 457 | 176 | 0.72 |
| 04* | 92 | 2.3 | 212 | 4.3\% | 9 | 4 | 5 | 51.6\% | 48.4\% | 109 | 102 | 0.52 |
| 05 | 600 | 2.3 | 1380 | 5.2\% | 72 | 34 | 38 | 47.2\% | 52.8\% | 652 | 728 | 0.47 |
| 06 | 505 | 2.3 | 1162 | 15.6\% | 181 | 76 | 105 | 42.0\% | 58.0\% | 488 | 674 | 0.42 |
| S1 | 694 | 2.3 | 1596 | 7.9\% | 126 | 70 | 56 | 55.6\% | 44.4\% | 887 | 709 | 0.56 |
| S2 | 39 | 2.3 | 90 | 6.7\% | 6 | 4 | 2 | 66.7\% | 33.3\% | 60 | 30 | 0.67 |
| Totals | 2371 |  | 5453 | 8.6\% | 467 | 241 | 226 |  |  | 2849 | 2604 | 0.52 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
| *Indicates '\%hatchery' and '\%wild' values were estimated from basin carcass totals (only performed when carcasses recovered represented <5\% of spawners for that reach) |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  | pHOS | 0.52 |
|  |  |  |  |  |  |  |  |  |  |  | effective pHOS | 0.47 |


| 2018 |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| reach | redds | fish per redd | spawners per reach | \% sampled | total carcasses | hatchery carcasses | wild carcasses | \%hatchery | \%wild | HOS | NOS | pHOS |
| O1* | 11 | 2.301 | 25 | 0.0\% | 0 | 0 | 0 | 31.6\% | 68.4\% | 8 | 17 | 0.32 |
| O2* | 74 | 2.301 | 170 | 0.0\% | 0 | 0 | 0 | 31.6\% | 68.4\% | 54 | 116 | 0.32 |
| 03 | 211 | 2.301 | 486 | 16.1\% | 78 | 40 | 38 | 51.3\% | 48.7\% | 249 | 237 | 0.51 |
| 04* | 133 | 2.301 | 306 | 2.6\% | 8 | 1 | 7 | 31.6\% | 68.4\% | 97 | 209 | 0.32 |
| 05 | 618 | 2.301 | 1422 | 9.4\% | 134 | 49 | 85 | 36.6\% | 63.4\% | 520 | 902 | 0.37 |
| 06 | 507 | 2.301 | 1167 | 16.3\% | 190 | 33 | 157 | 17.4\% | 82.6\% | 203 | 964 | 0.17 |
| S1 | 501 | 2.301 | 1153 | 11.4\% | 131 | 48 | 83 | 36.6\% | 63.4\% | 422 | 730 | 0.37 |
| S2* | 57 | 2.301 | 131 | 4.6\% | 6 | 2 | 4 | 31.6\% | 68.4\% | 41 | 90 | 0.32 |
| Totals | 2112 |  | 4860 | 11.3\% | 547 | 173 | 374 |  |  | 1594 | 3266 | 0.33 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
| *Indicates '\%hatchery' and '\%wild' values were estimated from basin carcass totals (only performed when carcasses recovered represented <5\% of spawners for that reach) |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  | pHOS | 0.33 |
|  |  |  |  |  |  |  |  |  |  |  | effective pHOS | 0.28 |


| 2017 |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| reach | redds | fish per redd | spawners per reach | \% sampled | total carcasses | hatchery carcasses | wild carcasses | \%hatchery | \%wild | HOS | NOS | pHOS |
| O1* | 2 | 2.039 | 4 | 0.0\% | 0 | 0 | 0 | 17.0\% | 83.0\% | 1 | 3 | 0.17 |
| 02 | 62 | 2.039 | 126 | 6.3\% | 8 | 4 | 4 | 50.0\% | 50.0\% | 63 | 63 | 0.50 |
| O3* | 192 | 2.039 | 391 | 2.3\% | 9 | 5 | 4 | 17.0\% | 83.0\% | 66 | 325 | 0.17 |
| 04 | 111 | 2.039 | 226 | 7.1\% | 16 | 5 | 11 | 31.3\% | 68.8\% | 71 | 156 | 0.31 |
| 05* | 830 | 2.039 | 1692 | 3.5\% | 60 | 10 | 50 | 17.0\% | 83.0\% | 287 | 1405 | 0.17 |
| 06 | 1237 | 2.039 | 2522 | 24.9\% | 628 | 66 | 562 | 10.5\% | 89.5\% | 265 | 2257 | 0.11 |
| S1 | 710 | 2.039 | 1448 | 31.3\% | 453 | 106 | 347 | 23.4\% | 76.6\% | 339 | 1109 | 0.23 |
| S2 | 77 | 2.039 | 157 | 17.2\% | 27 | 8 | 19 | 29.6\% | 70.4\% | 47 | 110 | 0.30 |
| Totals | 3221 |  | 6568 | 18.3\% | 1201 | 204 | 997 |  |  | 1139 | 5429 | 0.17 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
| *Indicates '\%hatchery' and '\%wild' values were estimated from basin carcass totals (only performed when carcasses recovered represented <5\% of spawners for that reach) |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  | pHOS | 0.17 |
|  |  |  |  |  |  |  |  |  |  |  | effective pHOS | 0.14 |


| 2016 |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| reach | redds | fish per redd | spawners per reach | \% sampled | total carcasses | hatchery carcasses | wild carcasses | \%hatchery | \%wild | HOS | NOS | pHOS |
| O1* | 2 | 2.01 | 4 | 0.0\% | 0 | 0 | 0 | 21.2\% | 78.8\% | 1 | 3 | 0.21 |
| 02 | 57 | 2.01 | 115 | 10.5\% | 12 | 6 | 6 | 50.0\% | 50.0\% | 57 | 57 | 0.50 |
| 03 | 52 | 2.01 | 105 | 13.4\% | 14 | 1 | 13 | 7.1\% | 92.9\% | 7 | 97 | 0.07 |
| 04* | 130 | 2.01 | 261 | 4.2\% | 11 | 4 | 7 | 21.2\% | 78.8\% | 55 | 206 | 0.21 |
| 05 | 907 | 2.01 | 1823 | 12.6\% | 230 | 44 | 186 | 19.1\% | 80.9\% | 349 | 1474 | 0.19 |
| 06 | 2338 | 2.01 | 4699 | 22.9\% | 1075 | 56 | 1019 | 5.2\% | 94.8\% | 245 | 4455 | 0.05 |
| S1 | 1645 | 2.01 | 3306 | 36.7\% | 1214 | 395 | 819 | 32.5\% | 67.5\% | 1076 | 2231 | 0.33 |
| S2 | 145 | 2.01 | 291 | 68.3\% | 199 | 78 | 121 | 39.2\% | 60.8\% | 114 | 177 | 0.39 |
| Totals | 5276 |  | 10605 | 26.0\% | 2755 | 584 | 2171 |  |  | 1905 | 8700 | 0.18 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
| *Indicates '\%hatchery' and '\%wild' values were estimated from basin carcass totals (only performed when carcasses recovered represented <5\% of spawners for that reach) |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  | pHOS | 0.18 |
|  |  |  |  |  |  |  |  |  |  |  | effective pHOS | 0.15 |


| 2015 |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| reach | redds | fish per redd | spawners per reach | \% sampled | total carcasses | hatchery carcasses | wild carcasses | \%hatchery | \%wild | HOS | NOS | pHOS |
| O1* | 36 | 3.215 | 116 | 0.0\% | 0 | 0 | 0 | 22.4\% | 77.6\% | 26 | 90 | 0.22 |
| O2* | 113 | 3.215 | 363 | 2.8\% | 10 | 5 | 5 | 22.4\% | 77.6\% | 81 | 282 | 0.22 |
| 03 | 284 | 3.215 | 913 | 6.7\% | 61 | 22 | 39 | 36.1\% | 63.9\% | 329 | 584 | 0.36 |
| 04* | 79 | 3.215 | 254 | 4.3\% | 11 | 2 | 9 | 18.2\% | 77.6\% | 46 | 197 | 0.19 |
| 05 | 1008 | 3.215 | 3241 | 8.7\% | 283 | 74 | 209 | 26.1\% | 73.9\% | 847 | 2393 | 0.26 |
| 06 | 859 | 3.215 | 2762 | 36.0\% | 994 | 63 | 931 | 6.3\% | 93.7\% | 175 | 2587 | 0.06 |
| S1 | 1611 | 3.215 | 5179 | 32.9\% | 1702 | 516 | 1186 | 30.3\% | 69.7\% | 1570 | 3609 | 0.30 |
| S2 | 286 | 3.215 | 919 | 25.2\% | 232 | 56 | 176 | 24.1\% | 75.9\% | 222 | 698 | 0.24 |
| Totals | 4276 |  | 13747 | 24.0\% | 3293 | 738 | 2555 |  |  | 3297 | 10439 | 0.24 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
| *Indicates '\%hatchery' and '\%wild' values were estimated from basin carcass totals (only performed when carcasses recovered represented <5\% of spawners for that reach) |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  | pHOS | 0.24 |
|  |  |  |  |  |  |  |  |  |  |  | effective pHOS | 0.20 |


| 2014 |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| reach | redds | fish per redd | spawners per reach | \% sampled | total carcasses | hatchery carcasses | wild carcasses | \%hatchery | \%wild | HOS | NOS | pHOS |
| O1* | 11 | 2.86 | 31 | 3.2\% | 1 | 1 | 0 | 13.4\% | 86.6\% | 4 | 27 | 0.13 |
| O2* | 57 | 2.86 | 163 | 0.6\% | 1 | 0 | 1 | 13.4\% | 86.6\% | 22 | 141 | 0.13 |
| 03 | 191 | 2.86 | 546 | 14.5\% | 79 | 19 | 60 | 24.1\% | 75.9\% | 131 | 415 | 0.24 |
| 04 | 111 | 2.86 | 317 | 17.0\% | 54 | 7 | 47 | 13.0\% | 87.0\% | 41 | 276 | 0.13 |
| 05 | 851 | 2.86 | 2434 | 11.3\% | 275 | 42 | 233 | 15.3\% | 84.7\% | 372 | 2062 | 0.15 |
| 06 | 1010 | 2.86 | 2889 | 27.1\% | 783 | 67 | 716 | 8.6\% | 91.4\% | 247 | 2641 | 0.09 |
| S1 | 1737 | 2.86 | 4968 | 15.5\% | 770 | 129 | 641 | 16.8\% | 83.2\% | 832 | 4136 | 0.17 |
| S2 | 285 | 2.86 | 815 | 60.0\% | 489 | 64 | 425 | 13.1\% | 86.9\% | 107 | 708 | 0.13 |
| Totals | 4253 |  | 12164 | 20.2\% | 2452 | 329 | 2123 |  |  | 1756 | 10407 | 0.14 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
| *Indicates '\%hatchery' and '\%wild' values were estimated from basin carcass totals (only performed when carcasses recovered represented <5\% of spawners for that reach) |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  | pHOS | 0.14 |
|  |  |  |  |  |  |  |  |  |  |  | effective pHOS | 0.12 |


| 2013 |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| reach | redds | fish per redd | spawners per reach | \% sampled | total carcasses | hatchery <br> carcasses | wild carcasses | \%hatchery | \%wild | HOS | NOS | pHOS |
| 01 | 3 | 2.31 | 7 | 0.0\% | 0 | 0 | 0 | 32.6\% | 67.4\% | 2 | 5 | 0.33 |
| O2* | 2 | 2.31 | 5 | 0.0\% | 0 | 0 | 0 | 32.6\% | 67.4\% | 2 | 3 | 0.33 |
| 03 | 158 | 2.31 | 365 | 8.2\% | 30 | 8 | 22 | 26.7\% | 73.3\% | 97 | 268 | 0.27 |
| 04 | 46 | 2.31 | 106 | 8.5\% | 9 | 2 | 7 | 22.2\% | 77.8\% | 24 | 83 | 0.22 |
| 05 | 397 | 2.31 | 917 | 5.7\% | 52 | 15 | 37 | 28.8\% | 71.2\% | 265 | 653 | 0.29 |
| 06 | 1661 | 2.31 | 3837 | 11.3\% | 432 | 80 | 352 | 18.5\% | 81.5\% | 711 | 3126 | 0.19 |
| S1 | 1254 | 2.31 | 2897 | 13.1\% | 379 | 188 | 191 | 49.6\% | 50.4\% | 1437 | 1460 | 0.50 |
| S2 | 26 | 2.31 | 60 | 13.3\% | 8 | 4 | 4 | 50.0\% | 50.0\% | 30 | 30 | 0.50 |
| Totals | 3547 |  | 8194 | 11.1\% | 910 | 297 | 613 |  |  | 2567 | 5627 | 0.31 |
| *Indicates '\%hatchery' and '\%wild' values were estimated from basin carcass totals (only performed when carcasses recovered represented <5\% of spawners for that reach) |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  | pHOS | 0.31 |
|  |  |  |  |  |  |  |  |  |  |  | effective pHOS | 0.27 |


| reach | redds | fish per redd | spawners per reach | \% sampled | total carcasses | hatchery carcasses | wild carcasses | \%hatchery | \%wild | HOS | NOS | pHOS |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| O1* | 12 | 3.07 | 37 | 2.7\% | 1 | 1 | 0 | 42.3\% | 57.7\% | 16 | 21 | 0.42 |
| O2* | 54 | 3.07 | 166 | 0.0\% | 0 | 0 | 0 | 42.3\% | 57.7\% | 70 | 96 | 0.42 |
| 03 | 159 | 3.07 | 488 | 11.5\% | 56 | 38 | 18 | 67.9\% | 32.1\% | 331 | 157 | 0.68 |
| 04 | 68 | 3.07 | 209 | 7.2\% | 15 | 6 | 9 | 40.0\% | 60.0\% | 84 | 125 | 0.40 |
| 05 | 555 | 3.07 | 1704 | 15.0\% | 256 | 123 | 133 | 48.0\% | 52.0\% | 819 | 885 | 0.48 |
| 06 | 765 | 3.07 | 2349 | 22.9\% | 537 | 110 | 427 | 20.5\% | 79.5\% | 481 | 1867 | 0.20 |
| S1 | 914 | 3.07 | 2806 | 17.6\% | 494 | 288 | 206 | 58.3\% | 41.7\% | 1636 | 1170 | 0.58 |
| S2 | 152 | 3.07 | 467 | 11.6\% | 54 | 31 | 23 | 57.4\% | 42.6\% | 268 | 199 | 0.57 |
| Totals | 2679 |  | 8225 | 17.2\% | 1413 | 597 | 816 |  |  | 3704 | 4521 | 0.45 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
| *Indicates '\%hatchery' and '\%wild' values were estimated from basin carcass totals (only performed when carcasses recovered represented <5\% of spawners for that reach) |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  | pHOS | 0.45 |
|  |  |  |  |  |  |  |  |  |  |  | effective <br> pHOS | 0.40 |


| 2011 |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| reach | redds | fish per redd | spawners per reach | \% sampled | total carcasses | hatchery carcasses | wild carcasses | \%hatchery | \%wild | HOS | NOS | pHOS |
| O1* | 3 | 3.1 | 9 | 0.0\% | 0 | 0 | 0 | 53.6\% | 46.4\% | 5 | 4 | 0.54 |
| 02* | 20 | 3.1 | 62 | 0.0\% | 0 | 0 | 0 | 53.6\% | 46.4\% | 33 | 29 | 0.54 |
| 03 | 101 | 3.1 | 313 | 17.6\% | 55 | 34 | 21 | 61.8\% | 38.2\% | 194 | 120 | 0.62 |
| 04 | 55 | 3.1 | 171 | 8.2\% | 14 | 10 | 4 | 71.4\% | 28.6\% | 122 | 49 | 0.71 |
| 05 | 593 | 3.1 | 1838 | 19.6\% | 361 | 160 | 201 | 44.3\% | 55.7\% | 815 | 1024 | 0.44 |
| 06 | 942 | 3.1 | 2920 | 16.4\% | 478 | 116 | 362 | 24.3\% | 75.7\% | 709 | 2212 | 0.24 |
| S1 | 1217 | 3.1 | 3773 | 20.0\% | 753 | 537 | 216 | 71.3\% | 28.7\% | 2690 | 1082 | 0.71 |
| S2 | 192 | 3.1 | 595 | 19.2\% | 114 | 95 | 19 | 83.3\% | 16.7\% | 496 | 99 | 0.83 |
| Totals | 3123 |  | 9681 | 18.3\% | 1775 | 952 | 823 |  |  | 5063 | 4618 | 0.52 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
| *Indicates '\%hatchery' and '\%wild' values were estimated from basin carcass totals (only performed when carcasses recovered represented <5\% of spawners for that reach) |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  | pHOS | 0.52 |
|  |  |  |  |  |  |  |  |  |  |  | effective pHOS | 0.47 |


| reach | redds | fish per redd | spawners per reach | \% sampled | total carcasses | hatchery carcasses | wild carcasses | \%hatchery | \%wild | HOS | NOS | pHOS |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 01 | 9 | 2.81 | 25 | 11.9\% | 3 | 2 | 1 | 66.7\% | 33.3\% | 17 | 8 | 0.67 |
| 02 | 58 | 2.81 | 163 | 6.1\% | 10 | 5 | 5 | 50.0\% | 50.0\% | 81 | 81 | 0.50 |
| 03 | 67 | 2.81 | 188 | 15.9\% | 30 | 11 | 19 | 36.7\% | 63.3\% | 69 | 119 | 0.37 |
| 04 | 89 | 2.81 | 250 | 16.8\% | 42 | 24 | 18 | 57.1\% | 42.9\% | 143 | 107 | 0.57 |
| O5 | 357 | 2.81 | 1003 | 24.0\% | 241 | 87 | 154 | 36.1\% | 63.9\% | 362 | 641 | 0.36 |
| 06 | 431 | 2.81 | 1211 | 29.1\% | 352 | 172 | 180 | 48.9\% | 51.1\% | 592 | 619 | 0.49 |
| S1 | 895 | 2.81 | 2515 | 24.9\% | 625 | 296 | 329 | 47.4\% | 52.6\% | 1191 | 1324 | 0.47 |
| S2 | 212 | 2.81 | 596 | 24.8\% | 148 | 79 | 69 | 53.4\% | 46.6\% | 318 | 278 | 0.53 |
| Totals | 2118 |  | 5952 | 24.4\% | 1451 | 676 | 775 |  |  | 2773 | 3178 | 0.47 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  | pHOS | 0.47 |
|  |  |  |  |  |  |  |  |  |  |  | effective pHOS | 0.41 |


| 2009 |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| reach | redds | fish per redd | spawners per reach | \% sampled | total carcasses | hatchery carcasses | wild carcasses | \%hatchery | \%wild | HOS | NOS | pHOS |
| 01 | 3 | 2.54 | 8 | 26.2\% | 2 | 0 | 2 | 0.0\% | 100.0\% | 0 | 8 | 0.00 |
| 02 | 32 | 2.54 | 81 | 8.6\% | 7 | 4 | 3 | 57.1\% | 42.9\% | 46 | 35 | 0.57 |
| 03 | 91 | 2.54 | 231 | 13.4\% | 31 | 18 | 13 | 58.1\% | 41.9\% | 134 | 97 | 0.58 |
| 04 | 138 | 2.54 | 351 | 9.1\% | 32 | 18 | 14 | 56.3\% | 43.8\% | 197 | 153 | 0.56 |
| 05 | 621 | 2.54 | 1577 | 22.1\% | 348 | 159 | 189 | 45.7\% | 54.3\% | 721 | 857 | 0.46 |
| 06 | 787 | 2.54 | 1999 | 25.0\% | 500 | 153 | 347 | 30.6\% | 69.4\% | 612 | 1387 | 0.31 |
| S1 | 1091 | 2.54 | 2771 | 25.4\% | 703 | 373 | 330 | 53.1\% | 46.9\% | 1470 | 1301 | 0.53 |
| S2 | 207 | 2.54 | 526 | 28.5\% | 150 | 75 | 75 | 50.0\% | 50.0\% | 263 | 263 | 0.50 |
| Totals | 2970 |  | 7544 | 23.5\% | 1773 | 800 | 973 |  |  | 3443 | 4100 | 0.46 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  | pHOS | 0.46 |
|  |  |  |  |  |  |  |  |  |  |  | effective pHOS | 0.40 |


| 2008 |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| reach | redds | fish per redd | spawners per reach | \% sampled | total carcasses | hatchery carcasses | wild carcasses | \%hatchery | \%wild | HOS | NOS | pHOS |
| 01 | 4 | 3.25 | 13 | 30.8\% | 4 | 2 | 2 | 50.0\% | 50.0\% | 7 | 7 | 0.50 |
| 02 | 51 | 3.25 | 166 | 6.0\% | 10 | 9 | 1 | 90.0\% | 10.0\% | 149 | 17 | 0.90 |
| 03 | 60 | 3.25 | 195 | 20.5\% | 40 | 26 | 14 | 65.0\% | 35.0\% | 127 | 68 | 0.65 |
| 04 | 96 | 3.25 | 312 | 11.5\% | 36 | 25 | 11 | 69.4\% | 30.6\% | 217 | 95 | 0.69 |
| 05 | 374 | 3.25 | 1216 | 20.4\% | 248 | 141 | 107 | 56.9\% | 43.1\% | 691 | 524 | 0.57 |
| 06 | 561 | 3.25 | 1823 | 36.5\% | 665 | 341 | 324 | 51.3\% | 48.7\% | 935 | 888 | 0.51 |
| S1 | 801 | 3.25 | 2603 | 33.0\% | 859 | 512 | 347 | 59.6\% | 40.4\% | 1552 | 1052 | 0.60 |
| S2 | 199 | 3.25 | 647 | 24.3\% | 157 | 116 | 41 | 73.9\% | 26.1\% | 478 | 169 | 0.74 |
| Totals | 2146 |  | 6975 | 28.9\% | 2019 | 1172 | 847 |  |  | 4155 | 2820 | 0.60 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  | pHOS | 0.60 |
|  |  |  |  |  |  |  |  |  |  |  | effective pHOS | 0.54 |


| 2007 |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| reach | redds | fish per redd | spawners per reach | \% sampled | total carcasses | hatchery carcasses | wild carcasses | \%hatchery | \%wild | HOS | NOS | pHOS |
| 01 | 3 | 2.2 | 7 | 30.3\% | 2 | 1 | 1 | 50.0\% | 50.0\% | 3 | 3 | 0.50 |
| O2* | 16 | 2.2 | 35 | 0.0\% | 0 | 0 | 0 | 38.1\% | 61.9\% | 13 | 22 | 0.38 |
| 03 | 116 | 2.2 | 255 | 21.6\% | 55 | 25 | 30 | 45.5\% | 54.5\% | 116 | 139 | 0.45 |
| 04* | 63 | 2.2 | 139 | 0.7\% | 1 | 0 | 1 | 38.1\% | 61.9\% | 53 | 86 | 0.38 |
| 05 | 549 | 2.2 | 1208 | 37.5\% | 453 | 169 | 284 | 37.3\% | 62.7\% | 451 | 757 | 0.37 |
| 06 | 554 | 2.2 | 1219 | 42.6\% | 519 | 197 | 322 | 38.0\% | 62.0\% | 463 | 756 | 0.38 |
| S1 | 652 | 2.2 | 1434 | 45.9\% | 658 | 253 | 405 | 38.4\% | 61.6\% | 552 | 883 | 0.38 |
| S2 | 55 | 2.2 | 121 | 24.0\% | 29 | 9 | 20 | 31.0\% | 69.0\% | 38 | 83 | 0.31 |
| Totals | 2008 |  | 4418 | 38.9\% | 1717 | 654 | 1063 |  |  | 1688 | 2730 | 0.38 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
| *Indicates '\%hatchery' and '\%wild' values were estimated from basin carcass totals (only performed when carcasses recovered represented <5\% of spawners for that reach) |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  | pHOS | 0.38 |
|  |  |  |  |  |  |  |  |  |  |  | effective pHOS | 0.33 |


| 2006 |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| reach | redds | fish per redd | spawners per reach | \% sampled | total carcasses | hatchery carcasses | wild carcasses | \%hatchery | \%wild | HOS | NOS | pHOS |
| 01 | 10 | 2.02 | 20 | 19.8\% | 4 | 2 | 2 | 50.0\% | 50.0\% | 10 | 10 | 0.50 |
| 02* | 56 | 2.02 | 113 | 2.7\% | 3 | 1 | 2 | 23.0\% | 77.0\% | 26 | 87 | 0.23 |
| 03 | 175 | 2.02 | 354 | 8.8\% | 31 | 9 | 22 | 29.0\% | 71.0\% | 103 | 251 | 0.29 |
| 04 | 145 | 2.02 | 293 | 5.5\% | 16 | 6 | 10 | 37.5\% | 62.5\% | 110 | 183 | 0.38 |
| 05 | 840 | 2.02 | 1697 | 7.1\% | 120 | 15 | 105 | 12.5\% | 87.5\% | 212 | 1485 | 0.13 |
| 06 | 1366 | 2.02 | 2759 | 10.5\% | 291 | 44 | 247 | 15.1\% | 84.9\% | 417 | 2342 | 0.15 |
| S1 | 1388 | 2.02 | 2804 | 18.1\% | 508 | 138 | 370 | 27.2\% | 72.8\% | 762 | 2042 | 0.27 |
| S2 | 278 | 2.02 | 562 | 18.9\% | 106 | 33 | 73 | 31.1\% | 68.9\% | 175 | 387 | 0.31 |
| Totals | 4258 |  | 8601 | 12.5\% | 1079 | 248 | 831 |  |  | 1814 | 6787 | 0.21 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
| *Indicates '\%hatchery' and '\%wild' values were estimated from basin carcass totals (only performed when carcasses recovered represented <5\% of spawners for that reach) |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  | pHOS | 0.21 |
|  |  |  |  |  |  |  |  |  |  |  | effective pHOS | 0.18 |

Table C 2. Number of hatchery- and natural-origin (wild) summer/fall Chinook carcasses collected in each reach of the Okanogan (01-06) and Similkameen rivers from 1993 to 2020.

| Survey year | Origin | Survey reach |  |  |  |  |  |  |  | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | O-1 | O-2 | O-3 | O-4 | O-5 | O-6 | S-1 | S-2 |  |
| $1993{ }^{\text {a }}$ | Wild | 0 | 0 | 3 | 0 | 13 | 4 | 48 | 1 | 69 |
|  | Hatchery | 0 | 2 | 0 | 0 | 10 | 9 | 25 | 0 | 46 |
| $1994{ }^{\text {b }}$ | Wild | 0 | 0 | 1 | 0 | 7 | 1 | 113 | 22 | 144 |
|  | Hatchery | 0 | 4 | 3 | 0 | 20 | 4 | 205 | 38 | 274 |
| 1995 | Wild | 0 | 0 | 1 | 0 | 10 | 0 | 66 | 4 | 81 |
|  | Hatchery | 0 | 0 | 1 | 0 | 20 | 0 | 173 | 11 | 205 |
| 1996 | Wild | 0 | 0 | 0 | 1 | 3 | 1 | 53 | 0 | 58 |
|  | Hatchery | 0 | 0 | 0 | 1 | 2 | 1 | 173 | 0 | 177 |
| 1997 | Wild | 0 | 0 | 1 | 0 | 0 | 3 | 83 | 0 | 87 |
|  | Hatchery | 0 | 0 | 1 | 0 | 9 | 0 | 142 | 1 | 153 |
| 1998 | Wild | 0 | 1 | 3 | 1 | 6 | 5 | 162 | 4 | 182 |
|  | Hatchery | 0 | 0 | 5 | 0 | 1 | 2 | 178 | 0 | 186 |
| 1999 | Wild | 0 | 0 | 0 | 0 | 9 | 23 | 293 | 9 | 334 |
|  | Hatchery | 0 | 0 | 3 | 2 | 14 | 30 | 473 | 39 | 561 |
| 2000 | Wild | 0 | 0 | 8 | 8 | 24 | 11 | 189 | 4 | 244 |
|  | Hatchery | 0 | 2 | 12 | 7 | 23 | 5 | 538 | 37 | 624 |
| 2001 | Wild | 0 | 10 | 23 | 5 | 67 | 42 | 390 | 54 | 591 |
|  | Hatchery | 0 | 16 | 52 | 5 | 60 | 70 | 751 | 51 | 1,005 |
| 2002 | Wild | 6 | 14 | 20 | 10 | 81 | 212 | 340 | 72 | 755 |
|  | Hatchery | 4 | 18 | 63 | 25 | 123 | 360 | 925 | 187 | 1,705 |
| $2003{ }^{\text {c }}$ | Wild | 0 | 0 | 13 | 0 | 12 | 152 | 231 | 124 | 532 |
|  | Hatchery | 0 | 0 | 15 | 0 | 5 | 91 | 365 | 257 | 733 |
| 2004 | Wild | 0 | 2 | 19 | 19 | 108 | 225 | 1,125 | 260 | 1,758 |
|  | Hatchery | 0 | 2 | 12 | 5 | 38 | 58 | 267 | 38 | 420 |
| 2005 | Wild | 0 | 5 | 51 | 21 | 256 | 364 | 531 | 176 | 1,404 |
|  | Hatchery | 0 | 3 | 42 | 16 | 115 | 70 | 200 | 100 | 546 |
| 2006 | Wild | 2 | 2 | 22 | 10 | 105 | 247 | 370 | 73 | 831 |
|  | Hatchery | 2 | 1 | 9 | 6 | 15 | 44 | 138 | 33 | 248 |
| 2007 | Wild | 1 | 0 | 30 | 1 | 284 | 322 | 405 | 20 | 1,063 |
|  | Hatchery | 1 | 0 | 25 | 0 | 169 | 197 | 253 | 9 | 654 |


| 2008 | Wild | 2 | 1 | 14 | 11 | 107 | 324 | 347 | 41 | 847 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Hatchery | 2 | 9 | 26 | 25 | 141 | 341 | 512 | 116 | 1,172 |
| 2009 | Wild | 2 | 3 | 13 | 14 | 189 | 347 | 330 | 75 | 973 |
|  | Hatchery | 0 | 4 | 18 | 18 | 159 | 153 | 373 | 75 | 800 |
| 2010 | Wild | 1 | 5 | 19 | 18 | 154 | 180 | 329 | 69 | 775 |
|  | Hatchery | 2 | 5 | 11 | 24 | 87 | 172 | 296 | 79 | 676 |
| 2011 | Wild | 0 | 0 | 21 | 4 | 201 | 362 | 216 | 19 | 823 |
|  | Hatchery | 0 | 0 | 34 | 10 | 160 | 116 | 537 | 95 | 952 |
| 2012 | Wild | 0 | 0 | 18 | 9 | 133 | 427 | 206 | 23 | 816 |
|  | Hatchery | 1 | 0 | 38 | 6 | 123 | 110 | 288 | 31 | 597 |
| $2013{ }^{\text {dee }}$ | Wild | 0 | 0 | 22 | 7 | 37 | 352 | 191 | 4 | 613 |
|  | Hatchery | 0 | 0 | 8 | 2 | 15 | 80 | 188 | 4 | 297 |
| 2014 | Wild | 0 | 1 | 60 | 47 | 233 | 716 | 641 | 425 | 2,123 |
|  | Hatchery | 1 | 0 | 19 | 7 | 42 | 67 | 129 | 64 | 329 |
| 2015 | Wild | 0 | 5 | 39 | 9 | 209 | 931 | 1186 | 176 | 2,555 |
|  | Hatchery | 0 | 5 | 22 | 2 | 74 | 63 | 516 | 56 | 738 |
| 2016 | Wild | 0 | 6 | 13 | 7 | 186 | 1019 | 819 | 121 | 2,171 |
|  | Hatchery | 0 | 6 | 1 | 4 | 44 | 56 | 395 | 78 | 584 |
| 2017 | Wild | 0 | 4 | 4 | 11 | 50 | 562 | 347 | 19 | 997 |
|  | Hatchery | 0 | 4 | 5 | 5 | 10 | 66 | 106 | 8 | 204 |
| 2018 | Wild | 0 | 0 | 38 | 7 | 85 | 157 | 83 | 4 | 374 |
|  | Hatchery | 0 | 0 | 40 | 1 | 49 | 33 | 48 | 2 | 173 |
| 2019 | Wild | 0 | 0 | 20 | 5 | 38 | 105 | 56 | 2 | 226 |
|  | Hatchery | 0 | 1 | 52 | 4 | 34 | 76 | 70 | 4 | 241 |
| 2020 | Wild | 2 | 8 | 58 | 19 | 283 | 726 | 698 | 114 | 1,908 |
|  | Hatchery | 5 | 17 | 84 | 24 | 110 | 117 | 295 | 44 | 696 |
| Avg. | Wild | 1 | 2 | 19 | 9 | 103 | 279 | 352 | 68 | 833 |
|  | Hatchery | 1 | 4 | 21 | 7 | 60 | 85 | 306 | 52 | 536 |

${ }^{\text {a }} 25$ additional carcasses were sampled on the Similkameen and 46 on the Okanogan without any reach designation.
${ }^{\mathrm{b}}$ One additional carcass was sampled on the Similkameen without any reach designation.
c 793 carcasses were sampled on the Similkameen before initiation of spawning (pre-spawn mortality) and an additional 40 carcasses were sampled on the Okanogan. The cause of the high mortality (Ichthyophthirius multifilis and Flavobacterium columnarae) was exacerbated by high river temperatures.
${ }^{\text {d }}$ In 2013, carcass recoveries were combined in reaches O-3 and O-4, and S-1 and S-2. Then re-apportioned based on redd counts within each reach.
e 2013 data have been updated to reflect age and origin data acquired from scale reading since the publication of the 2013 annual report

## Age at Maturity

Table C 2. Salt age of recovered carcasses in the Okanogan and Similkameen Rivers.

| Hatchery-Origin Male <br> Salt Age Carcasses Recovered |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Survey Year | 0 | 1 | 2 | 3 | 4 | 5 | Total |
| 1993 | 0 | 0 | 33 | 0 | 0 | 0 | 33 |
| 1994 | 0 | 5 | 23 | 92 | 0 | 0 | 120 |
| 1995 | 0 | 2 | 23 | 27 | 17 | 0 | 69 |
| 1996 | 0 | 3 | 17 | 24 | 5 | 0 | 49 |
| 1997 | 0 | 0 | 1 | 25 | 2 | 0 | 28 |
| 1998 | 0 | 9 | 64 | 12 | 9 | 0 | 94 |
| 1999 | 2 | 0 | 35 | 74 | 2 | 0 | 113 |
| 2000 | 7 | 65 | 6 | 104 | 8 | 0 | 190 |
| 2001 | 0 | 47 | 625 | 3 | 11 | 0 | 686 |
| 2002 | 0 | 10 | 267 | 419 | 0 | 1 | 697 |
| 2003 | 0 | 18 | 30 | 146 | 27 | 0 | 221 |
| 2004 | 0 | 2 | 100 | 67 | 18 | 0 | 187 |
| 2005 | 0 | 12 | 19 | 104 | 15 | 0 | 150 |
| 2006 | 0 | 7 | 15 | 11 | 27 | 0 | 60 |
| 2007 | 0 | 122 | 116 | 56 | 5 | 3 | 302 |
| 2008 | 0 | 18 | 460 | 137 | 3 | 0 | 618 |
| 2009 | 0 | 43 | 33 | 158 | 2 | 0 | 236 |
| 2010 | 4 | 20 | 293 | 29 | 7 | 0 | 353 |
| 2011 | 0 | 144 | 47 | 118 | 0 | 0 | 309 |
| 2012 | 1 | 31 | 168 | 63 | 7 | 0 | 270 |
| 2013 | 0 | 7 | 27 | 22 | 2 | 1 | 59 |
| 2014 | 0 | 55 | 58 | 39 | 0 | 0 | 152 |
| 2015 | 0 | 17 | 234 | 49 | 0 | 0 | 300 |
| 2016 | 0 | 6 | 15 | 74 | 4 | 0 | 99 |
| 2017 | 0 | 3 | 19 | 20 | 5 | 0 | 47 |
| 2018 | 0 | 0 | 32 | 7 | 1 | 0 | 40 |
| 2019 | 0 | 3 | 21 | 18 | 1 | 0 | 43 |
| 2020 | 0 | 0 | 5 | 8 | 0 | 0 | 13 |
| Average | 1 | 23 | 100 | 68 | 6 | 0 | 198 |

## Hatchery-Origin Female

Salt Age Carcasses Recovered

| Survey Year | 0 | 1 | 2 | 3 | 4 | 5 | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1993 | 0 | 0 | 10 | 1 | 0 | 0 | 11 |
| 1994 | 0 | 0 | 3 | 141 | 1 | 0 | 145 |
| 1995 | 0 | 0 | 9 | 44 | 82 | 0 | 135 |
| 1996 | 0 | 0 | 21 | 74 | 31 | 1 | 127 |
| 1997 | 0 | 0 | 2 | 107 | 16 | 0 | 125 |
| 1998 | 0 | 1 | 28 | 30 | 32 | 0 | 91 |
| 1999 | 1 | 0 | 31 | 393 | 13 | 2 | 440 |
| 2000 | 0 | 1 | 4 | 307 | 49 | 0 | 361 |
| 2001 | 0 | 1 | 256 | 19 | 42 | 0 | 318 |
| 2002 | 0 | 0 | 54 | 921 | 9 | 0 | 984 |
| 2003 | 0 | 1 | 9 | 368 | 54 | 0 | 432 |
| 2004 | 0 | 0 | 22 | 103 | 69 | 0 | 194 |
| 2005 | 0 | 0 | 11 | 303 | 64 | 2 | 380 |
| 2006 | 0 | 0 | 10 | 21 | 48 | 0 | 79 |
| 2007 | 0 | 0 | 53 | 178 | 22 | 4 | 257 |
| 2008 | 0 | 0 | 197 | 267 | 25 | 1 | 490 |
| 2009 | 0 | 0 | 9 | 516 | 22 | 0 | 547 |
| 2010 | 0 | 0 | 155 | 120 | 42 | 1 | 318 |
| 2011 | 0 | 1 | 22 | 602 | 6 | 0 | 631 |
| 2012 | 0 | 1 | 153 | 140 | 25 | 0 | 319 |
| 2013 | 1 | 0 | 34 | 188 | 7 | 0 | 230 |
| 2014 | 0 | 0 | 23 | 127 | 5 | 0 | 155 |
| 2015 | 0 | 1 | 138 | 102 | 5 | 0 | 246 |
| 2016 | 0 | 0 | 6 | 283 | 13 | 0 | 302 |
| 2017 | 0 | 1 | 19 | 38 | 37 | 0 | 95 |
| 2018 | 0 | 0 | 46 | 59 | 7 | 0 | 112 |
| 2019 | 0 | 0 | 3 | 10 | 0 | 0 | 13 |
| 2020 | 0 | 0 | 9 | 3 | 0 | 0 | 12 |
| Average | 0 | 0 | 48 | 195 | 26 | 0 | 270 |


| Natural-Origin Male <br> Salt Age Carcasses Recovered |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Survey Year | 0 | 1 | 2 | 3 | 4 | 5 | Total |
| 1993 | 0 | 0 | 8 | 19 | 3 | 0 | 30 |
| 1994 | 0 | 3 | 13 | 22 | 10 | 0 | 48 |
| 1995 | 0 | 0 | 6 | 11 | 4 | 0 | 21 |
| 1996 | 0 | 1 | 7 | 4 | 1 | 0 | 13 |
| 1997 | 0 | 3 | 8 | 8 | 1 | 0 | 20 |
| 1998 | 0 | 3 | 32 | 27 | 5 | 0 | 67 |
| 1999 | 0 | 0 | 22 | 39 | 8 | 1 | 70 |
| 2000 | 0 | 6 | 24 | 27 | 12 | 0 | 69 |
| 2001 | 0 | 13 | 82 | 168 | 8 | 0 | 271 |
| 2002 | 0 | 15 | 85 | 232 | 52 | 1 | 385 |
| 2003 | 0 | 12 | 55 | 171 | 34 | 0 | 272 |
| 2004 | 0 | 19 | 226 | 166 | 303 | 3 | 717 |
| 2005 | 0 | 1 | 129 | 447 | 28 | 4 | 609 |
| 2006 | 0 | 1 | 14 | 189 | 116 | 0 | 320 |
| 2007 | 0 | 17 | 67 | 53 | 226 | 5 | 368 |
| 2008 | 0 | 8 | 258 | 263 | 13 | 2 | 544 |
| 2009 | 0 | 10 | 21 | 276 | 31 | 0 | 338 |
| 2010 | 0 | 3 | 90 | 123 | 50 | 0 | 266 |
| 2011 | 0 | 10 | 46 | 228 | 17 | 0 | 301 |
| 2012 | 1 | 14 | 160 | 112 | 58 | 0 | 345 |
| 2013 | 0 | 6 | 83 | 140 | 12 | 0 | 241 |
| 2014 | 0 | 43 | 135 | 633 | 76 | 0 | 887 |
| 2015 | 0 | 8 | 809 | 402 | 113 | 0 | 1,332 |
| 2016 | 0 | 1 | 53 | 548 | 109 | 1 | 712 |
| 2017 | 0 | 0 | 15 | 176 | 159 | 3 | 353 |
| 2018 | 0 | 2 | 29 | 49 | 25 | 0 | 105 |
| 2019 | 0 | 0 | 40 | 42 | 6 | 0 | 88 |
| 2020 | 0 | 2 | 92 | 518 | 41 | 0 | 653 |
| Average | 0 | 7 | 93 | 182 | 54 | 1 | 337 |


| Natural-Origin Female <br> Salt Age Carcasses Recovered |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Survey Year | 0 | 1 | 2 | 3 | 4 | 5 | Total |
| 1993 | 0 | 0 | 5 | 25 | 3 | 0 | 33 |
| 1994 | 0 | 0 | 2 | 36 | 29 | 0 | 67 |
| 1995 | 0 | 0 | 7 | 27 | 11 | 0 | 45 |
| 1996 | 0 | 0 | 3 | 18 | 2 | 0 | 23 |
| 1997 | 0 | 0 | 12 | 31 | 10 | 0 | 53 |
| 1998 | 0 | 0 | 21 | 51 | 12 | 0 | 84 |
| 1999 | 0 | 0 | 32 | 132 | 34 | 0 | 198 |
| 2000 | 0 | 0 | 9 | 106 | 32 | 0 | 147 |
| 2001 | 0 | 0 | 11 | 237 | 12 | 0 | 260 |
| 2002 | 0 | 0 | 18 | 199 | 90 | 0 | 307 |
| 2003 | 2 | 2 | 29 | 130 | 45 | 0 | 208 |
| 2004 | 0 | 0 | 37 | 233 | 539 | 2 | 811 |
| 2005 | 0 | 0 | 28 | 566 | 71 | 7 | 672 |
| 2006 | 0 | 0 | 2 | 250 | 256 | 2 | 510 |
| 2007 | 0 | 0 | 8 | 72 | 601 | 12 | 693 |
| 2008 | 0 | 0 | 12 | 269 | 19 | 3 | 303 |
| 2009 | 0 | 0 | 3 | 473 | 112 | 0 | 588 |
| 2010 | 0 | 0 | 20 | 195 | 226 | 1 | 442 |
| 2011 | 0 | 0 | 12 | 416 | 58 | 0 | 486 |
| 2012 | 0 | 0 | 15 | 195 | 196 | 0 | 406 |
| 2013 | 0 | 0 | 5 | 254 | 27 | 0 | 286 |
| 2014 | 0 | 3 | 24 | 809 | 189 | 0 | 1,025 |
| 2015 | 0 | 0 | 66 | 342 | 426 | 1 | 835 |
| 2016 | 0 | 0 | 4 | 927 | 288 | 4 | 1,223 |
| 2017 | 0 | 0 | 4 | 127 | 367 | 7 | 505 |
| 2018 | 0 | 0 | 10 | 102 | 63 | 0 | 175 |
| 2019 | 0 | 0 | 0 | 87 | 22 | 0 | 109 |
| 2020 | 0 | 0 | 4 | 720 | 102 | 0 | 826 |
| Average | 0 | 0 | 18 | 249 | 136 | 1 | 404 |

Table C 3. Salt age structure (percent of recovered carcasses) for sex-origin classes.

## Hatchery-Origin Male

Salt Age - Percent of carcasses recovered within origin/sex class

| Survey Year | $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\mathbf{1 9 9 3}$ | $0 \%$ | $0 \%$ | $100 \%$ | $0 \%$ | $0 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{1 9 9 4}$ | $0 \%$ | $4 \%$ | $19 \%$ | $77 \%$ | $0 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{1 9 9 5}$ | $0 \%$ | $3 \%$ | $33 \%$ | $39 \%$ | $25 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{1 9 9 6}$ | $0 \%$ | $6 \%$ | $35 \%$ | $49 \%$ | $10 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{1 9 9 7}$ | $0 \%$ | $0 \%$ | $4 \%$ | $89 \%$ | $7 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{1 9 9 8}$ | $0 \%$ | $10 \%$ | $68 \%$ | $13 \%$ | $10 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{1 9 9 9}$ | $2 \%$ | $0 \%$ | $31 \%$ | $65 \%$ | $2 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 0 0}$ | $4 \%$ | $34 \%$ | $3 \%$ | $55 \%$ | $4 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 0 1}$ | $0 \%$ | $7 \%$ | $91 \%$ | $0 \%$ | $2 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 0 2}$ | $0 \%$ | $1 \%$ | $38 \%$ | $60 \%$ | $0 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 0 3}$ | $0 \%$ | $8 \%$ | $14 \%$ | $66 \%$ | $12 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 0 4}$ | $0 \%$ | $1 \%$ | $53 \%$ | $36 \%$ | $10 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 0 5}$ | $0 \%$ | $8 \%$ | $13 \%$ | $69 \%$ | $10 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 0 6}$ | $0 \%$ | $12 \%$ | $25 \%$ | $18 \%$ | $45 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 0 7}$ | $0 \%$ | $40 \%$ | $38 \%$ | $19 \%$ | $2 \%$ | $1 \%$ | $100 \%$ |
| $\mathbf{2 0 0 8}$ | $0 \%$ | $3 \%$ | $74 \%$ | $22 \%$ | $0 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 0 9}$ | $0 \%$ | $18 \%$ | $14 \%$ | $67 \%$ | $1 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 1 0}$ | $1 \%$ | $6 \%$ | $83 \%$ | $8 \%$ | $2 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 1 1}$ | $0 \%$ | $47 \%$ | $15 \%$ | $38 \%$ | $0 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 1 2}$ | $0 \%$ | $11 \%$ | $62 \%$ | $23 \%$ | $3 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 1 3}$ | $0 \%$ | $12 \%$ | $46 \%$ | $37 \%$ | $3 \%$ | $2 \%$ | $100 \%$ |
| $\mathbf{2 0 1 4}$ | $0 \%$ | $36 \%$ | $38 \%$ | $26 \%$ | $0 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 1 5}$ | $0 \%$ | $6 \%$ | $78 \%$ | $16 \%$ | $0 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 1 6}$ | $0 \%$ | $6 \%$ | $15 \%$ | $75 \%$ | $4 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 1 7}$ | $0 \%$ | $6 \%$ | $40 \%$ | $43 \%$ | $7 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 1 8}$ | $0 \%$ | $0 \%$ | $80 \%$ | $18 \%$ | $3 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 1 9}$ | $0 \%$ | $7 \%$ | $49 \%$ | $42 \%$ | $2 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{A v e r a g e}$ | $0 \%$ | $0 \%$ | $38 \%$ | $62 \%$ | $0 \%$ | $0 \%$ | $100 \%$ |
|  | $0 \%$ | $10 \%$ | $43 \%$ | $40 \%$ | $6 \%$ | $0 \%$ | $100 \%$ |

## Hatchery-Origin Female

Salt Age - Percent of carcasses recovered within origin/sex class

| Survey Year | $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\mathbf{1 9 9 3}$ | $0 \%$ | $0 \%$ | $91 \%$ | $9 \%$ | $0 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{1 9 9 4}$ | $0 \%$ | $0 \%$ | $2 \%$ | $97 \%$ | $1 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{1 9 9 5}$ | $0 \%$ | $0 \%$ | $7 \%$ | $33 \%$ | $61 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{1 9 9 6}$ | $0 \%$ | $0 \%$ | $17 \%$ | $58 \%$ | $24 \%$ | $1 \%$ | $100 \%$ |
| $\mathbf{1 9 9 7}$ | $0 \%$ | $0 \%$ | $2 \%$ | $86 \%$ | $13 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{1 9 9 8}$ | $0 \%$ | $1 \%$ | $31 \%$ | $33 \%$ | $35 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{1 9 9 9}$ | $0 \%$ | $0 \%$ | $7 \%$ | $89 \%$ | $3 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 0 0}$ | $0 \%$ | $0 \%$ | $1 \%$ | $85 \%$ | $14 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 0 1}$ | $0 \%$ | $0 \%$ | $81 \%$ | $6 \%$ | $13 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 0 2}$ | $0 \%$ | $0 \%$ | $5 \%$ | $94 \%$ | $1 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 0 3}$ | $0 \%$ | $0 \%$ | $2 \%$ | $85 \%$ | $13 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 0 4}$ | $0 \%$ | $0 \%$ | $11 \%$ | $53 \%$ | $36 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 0 5}$ | $0 \%$ | $0 \%$ | $3 \%$ | $80 \%$ | $17 \%$ | $1 \%$ | $100 \%$ |
| $\mathbf{2 0 0 6}$ | $0 \%$ | $0 \%$ | $13 \%$ | $27 \%$ | $61 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 0 7}$ | $0 \%$ | $0 \%$ | $21 \%$ | $69 \%$ | $9 \%$ | $2 \%$ | $100 \%$ |
| $\mathbf{2 0 0 8}$ | $0 \%$ | $0 \%$ | $40 \%$ | $54 \%$ | $5 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 0 9}$ | $0 \%$ | $0 \%$ | $2 \%$ | $94 \%$ | $4 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 1 0}$ | $0 \%$ | $0 \%$ | $49 \%$ | $38 \%$ | $13 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 1 1}$ | $0 \%$ | $0 \%$ | $3 \%$ | $95 \%$ | $1 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 1 2}$ | $0 \%$ | $0 \%$ | $48 \%$ | $44 \%$ | $8 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 1 3}$ | $0 \%$ | $0 \%$ | $15 \%$ | $82 \%$ | $3 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 1 4}$ | $0 \%$ | $0 \%$ | $15 \%$ | $82 \%$ | $3 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 1 5}$ | $0 \%$ | $0 \%$ | $56 \%$ | $41 \%$ | $2 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 1 6}$ | $0 \%$ | $0 \%$ | $2 \%$ | $94 \%$ | $4 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 1 7}$ | $0 \%$ | $1 \%$ | $20 \%$ | $40 \%$ | $39 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 1 8}$ | $0 \%$ | $0 \%$ | $41 \%$ | $53 \%$ | $6 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 1 9}$ | $0 \%$ | $0 \%$ | $23 \%$ | $77 \%$ | $0 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 2 0}$ | $0 \%$ | $0 \%$ | $75 \%$ | $25 \%$ | $0 \%$ | $0 \%$ | $100 \%$ |
| $\boldsymbol{A v e r a g e}$ | $0 \%$ | $0 \%$ | $24 \%$ | $62 \%$ | $14 \%$ | $0 \%$ | $100 \%$ |
|  |  |  |  |  |  |  |  |

## Natural-Origin Male

Salt Age - Percent of carcasses recovered within origin/sex class

| Survey Year | $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{T o t a l}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\mathbf{1 9 9 3}$ | $0 \%$ | $0 \%$ | $27 \%$ | $63 \%$ | $10 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{1 9 9 4}$ | $0 \%$ | $6 \%$ | $27 \%$ | $46 \%$ | $21 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{1 9 9 5}$ | $0 \%$ | $0 \%$ | $29 \%$ | $52 \%$ | $19 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{1 9 9 6}$ | $0 \%$ | $8 \%$ | $54 \%$ | $31 \%$ | $8 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{1 9 9 7}$ | $0 \%$ | $15 \%$ | $40 \%$ | $40 \%$ | $5 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{1 9 9 8}$ | $0 \%$ | $4 \%$ | $48 \%$ | $40 \%$ | $7 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{1 9 9 9}$ | $0 \%$ | $0 \%$ | $31 \%$ | $56 \%$ | $11 \%$ | $1 \%$ | $100 \%$ |
| $\mathbf{2 0 0 0}$ | $0 \%$ | $9 \%$ | $35 \%$ | $39 \%$ | $17 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 0 1}$ | $0 \%$ | $5 \%$ | $30 \%$ | $62 \%$ | $3 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 0 2}$ | $0 \%$ | $4 \%$ | $22 \%$ | $60 \%$ | $14 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 0 3}$ | $0 \%$ | $4 \%$ | $20 \%$ | $63 \%$ | $13 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 0 4}$ | $0 \%$ | $3 \%$ | $32 \%$ | $23 \%$ | $42 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 0 5}$ | $0 \%$ | $0 \%$ | $21 \%$ | $73 \%$ | $5 \%$ | $1 \%$ | $100 \%$ |
| $\mathbf{2 0 0 6}$ | $0 \%$ | $0 \%$ | $4 \%$ | $59 \%$ | $36 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 0 7}$ | $0 \%$ | $5 \%$ | $18 \%$ | $14 \%$ | $61 \%$ | $1 \%$ | $100 \%$ |
| $\mathbf{2 0 0 8}$ | $0 \%$ | $1 \%$ | $47 \%$ | $48 \%$ | $2 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 0 9}$ | $0 \%$ | $3 \%$ | $6 \%$ | $82 \%$ | $9 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 1 0}$ | $0 \%$ | $1 \%$ | $34 \%$ | $46 \%$ | $19 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 1 1}$ | $0 \%$ | $3 \%$ | $15 \%$ | $76 \%$ | $6 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 1 2}$ | $0 \%$ | $4 \%$ | $46 \%$ | $32 \%$ | $17 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 1 3}$ | $0 \%$ | $2 \%$ | $34 \%$ | $58 \%$ | $5 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 1 4}$ | $0 \%$ | $5 \%$ | $15 \%$ | $71 \%$ | $9 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 1 5}$ | $0 \%$ | $1 \%$ | $61 \%$ | $30 \%$ | $8 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 1 6}$ | $0 \%$ | $7 \%$ | $77 \%$ | $15 \%$ | $0 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 1 7}$ | $0 \%$ | $0 \%$ | $4 \%$ | $50 \%$ | $45 \%$ | $1 \%$ | $100 \%$ |
| $\mathbf{2 0 1 8}$ | $0 \%$ | $2 \%$ | $28 \%$ | $47 \%$ | $24 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 1 9}$ | $0 \%$ | $0 \%$ | $45 \%$ | $48 \%$ | $7 \%$ | $0 \%$ | $100 \%$ |
| $\mathbf{2 0 2 0}$ | $0 \%$ | $0 \%$ | $14 \%$ | $79 \%$ | $6 \%$ | $0 \%$ | $100 \%$ |
| Average | $0 \%$ | $3 \%$ | $31 \%$ | $50 \%$ | $15 \%$ | $0 \%$ | $100 \%$ |
|  |  |  |  |  |  |  |  |

Natural-Origin Female
Salt Age - Percent of carcasses recovered within origin/sex class

| Survey Year | 0 | 1 | 2 | 3 | 4 | 5 | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1993 | 0\% | 0\% | 15\% | 76\% | 9\% | 0\% | 100\% |
| 1994 | 0\% | 0\% | 3\% | 54\% | 43\% | 0\% | 100\% |
| 1995 | 0\% | 0\% | 16\% | 60\% | 24\% | 0\% | 100\% |
| 1996 | 0\% | 0\% | 13\% | 78\% | 9\% | 0\% | 100\% |
| 1997 | 0\% | 0\% | 23\% | 58\% | 19\% | 0\% | 100\% |
| 1998 | 0\% | 0\% | 25\% | 61\% | 14\% | 0\% | 100\% |
| 1999 | 0\% | 0\% | 16\% | 67\% | 17\% | 0\% | 100\% |
| 2000 | 0\% | 0\% | 6\% | 72\% | 22\% | 0\% | 100\% |
| 2001 | 0\% | 0\% | 4\% | 91\% | 5\% | 0\% | 100\% |
| 2002 | 0\% | 0\% | 6\% | 65\% | 29\% | 0\% | 100\% |
| 2003 | 1\% | 1\% | 14\% | 63\% | 22\% | 0\% | 100\% |
| 2004 | 0\% | 0\% | 5\% | 29\% | 66\% | 0\% | 100\% |
| 2005 | 0\% | 0\% | 4\% | 84\% | 11\% | 1\% | 100\% |
| 2006 | 0\% | 0\% | 0\% | 49\% | 50\% | 0\% | 100\% |
| 2007 | 0\% | 0\% | 1\% | 10\% | 87\% | 2\% | 100\% |
| 2008 | 0\% | 0\% | 4\% | 89\% | 6\% | 1\% | 100\% |
| 2009 | 0\% | 0\% | 1\% | 80\% | 19\% | 0\% | 100\% |
| 2010 | 0\% | 0\% | 5\% | 44\% | 51\% | 0\% | 100\% |
| 2011 | 0\% | 0\% | 2\% | 86\% | 12\% | 0\% | 100\% |
| 2012 | 0\% | 0\% | 4\% | 48\% | 48\% | 0\% | 100\% |
| 2013 | 0\% | 0\% | 2\% | 89\% | 9\% | 0\% | 100\% |
| 2014 | 0\% | 0\% | 2\% | 79\% | 18\% | 0\% | 100\% |
| 2015 | 0\% | 0\% | 8\% | 41\% | 51\% | 0\% | 100\% |
| 2016 | 0\% | 0\% | 0\% | 76\% | 24\% | 0\% | 100\% |
| 2017 | 0\% | 0\% | 1\% | 25\% | 73\% | 1\% | 100\% |
| 2018 | 0\% | 0\% | 6\% | 58\% | 36\% | 0\% | 100\% |
| 2019 | 0\% | 0\% | 0\% | 80\% | 20\% | 0\% | 100\% |
| 2020 | 0\% | 0\% | 0\% | 87\% | 12\% | 0\% | 100\% |
| Average | 0\% | 0\% | 10\% | 62\% | 89\% | 0\% | 100\% |

## Contribution to Fisheries

Table C 4. Estimated number and percent of hatchery-origin Okanogan/Similkameen summer/fall Chinook captured in different fisheries, brood years 1989-2013.

| Brood year | Ocean fisheries | Columbia River Fisheries |  |  | Total |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Tribal | Commercial <br> (Zones 1-5) | Recreational (sport) |  |
| 1989 | $2,360(80)$ | $553(19)$ | $0(0)$ | $53(2)$ | 2,966 |
| 1990 | $355(89)$ | $34(8)$ | $0(0)$ | $12(3)$ | 401 |
| 1991 | $220(86)$ | $37(14)$ | $0(0)$ | $0(0)$ | 257 |
| 1992 | $422(91)$ | $28(6)$ | $2(0)$ | $10(2)$ | 462 |
| 1993 | $24(80)$ | $6(20)$ | $0(0)$ | $0(0)$ | 30 |
| 1994 | $372(92)$ | $23(6)$ | $2(0)$ | $7(2)$ | 404 |
| 1995 | $643(93)$ | $9(1)$ | $12(2)$ | $25(4)$ | 689 |
| 1996 | $6(100)$ | $0(0)$ | $0(0)$ | $0(0)$ | 6 |
| 1997 | $6,483(92)$ | $136(2)$ | $36(0)$ | $424(6)$ | 7,079 |
| 1998 | $4,414(89)$ | $251(5)$ | $45(1)$ | $223(5)$ | 4,933 |
| 1999 | $1,359(68)$ | $224(11)$ | $31(2)$ | $384(19)$ | 1,998 |
| 2000 | $3,139(69)$ | $533(12)$ | $222(5)$ | $675(15)$ | 4,559 |
| 2001 | $184(58)$ | $81(25)$ | $31(10)$ | $23(7)$ | 319 |
| 2002 | $706(56)$ | $200(16)$ | $90(7)$ | $258(21)$ | 1,254 |
| 2003 | $711(38)$ | $568(30)$ | $130(7)$ | $466(25)$ | 1,875 |
| 2004 | $3,153(39)$ | $2,162(26)$ | $694(8)$ | $2,168(27)$ | 8,177 |
| 2005 | $470(46)$ | $306(30)$ | $79(8)$ | $167(16)$ | 1,022 |
| 2006 | $3,136(37)$ | $3,352(40)$ | $469(6)$ | $1,419(17)$ | 8,376 |
| 2007 | $1,549(44)$ | $992(28)$ | $67(2)$ | $905(26)$ | 3,513 |
| 2008 | $4,226(38)$ | $2,576(39)$ | $218(2)$ | $3,969(36)$ | 10,989 |
| 2009 | $2,005(36)$ | $2,155(39)$ | $207(5)$ | $1,138(21)$ | 5,505 |
| 2010 | $3,193(38)$ | $3,933(46)$ | $247(4)$ | $1,110(13)$ | 8,483 |
| 2011 | $5,801(40)$ | $5,812(40)$ | $456(3)$ | $2,598(18)$ | 14,667 |
| 2012 | $771(35)$ | $827(37)$ | $13(1)$ | $619(28)$ | 2,230 |
| 2013 | $1,640(684)$ | $2,671(820)$ | $26(10)$ | $1,354(209)$ | 5,691 |
| 2014 | $738(335)$ | $1,602(483)$ | $9(4)$ | $1,015(117)$ | 3,364 |
| Average | $\mathbf{1 , 8 4 9 ( 9 8 )}$ | $\mathbf{1 , 1 1 8 ( 6 9 )}$ | $\mathbf{1 1 9 ( 3 )}$ | $732(25)$ | $\mathbf{3 , 8 1 7}$ |
| Median | $\mathbf{1 , 0 6 5 ( 6 9 )}$ | $\mathbf{4 2 0 ( 2 3 )}$ | $\mathbf{3 4 ( 2 )}$ | $\mathbf{4 0 4 ( 1 6 )}$ | $\mathbf{2 , 5 9 8}$ |

## APPENDIX D

## Glossary of Terms, Acronyms, and Abbreviations

The following is a list of key terms and variables used in the Chief Joseph Hatchery Program and in this Annual Report. This is not a complete list but provides many of the main terms used in this report or that will likely be used in future CJHP Annual Report.

Accord/MOA = A ten-year agreement (2008-2018) between BPA and the CCT whereas BPA agreed to fund pre-determined fish and wildlife projects and CCT agreed not to sue the Action Agencies regarding the BiOp for the FCRPS.

CJHP Master Plan = A three-step development and review process required for all new hatcheries funded by BPA in the Columbia Basin.
eDNA = environmental DNA; dissolved or cell-bound DNA that persists in the environment.
Escapement Target = Number of fish of all origins targeted to pass upstream of the Okanogan Adult Fish weir
$\mathbf{H O B}=$ the number of hatchery-origin fish used as hatchery broodstock.
HOR = hatchery-origin recruit. The number of HORs equals the sum of HOS + HOB + hatchery-origin fish intercepted in fisheries.

HOR Terminal Run Size = Number of Chief Joseph Hatchery HORs returning to Wells Dam HOS = the number of hatchery-origin fish spawning naturally.

Juvenile Abundance = annual abundance of out-migrant juveniles estimated by expanding data from juveniles captured at the rotary screw trap.

Met Comp = Methow composite Spring Chinook. These fish are part of the Winthrop NFH program and are intended to be used for the Okanogan reintroduction pending approval under section 10(j) of the ESA.

NOB = the number of natural-origin fish used as hatchery broodstock.
NOR = natural-origin recruit. The number of NOR's equals the sum of NOB, + NOS + natural-origin fish intercepted in fisheries.

NOR Terminal Run Size = Number of Okanogan (and Similkameen, combined) NOR's returning to Wells Dam.

NOS $=$ the number of natural-origin fish spawning naturally.
$\mathbf{p H O S}=$ proportion of natural spawners composed of HORs. Equals HOS/ (NOS + HOS).

PNI = proportion of natural influence on a composite hatchery-/natural-origin population. Can also be thought of as the percentage of time the genes of a composite population spend in the natural environment. Equals $1-\mathrm{pNOB} /(\mathrm{pNOB}+\mathrm{pHOS})$.
pNOB = proportion of hatchery broodstock composed of NORs. Equals NOB/ (HOB + NOB).
$\mathbf{S A R}=$ smolt to adult return.
Recovery Plans = Federally required plans under the Endangered Species Act that describe species status, recovery criteria and expected restoration actions.

Relative Reproductive Success = The probability that an HOR produce adult offspring and summer/fall expressed as a fraction of the same probability for a NOR

Spatial Distribution = Geographic spawning distribution of adult salmon.
Spawner Abundance $=$ Total number of adult spawners each year.
Subbasin Plans = Plans developed in the early 2000s for the NPCC project funding process describing "limiting factors" used for development of regional recovery and protection strategies.

Total NOR Recruitment = Annual number of adult recruits (catch plus escapement)

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AHA = All H Analyzer
APPT = Annual Program Planning Tool
APR = Annual Program Review
BiOp = Biological Opinion
BKD = Bacterial Kidney Disease
BPA = Bonneville Power Administration
CA = Coordinated Assessments
CBFWA = Columbia Basin Fish and Wildlife Authority
CCT = Confederated Tribes of the Colville Indian Reservation
cfs = Cubic feet per second
CJH = Chief Joseph Hatchery
CJHP = Chief Joseph Hatchery Program
Colville Tribes = Confederated Tribes of the Colville Reservation
CTFWP = Colville Tribes Fish &Wildlife Program
CRITFC = Columbia River Inter-Tribal Fish Commission
CWT = Coded Wire Tag
DI = Density Index
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DPS = Distinct Population Segment
EDT = Ecosystem Diagnostic & Treatment
ELISA = Enzyme-Linked Immunosorbent Assay
ESA = Endangered Species Act
ESU = Evolutionarily Significant Unit
FCRPS = Federal Columbia River Power System
FI = Flow Index
FPP = Fish per pound
FWS = U.S. Fish and Wildlife Service
GIS = Geographic Information System
gpm = gallons per minute
GPS = Global Positioning System
HCP = Habitat Conservation Plan(s)
HGMP = Hatchery Genetic Management Plan(s)
HSRG = Hatchery Science Review Group
ISIT = In-season Implementation Tool
ISRP = Independent Scientific Review Panel
KMQ = Key Management Questions
LNFH = Leavenworth National Fish Hatchery
NEPA = National Environmental Policy Act
NMFS = National Marine Fisheries Service
NOAA = National Oceanic and Atmospheric Administration
NPCC = Northwest Power and Conservation Council
OBMEP = Okanogan Basin Monitoring and Evaluation Program
ODFW = Oregon Department of Fish and Wildlife
ONA = Okanagan Nation Alliance
PBT = Parental Based Tagging
PIT = Passive Integrated Transponder
PNAMP = Pacific Northwest Aquatic Monitoring Partnership
PSMFC = Pacific States Marine Fisheries Commission
PTAGIS = PIT Tag Information System
PUD = Public Utility District
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RKM = River Kilometer
RM = River Mile
RMIS = Regional Mark Information System
RM\&E = Research, Monitoring, and Evaluation
RST = Rotary Screw Trap
SNP = Single Nucleotide Polymorphism
TAC $=$ Technical Advisory Committee
TRMP = Tribal Resources Management Plan
$\mathbf{T U}=$ Temperature Unit
UCSRB = Upper Columbia Salmon Recovery Board
USGS = U.S. Geological Survey
WDFW = Washington Department of Fish and Wildlife
WNFH = Winthrop National Fish Hatchery

## Appendix E

# Technical Memorandum: Minijack Rates for 2020 Chief Joseph Hatchery Integrated and Segregated Chinook Releases 



Date: July 29, 2020
From: Andrea Pearl; andrea.pearl@colvilletribes.com (509) 634-1364
To: Matthew McDaniel, Casey Baldwin, Anthony Cleveland, Jim Andrews
CC: Kirk Truscott
Subject: Minijack rates for 2020 Chief Joseph Hatchery Chinook release groups

## Background

This technical memorandum will summarize the results of gonadal-somatic index (GSI) sampling conducted by the Chief Joseph Hatchery Program (CJHP) in May 2020, and provide estimates for the rate of early maturation ("minijack rate") from each yearling group released in 2020 (brood year 2018).

Early maturation of male hatchery-origin Chinook salmon is a concern throughout the Columbia river basin, with some hatchery releases exhibiting minijack rates of over 70\% (Harstad et al. 2014). The production of high levels of minijacks is not consistent with the goals and objectives of the CJHP, which intends to produce adult fish for harvest and conservation. Additionally, the National Marine Fisheries Service (NMFS) requested that the Confederated Tribes of the Colville Indian Reservation (CCT) include an evaluation of early maturation on all yearling Chinook programs because early maturation is considered a 'take surrogate' for potential competitive interactions with natural-origin fish (NMFS 2017). The reporting requirements of NMFS were based on the methodology described in Harstad et al. (2014) that used a blood plasma test to evaluate the level of 11-ketotestosterone to estimate initiation of male maturation as mini-jacks. Absent funding to implement the 11-KT method, the CJHP elected to use a visual and GSI approach to evaluate early maturation. The GSI approach has been implemented by the

USFWS for the Leavenworth complex for a number of years with good success (Matt Cooper, personal communication). The CJHP staff believe the GSI evaluation presented herein meets the intent of the reporting requirement (\#6) described in the NMFS determination letter.

## Methods

Prior to release, approximately 300 fish were collected from each yearling 2020 Chief Joseph Hatchery (CJH) release group for dissection and examination. Similar to 2019, these fish were held at CJH after their cohorts had been released for approximately one month. This was to allow for additional maturation and facilitate distinction between mature and immature fish. The release groups are:

- Segregated spring Chinook; released from Chief Joseph Hatchery, hatchery-origin broodstock collected at the Chief Joseph Hatchery Ladder
- Segregated summer Chinook; released from Chief Joseph Hatchery, hatchery-origin broodstock collected from the Columbia River near the mouth of the Okanogan River
- Integrated spring Chinook; released from the Riverside Acclimation Pond, natural-origin MetComp broodstock from Winthrop National Fish Hatchery
- Integrated summer Chinook; released from the Omak Acclimation Pond, natural- and hatchery-origin broodstock primarily of Okanogan-origin stock
- Integrated summer Chinook; released from the Similkameen Acclimation Pond, naturaland hatchery-origin broodstock primarily of Okanogan-origin stock

Fish were euthanized with MS-222, and processed in accordance with the USFWS GSI sampling protocol (Pfannenstein 2016, see Appendix A). Males were classified as either mature or immature based on a visual inspection of the gonads, and the gonadal-somatic index (GSI) was also calculated for statistical estimation of minijack rates for each release group.

After data was collected, GSI values were analyzed using a mixture model (Medeiros, see Appendix B) in an attempt to identify immature and mature sub-populations and estimate the minijack rate within each sampled release group.

## Results

Based on the visual assessment of maturity, CJH yearlings overall displayed moderate rates of early maturity ( $11.11 \%-49.66 \%$, Table 1 ). The mixture model was fit to all release groups and encompassed a larger range of expected rates of early maturation ( $19.26 \%-65.06 \%$, Table 1). There was no distinct separation in Log 10 GSI between immature and mature fish in any of the release groups. Such a break almost occurred in the Omak integrated summer Chinook release group, but it wasn't completed separated. Nevertheless, a cutoff value for classifying sampled fish as mature or immature, and therefore a minijack rate, could be modeled for all groups (Figures 1-5). Histograms that display the distribution of Log10 GSI for each sampled release group are presented in Figures 1-5. Annual rates of early maturation are recorded in Table 2.

Table 1. Mini-jack rate for each Chief Joseph Hatchery release group from brood year 2018.

| Release <br> Group | Release <br> Location | Males <br> Examined | Visually <br> classified <br> immature | Visually <br> classified <br> mature | Visual <br> mini-jack <br> Rate | Modeled <br> mini-jack <br> rate |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Segregated <br> Spring <br> Yearlings | Chief Joseph <br> Hatchery | 135 | 120 | 15 | $11.11 \%$ | $19.26 \%$ |
| Segregated <br> Summer <br> Yearlings | Chief Joseph <br> Hatchery | 166 | 124 | 42 | $25.30 \%$ | $65.06 \%$ |
| Integrated <br> Spring <br> Yearlings | Riverside <br> Acclimation <br> Pond | 139 | 106 | 33 | $23.74 \%$ | $43.88 \%$ |
| Integrated <br> Summer <br> Yearlings | Omak <br> Acclimation <br> Pond | 149 | 75 | 74 | $49.66 \%$ | $54.36 \%$ |
| Integrated <br> Summer <br> Yearlings | Similkameen <br> Acclimation <br> Pond | 144 | 115 | 29 | $20.14 \%$ | $46.53 \%$ |
|  |  |  |  |  |  |  |

BY18 CJH Segregated Spring Chinook


Figure 1. Distribution of Log 10 GSI for the segregated spring Chinook released from the Chief Joseph Hatchery. The cutoff value is marked by the vertical green dashed line. It marks the point of differentiation between immature fish (appearing to the left of the cutoff line) and mature fish (appearing to the right of the line). The solid blue line shows the distribution function of immature fish, and the solid red line shows the distribution function of mature fish.

## BY18 CJH Segregated Summer Chinook



Figure 2. Distribution of Log 10 GSI for the segregated summer/fall Chinook released from the Chief Joseph Hatchery. The cutoff value is marked by the vertical green dashed line. It marks the point of differentiation between immature fish (appearing to the left of the cutoff line) and mature fish (appearing to the right of the line). The solid blue line shows the distribution function of immature fish, and the solid red line shows the distribution function of mature fish.

BY18 CJH Integrated Spring Chinook


Figure 3. Distribution of Log10 GSI for the integrated spring Chinook released from the Riverside Acclimation Pond. The cutoff value is marked by the vertical green dashed line. It marks the point of differentiation between immature fish (appearing to the left of the cutoff line) and mature fish (appearing to the right of the line). The solid blue line shows the distribution function of immature fish, and the solid red line shows the distribution function of mature fish.

## BY18 Omak Integrated Summer Chinook



Figure 4. Distribution of Log10 GSI for the integrated summer/fall Chinook released from the Omak Acclimation Pond. The cutoff value is marked by the vertical green dashed line. It marks the point of differentiation between immature fish (appearing to the left of the cutoff line) and mature fish (appearing to the right of the line). The solid blue line shows the distribution function of immature fish, and the solid red line shows the distribution function of mature fish.

## BY18 Similkameen Integrated Summer Chinook



Figure 5. Distribution of Log10 GSI for the integrated summer/fall Chinook released from the Similkameen Acclimation Pond. Since a cutoff value differentiating immature and mature subpopulations was not determinable, subpopulations distribution functions and the cutoff value are not displayed.

Table 2. Annual predicted minijack rate for all CJH release groups.

| Year |  | CJH <br> Segregated <br> Spring <br> Chinook | CJH <br> Segregated <br> Summer <br> Chinook | Riverside <br> Integrated <br> Spring <br> Chinook | Omak <br> Integrated <br> Summer <br> Chinook | Similkameen <br> Integrated <br> Summer <br> Chinook |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Visual <br> Estimate | $3.23 \%$ | $4.29 \%$ | $1.34 \%$ | $0.00 \%$ | $0.75 \%$ |
|  | Modeled <br> Estimate | $4.52 \%$ | N/A | N/A | N/A | N/A |
|  | Visual <br> Estimate | $31.29 \%$ | $14.29 \%$ | $37.41 \%$ | $19.63 \%$ | $14.25 \%$ |
|  | Modeled <br> Estimate | $19.02 \%$ | $43.06 \%$ | $42.17 \%$ | $29.63 \%$ | N/A |
|  | Visual <br> Estimate | $11.11 \%$ | $25.30 \%$ | $23.74 \%$ | $49.66 \%$ | $20.14 \%$ |
|  | Modeled | $19.26 \%$ | $65.06 \%$ | $43.88 \%$ | $54.36 \%$ | $46.53 \%$ |

## Discussion and Recommendations

The data and analyses presented herein suggest that the early maturation rates for brood year 2018 releases were much higher than that of brood year 2016 and 2017 Chinook for some of the release groups. The increase in minijack rates occurred with all of the summer Chinook release groups with almost a two-fold increase in the integrated group and a one and half increase in the segregated group. A potential cause for this increase in minijack rates could be due to the failure of the chiller during the incubation stage in November and December of 2018. These release groups were not incubated under chilled water during the eye up stage and were therefore ponded earlier than expected due to premature hatching. The spring Chinook release groups had similar minijack rates to those in 2019 and were still comparable to other Columbia River hatchery programs (Harstad et al. 2014).

Although the range of rates of minijacking between release groups estimated by visual assessment and the mixture model were similar for some groups, there was not perfect agreement between the two methodologies. This predictive exercise should be paired with a retrospective analysis which uses PIT tag data to estimate actual rates of minijacking within each release group. Such an analysis could shed light on whether one method of estimating minijack rate is more accurate than the other. Or, if PIT analysis shows rates of early maturation that are
strongly divergent from both of the GSI-based estimates, that could provide a basis for future implementation of 11-KT testing.

Visual determination of maturity state is subjective and is likely only useful when the state of maturity has progressed to the point where it becomes so clear that observer error or bias can be overcome. Similarly, the mixture model relies on an ability to differentiate between two distinct, normally distributed populations within a sample. Holding the fish for an additional month postrelease allowed more time for gonadal development in the early maturing fish. Similar to the 2019 releases, this allowed for mixture model convergence at a much higher rate than in 2018, and may have contributed to reducing Type II error in the visual determination. Although this implies that the minijack rates reported in 2019 may have been artificially low, such a determination cannot be confidently made without supportive PIT tag data. It is recommended that a holdover period similar to what was employed in 2019 and 2020 be maintained in future years.

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# ‘NAD Sampling Protocols 

## Supplies List

Sampling How-To
Data Summary and Analysis Methods
Notes from 2016


By Katy Pfannenstein

Mid-Columbia River Fishery Resource Office

US Fish and Wildlife Service

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NAD Supplies List [Bracketed numbers are minimum numbers needed for ONE CREW, 4-6 people, for 300 fish]

Daily consumables:

- Data sheets: Length/weight sheet AND gonad weight sheet (Rite in the Rain) Paper number tabs (Rite in the Rain)
- Paper towels (brown single fold, $\sim 100 /$ pack)

General:

- [3] Clipboards
- [3] Mechanical pencils + lead
- [2] Tables
- [4] Chairs
- [4] Buckets to raise table (small white)
- [2] Power strips
- [2] Extension cords
- Garbage bags
- Absorbent lab paper to cover work surfaces (roll)
- Duct tape
- Large scissors and a sharpie
- Extra batteries (9 volt + AA)
- Buckets + aerators
- Counting clickers
- Camera/iPad

Length and weight station:

- Tricane Methanesulfonate (MS 222)
- [1] Tub for fish
- [1] Dip net
- [1] Pit scanner + [1] stand
- [4] large sponges + [1] cookie tray
- [1] Scale for weights + [1] smolt weight pan
- [1] Length board


## Dissecting station:

- [1 or 2] Micro scale (minimum power 0.001 g ) + power cords
- [4] Scissors + [4] tweezers
- [2] Buckets for garbage (5 gallon)
- S/M/L glove boxes
- Weigh boats for scales
- Portable lights


## 'NAD Sampling How-To

1. Prepare TWO different data sheets: one with fish ID, fork length, weight, smolt index (0-3), pit \#, and the other with fish ID, sex (M/F), maturation (0-2), gonad weight. Each fish will have an individual fish ID number, which will be matched up during data entry. Measure fish body weight to the nearest 0.1 g and gonad weight to 0.0001 g .

## PRE-RELEASE JUVENILE SAMPLING DATA SHEET

Page $\qquad$ of $\qquad$

Date: $\qquad$ Samplers: $\qquad$
Hatchery: $\qquad$ Species/Stock $\qquad$
Group: $\qquad$ Bank: $\qquad$ Raceway(s) $\qquad$
Other: $\qquad$
Smolt index ( $0=$ unk, $1=$ parr, $2=$ trans, $3=$ smolt ) Maturity ( $0=$ unknown, $1=$ immature, $2=$ mature )

| Fish ID\# | Fork Ln <br> (mm) | WGHT <br> (gms) | Smolt <br> Index <br> $(0-3)$ | PIT \# <br> (last 4) | CWT ID \# | Sex <br> (M/F) | Maturity <br> (0-2) | Gonad <br> Wt. (gms) | Comment |
| :--- | :--- | :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
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PRE-RELEASE JUVENILE SAMPLING DATA SHEET
Page $\qquad$ of $\qquad$

Date: $\qquad$ _/20 $\qquad$
Hatchery: $\qquad$ Species/Stock $\qquad$
Group: $\qquad$ Bank: $\qquad$ Raceway(s) $\qquad$
Other:
Smolt index (0= unk, 1= parr, 2= trans, 3=smolt) Maturity (0=unknown, 1=immature, 2=mature)

| Fish ID\# | Sex (M/F) | Maturity (0-2) | Gonad Wt. <br> (gms) | Comment |
| :---: | :---: | :---: | :---: | :---: |
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2. Collect fish from hatchery ponds. Random sample? Keep different ponds separate? CWT? Pit Tag?
3. Set up stations. Note length/weight station is at standing height.

4. Smolt index: 1. Parr, dark marks (bottom fish), 2. Transitional, faded marks (middle fish), 3. Smolt, silver, no marks (top fish)

5. Set out 15-20 fish in a row on the sponges. Add number tags to fish. Assess smolt index while all fish are in the line. Obtain weights and lengths, place on paper towel to pass to the dissecting crew.

6. Fish dissection: Cut open belly from vent (shallow incision), cut behind gill, open fish and gently remove guts to expose air bladder. Both male and female gonads are located on the top/edge of the air bladder (orange arrow on mature male).

7. Female identification: 1. Ovary forms a point and then narrows to oviduct - thread like (green arrow) 2. Ovary is angular, has ridge (blue arrow), 3. Granulated
(orange arrow), 4. Color (red arrow) is not a good indicator as it can vary from pink to white.

8. Immature male identification: Testes are thready throughout, smooth and round, no development or thickness (green arrows).

9. Mature male identification: Testes thicken, become white/translucent, smooth, tapers to tail.

10. Visually identify fish sex. If female, record fish number and sex on datasheet. If male, visually identify if immature or mature PRIOR to weighing gonads, record visual call and then remove and weigh gonads.
11. Removal of testes for weighing: Use a fine point tweezers, start as near to the anterior insertion as possible (orange arrow), gently lift the entirety of the 'nad off of air bladder down to the tail (blue arrow). Place on the back of your hand and remove second 'nad. Weigh both complete testes. If you were only able to remove one, double the weight on the datasheet, and note that only one was weighed.

12. To use the scale: Close all doors, zero balance, open door, place 'nads in weight boat, close doors, wait for number to stabilize. 'Nads will evaporate and become lighter in a short period of time.
13. Enjoy all the 'nad jokes you can handle and interagency mingling!


## NAD Data Summary and Analysis Methods

- Enter data and QA/QC work, make sure to include specific banks/raceways.
- Calculate Gonadosomatic Index (GSI = gonad weight (g)/weight (g) *100).
- Calculate Condition Factor (K= (105) *weight/length ${ }^{3}$ ).
- Calculate the Log10(GSI) and graph the frequencies in a histogram to visually see the bimodal pattern of the immature and mature males. Use this graph to determine the GSI threshold that separates immature and mature males.

- From the GSI threshold, calculate the counts, percentages, average length, weight, and condition factor for immature and mature males.
- In a summary table, for both males and females, include gender counts, percentages, and average length, weight, and condition factors. For males, summarize visual counts for immature and mature fish and the percentage of mature fish. Summarize GSI counts and percent for immature and mature fish and list the average length, weigh, and condition factor for each group. Make sure to note what GSI threshold was used.

Table x. Leavenworth National Fish Hatchery Complex juvenile pre-release/early-maturation sampling, April 5-8, 2016.

| Pre-Release Data |  |  |  |  |  |  |  | Visual Count |  |  | GSI* Count |  |  | GSI Immature Male Averages |  |  | GSI Mature Male Averages |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Site | Species | Gender | Count | Percent | Ln | Wt | K | Immature | Mature | \% | Immature | Mature | \% | Ln | Wt | K | Ln | Wt | K |

- Perform additional statistics as desired (Were the raceways different? Feed differences? Circular tanks vs. raceways, differences between years, etc). Normality, chi-squared goodness of fit, t-test, Anova, etc.


## NAD Sampling Notes (What worked? What didn't?)

- Print off more data sheets than you think you need. The two data sheet system works best; the dissectors can record their own data.
- Have two people per dissection scale- the more people that use the scale, the more awkward it gets.
- Weighing all male gonads vs. writing " T " for threads/trace? What is best for level of accuracy desired?
- Can we eyeball maturation, i.e., distinguish between 1 (immature) and 2 (mature)?
- Can maturation be determined by gonad weight or \% GSI? OR is maturation highly variable and dependent on stock and/or sampling date?
- For data analysis, " T " weight gonads were given a gonad weight of 0.00001 g for a visual representation on the graphs.
- Steelhead that were expressing milt were assigned a maturity level of 3 , and were counted, but not weighed. For data analysis, they were assigned a gonad weight of 1.0 g in order to calculate GSI and to be visually represented on the graphs.

Thank you to everyone who participated in the 2016 'NAD sampling: USFWS, WDFW, Chelan PUD, Douglas PUD and Grant PUD!

## References:

Larsen, D. A., B. R. Beckman, K. A. Cooper, D. Barrett, M. Johnston, P. Swanson, and W. W. Dickhoff. 2004. Assessment of high rates of precocious male maturation in a spring Chinook salmon supplementation hatchery program. Transactions of the American Fisheries Society 133:98-120.

Harstad, D. L., D. A. Larsen, and B. R. Beckman. 2014. Variation in minijack rate among hatchery populations of Columbia River basin Chinook salmon. Transactions of the American Fisheries Society 143:768-778.

## Mixture model and maturity cutoff calculation

## For Data Analyses: Determine cutoff for maturing vs. non-maturing fish

From Dr. Lea Medeiros, University of Idaho Post-Doc
\# Example using C16 11-kT data from minijack study
Export list of Log(conc) or Conc (and convert to Log(conc) once imported into R studio) Import C16 CSV using import button in rStudio

- Make sure that the separator is set to "Comma" if importing a CSV... sometimes wants to import as whitespace
Copy and paste the code below the line into rStudio
\# Load the appropriate packages
library(mixtools)
library(diptest)
library(Hmisc)
\# Define variables (columns in imported CSV)
LC=C16\$Log
\# Only define variables for which you have columns
\# If value shows up as factor instead of num you have a non-numeric value in the CSV
\# Determine if distribution is bimodal
dip.test(LC) \# returns dip statistic (D) and p-value, as well as what hypothesis (i.e., initial or alternate) to accept. If alternate is accepted, proceed.
\# Determine the variables for the normal curves in the bimodal distribution
model=normalmixEM(LC)
plot(model, whichplots = 2)
\#Make sure things look right but won't actually use this graph as it plots on a density scale and may cause confusion. However, this should look pretty spot on (final graph will just be scaled up by a constant determined later on) so make sure that the point where the two curves intersect is where you are expecting the cutoff to be
\# Determine cutoff
index.lower <- which.min(model\$mu)
find.cutoff $<-$ function(proba=0.5, $\mathrm{i}=$ index.lower) \{
\#\# Cutoff such that $\operatorname{Pr}[$ drawn from bad component] $==$ proba
f $<$ - function( x ) \{
proba - (model\$lambda[i]*dnorm(x, model\$mu[i], model\$sigma[i]) /
(model\$lambda[1]*dnorm(x, model\$mu[1], model\$sigma[1]) +
model\$lambda[2]*dnorm(x, model\$mu[2], model\$sigma[2])))
\}
return(uniroot(f=f, lower=-2, upper=2)\$root) \# Careful with division by zero if changing lower and upper
cutoff <- c(find.cutoff(proba=0.5)) \# Can change to have range around $50 / 50$ probability, but this is the value we use to determine if a fish is maturing or not
\# Define curves from normalmixEM for plotting on histogram
$\mathrm{h}<-$ hist(LC,ylim=c(0,140),breaks=20) \# will produce basic histogram of data used for stats it produces; may need to alter ylim to reflect frequency of tallest bin and breaks xfit $<-\operatorname{seq}(-0.7,1.4$,length $=200$ )
\#First number should minimum bin, second number should be maximum bin, length is number of plots pointed (higher number = smoother curve... to a point)
yfit1 <- model\$lambda[1]*dnorm(xfit,mean=model\$mu[1],sd=model\$sigma[1])
yfit2 <- model\$lambda[2]*dnorm(xfit,mean=model\$mu[2],sd=model\$sigma[2])
yfit1 <- yfit1*diff(h\$mids[1:2])*length(LC)
yfit2 <- yfit2*diff(h\$mids[1:2])*length(LC)
\# Plot pretty graph
v1 = seq(-0.65,1.35,length=11) \# offset from minimum bin by 0.05 so that ticks are in middle of bins
$\mathrm{v} 2=\mathrm{c}(0.2,0.32,0.50,0.80,1.26,2.0,3.2,5.0,7.9,12.6,20.0) \#$ actual $\mathrm{ng} / \mathrm{mL}$ values on $\log$ scale
hist(LC, breaks = 20, density = 10, col = "purple", xaxt="n", xlab = "Plasma [11-kt] (ng/mL)", ylim = c(0, 140), main = "Plasma [11-kT] in Yakima River Juvenile Males")
lines(xfit, yfit1, col="red", lwd=2)
lines(xfit, yfit2, col="blue", lwd=2)
axis(side $=1$, at $=v 1$, labels $=v 2$ )
abline( $\mathrm{v}=$ cutoff, col="green", lty=2, lwd=2)
text(0.05,135, paste("Minijack cutoff", " $\mathrm{n}=$ =", round(10^(cutoff), 2),"(ng/mL)" ))


[^0]:    ${ }^{1}$ Adapted from the Hatchery Reform Project, the Hatchery Science Review Group reports and independent science review.

[^1]:    ${ }^{2}$ website: http://www.psmfc.org/Regional Mark Processing_Center_RMPC

[^2]:    ${ }^{3}$ There could have been some hatchery-origin fish with an intact adipose fin. Although all summer/fall Chinook hatchery programs in the Upper Columbia strive for a $100 \%$ adipose fin clip rate, a small percentage ( $\sim 1 \%$ ) may not receive the fin clip due to mechanical failure in the marking trailer. Additionally, not all fall Chinook programs, such as Priest Rapids Hatchery, clip the adipose fin of their releases.

[^3]:    ${ }^{4}$ This $80 \%$ correction factor has also been suggested by the HSRG as a default value when no direct estimates are available (HSRG 2009). Also see HSRG 2014 for a discussion about the definition and calculation effective pHOS.

[^4]:    ${ }^{5}$ A radio tracking study showed that fewer than $50 \%$ of the natural-origin fish tagged at Wells Dam ended up in the Okanogan in 2011 and 2012 (Mann and Snow 2013).

[^5]:    ${ }^{\mathrm{a}}$ Estimates for weir efficiency are adjusted for prespawn mortality and include Chinook adults that are harvested, released, and collected for brood.
    ${ }^{\mathrm{b}}$ Estimates for weir effectiveness are adjusted for prespawn mortality and include Chinook adults that are harvested or removed for pHOS management.
    c Estimates do not include Chinook Zosel Dam counts.

[^6]:    ${ }^{6}$ Origin assignments take into account all scale, ad-mark, coded wire tag and PIT tag information available at time of publication. Values may be updated in future annual reports depending on availability of data.

[^7]:    ${ }^{7}$ From McElhany, 2000 (NOAA), a viable salmonid population is an independent population of any Pacific salmonid (genus Oncorhynchus) that has a negligible risk of extinction due to threats from demographic variation, local environmental variation, and genetic diversity changes over a 100-year time frame. The four VSP parameters are abundance, productivity, spatial structure, and diversity.

[^8]:    ${ }^{(1)}$ Includes mini jacks
    ${ }^{(2)} 24 \% \mathrm{AD}$ Present steelhead were HORs
    ${ }^{(3)} 67 \% \mathrm{AD}$ Present steelhead were HORs
    ${ }^{(4)} 147$ adults ( 80 males, 67 females) taken for transfer to Eastbank Hatchery
    (5) 98 males and 98 females taken in July and August,
    ${ }^{(6)}$ Surplussed fish
    RTS $=$ Return to stream

